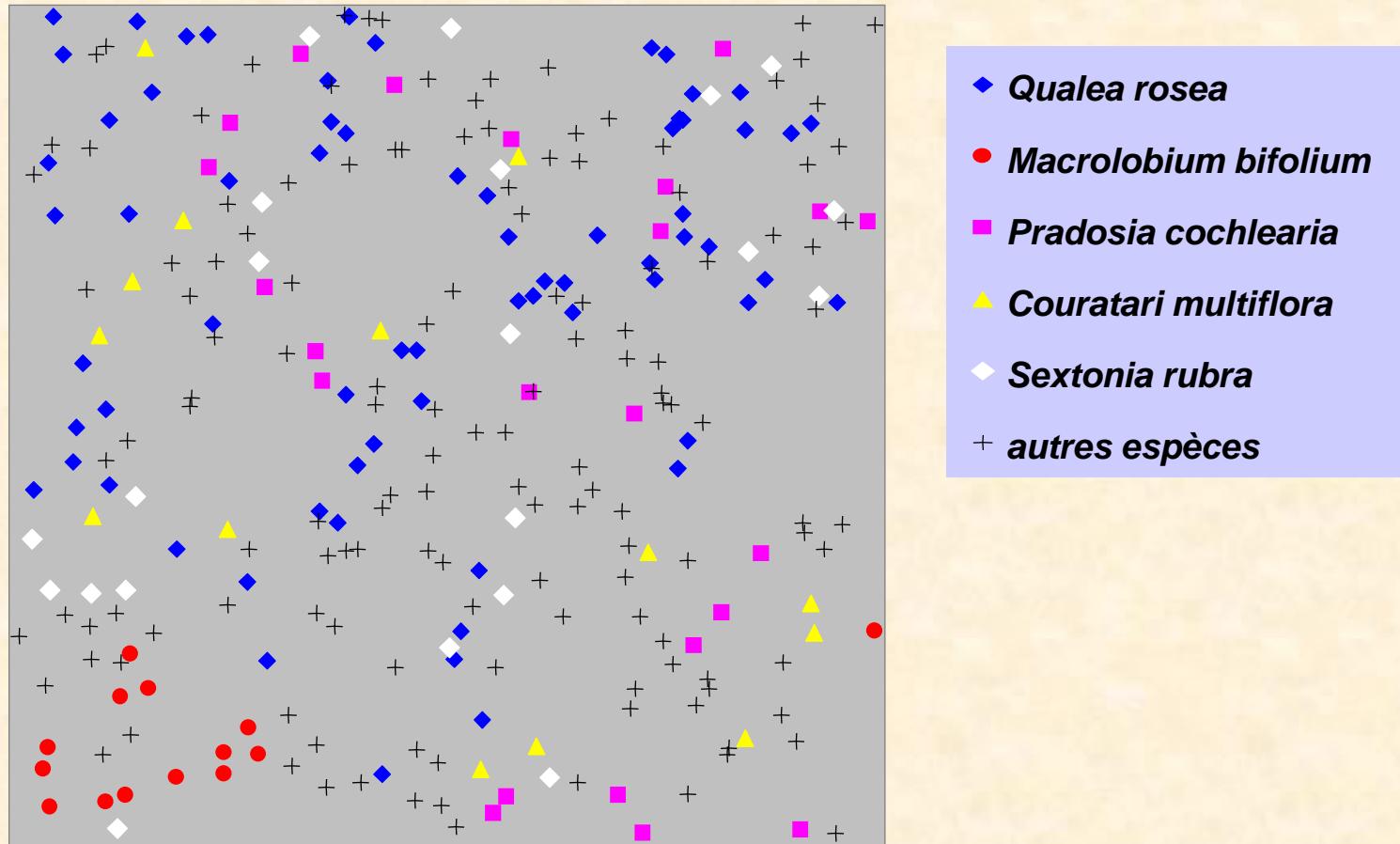


Distribution spatiale des individus et modèles neutres / de niches

Distribution de 5 espèces d'arbres dans une parcelle de 25 ha (Guyane française)

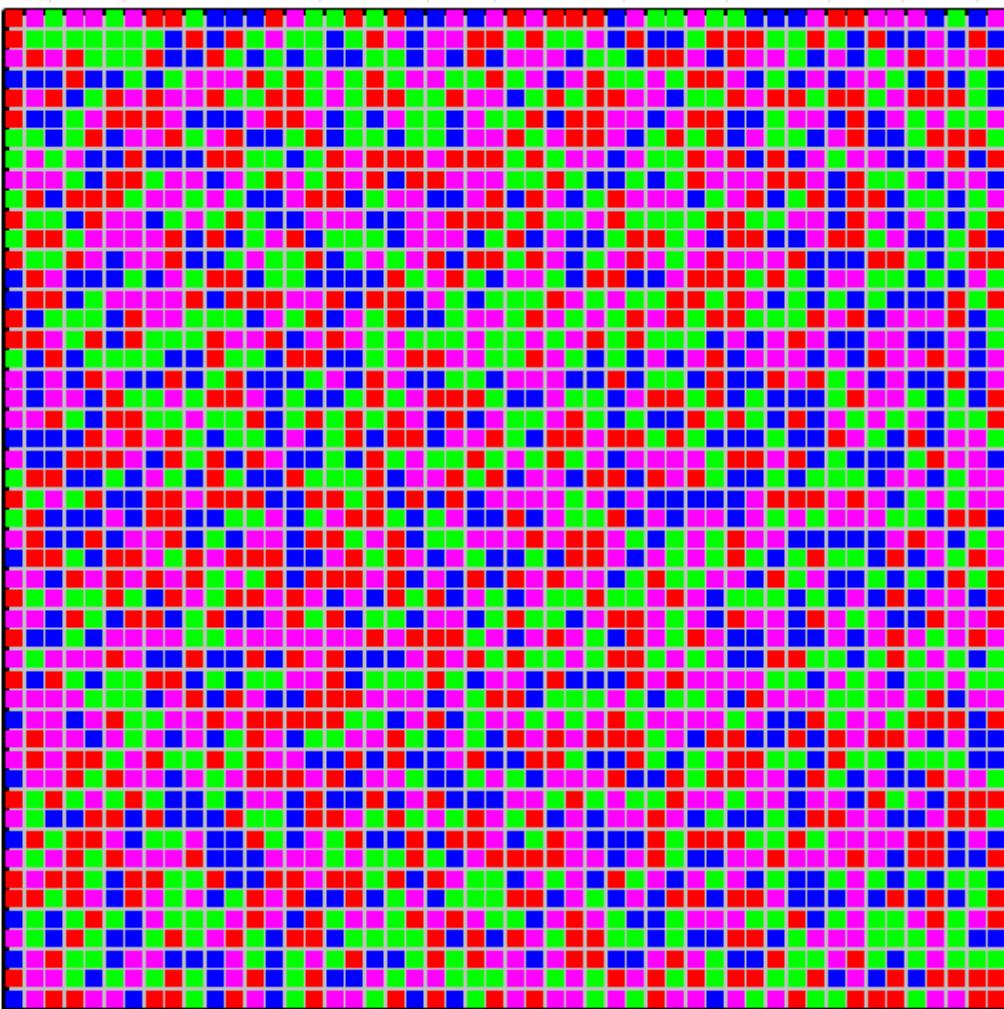


Ces distributions spatiales reflètent-elles

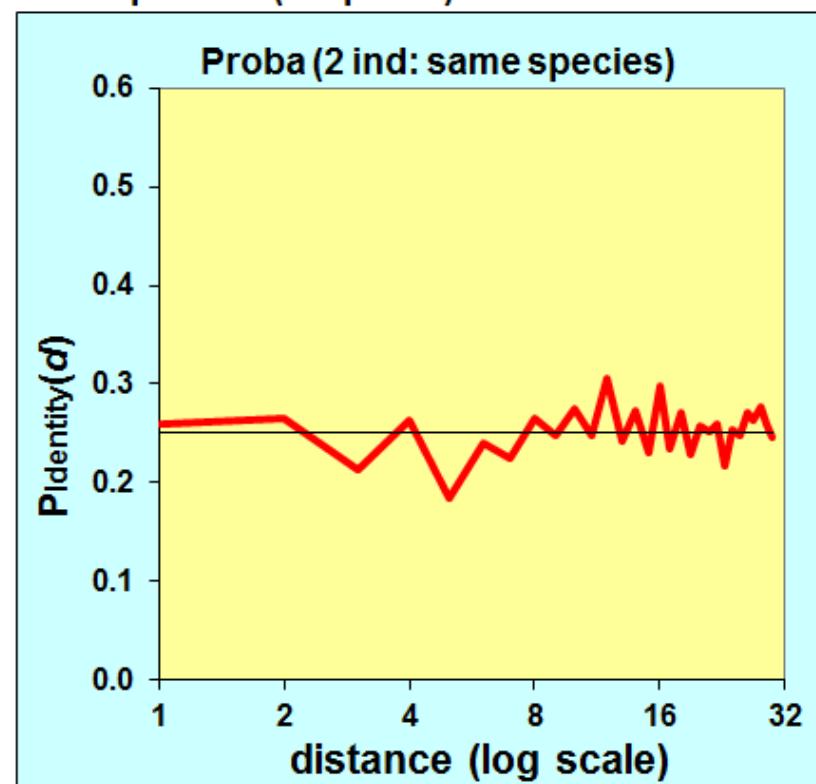
- l'effet de l'hétérogénéité des conditions du milieux ? -> niche
- une dispersion limitée des espèces ? -> neutre

Simulateur (4 espèces sur une grille de 50x50) : état initial

Mortality rate = 50% Dispersal distance = 1 step Time = 0
Habitat selection = 0 R(sp/envi)= -0



Time = 0
■ species1 (freq=0.24)
■ species2 (freq=0.26)
■ species3 (freq=0.26)
■ species4 (freq=0.24)



Neutre + dispersion très limitée (10 pas de temps)

Mortality rate =

50%

Dispersal distance =

1 step

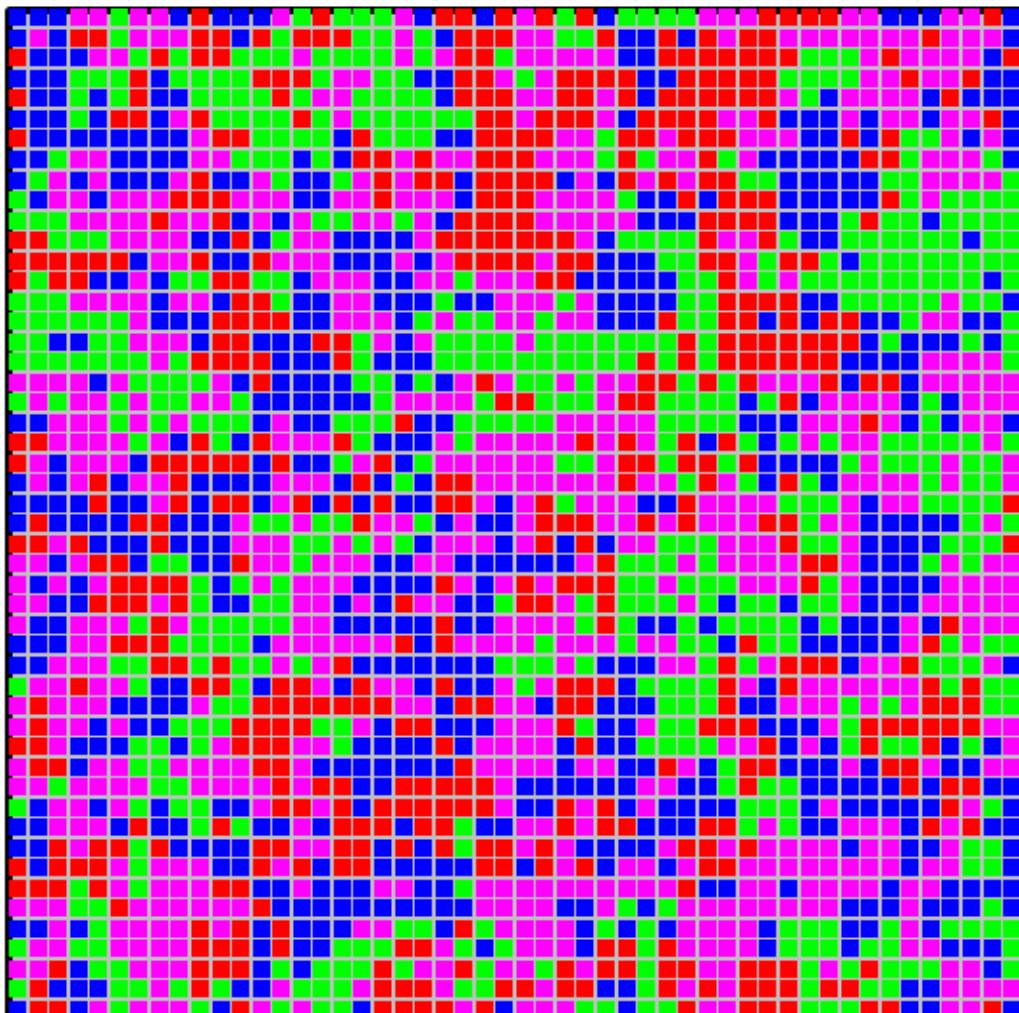
Time = 10

Habitat selection =

0

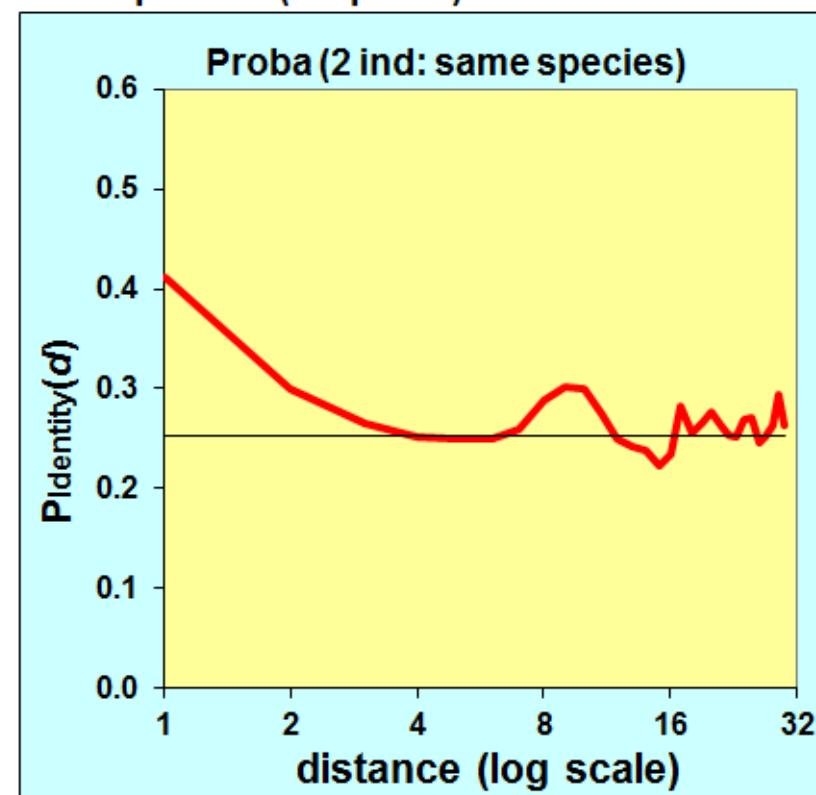
R(sp/envi)=

-0



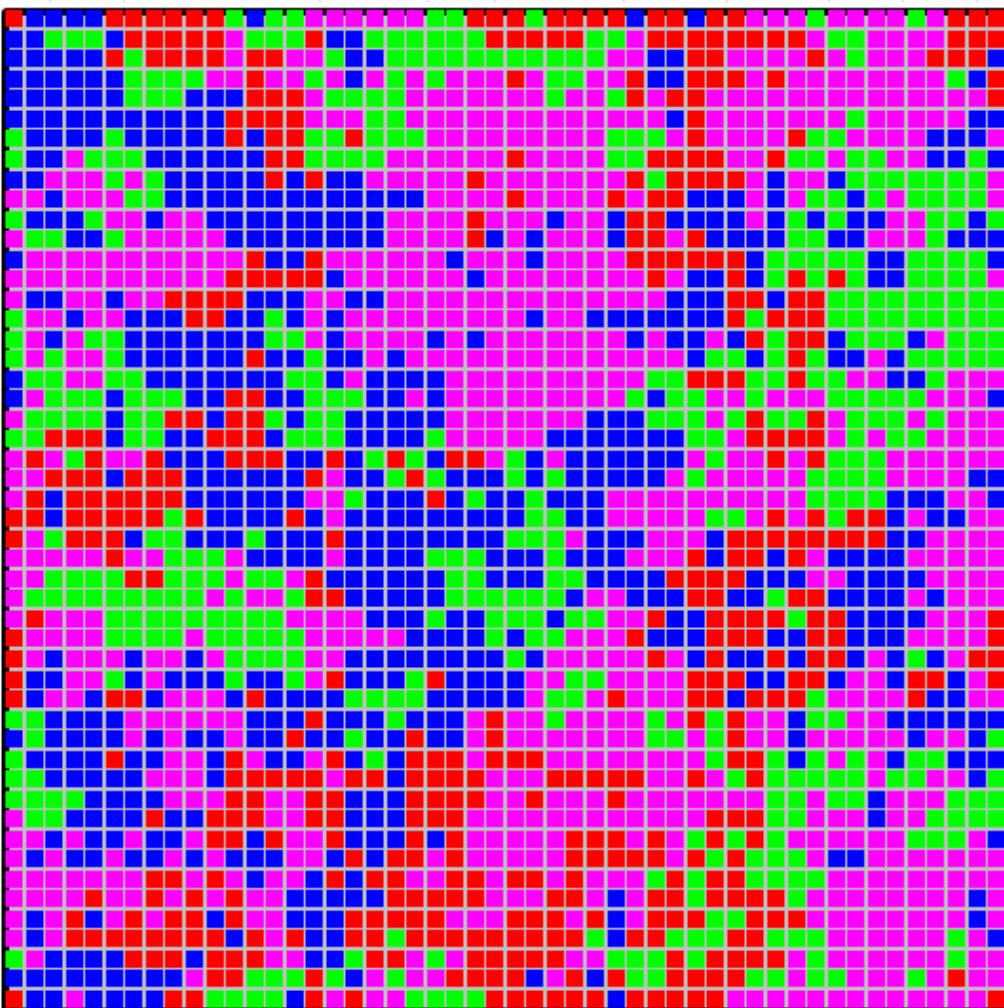
Time = 10

- species1 (freq=0.26)
- species2 (freq=0.22)
- species3 (freq=0.29)
- species4 (freq=0.23)

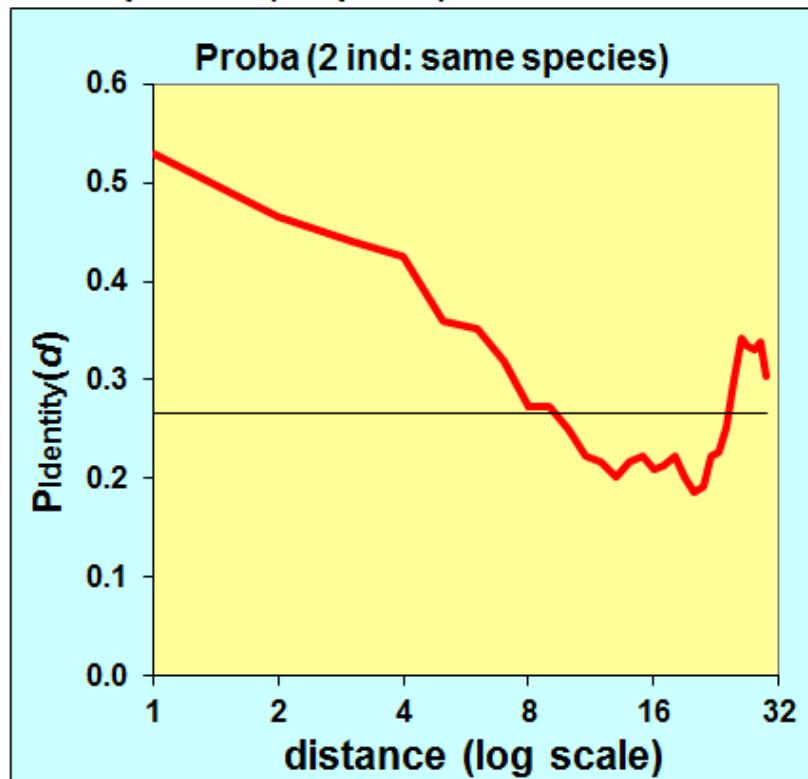


Neutre + dispersion très limitée (50 pas de temps)

Mortality rate = 50% Dispersal distance = 1 step Time = 53
Habitat selection = 0 R(sp/envi)= -0

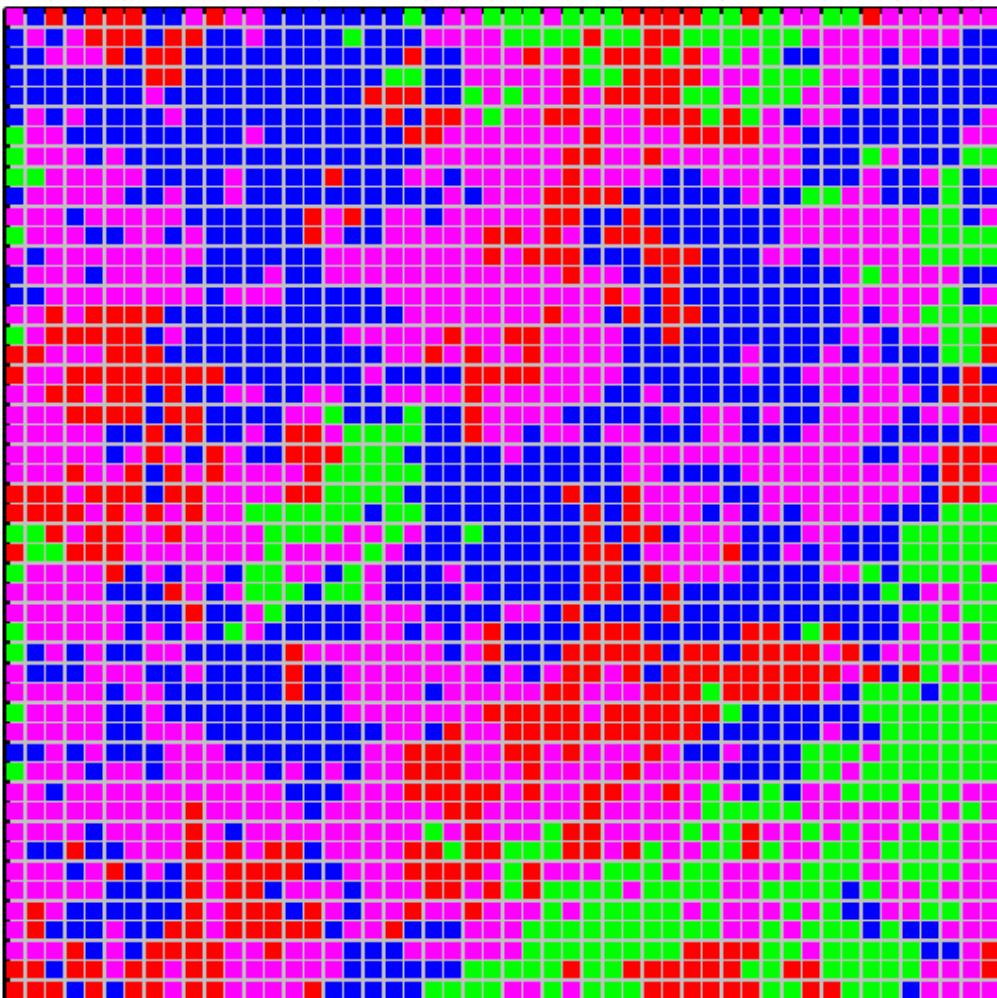


Time = 53
■ species1 (freq=0.27)
■ species2 (freq=0.2)
■ species3 (freq=0.34)
■ species4 (freq=0.19)

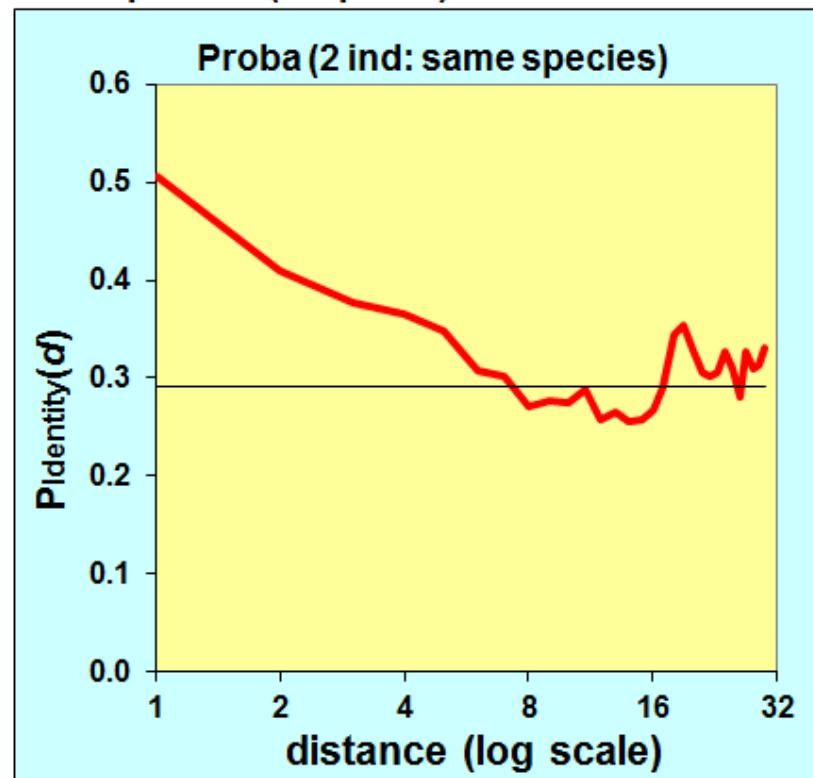


Neutre + dispersion très limitée (150 pas de temps)

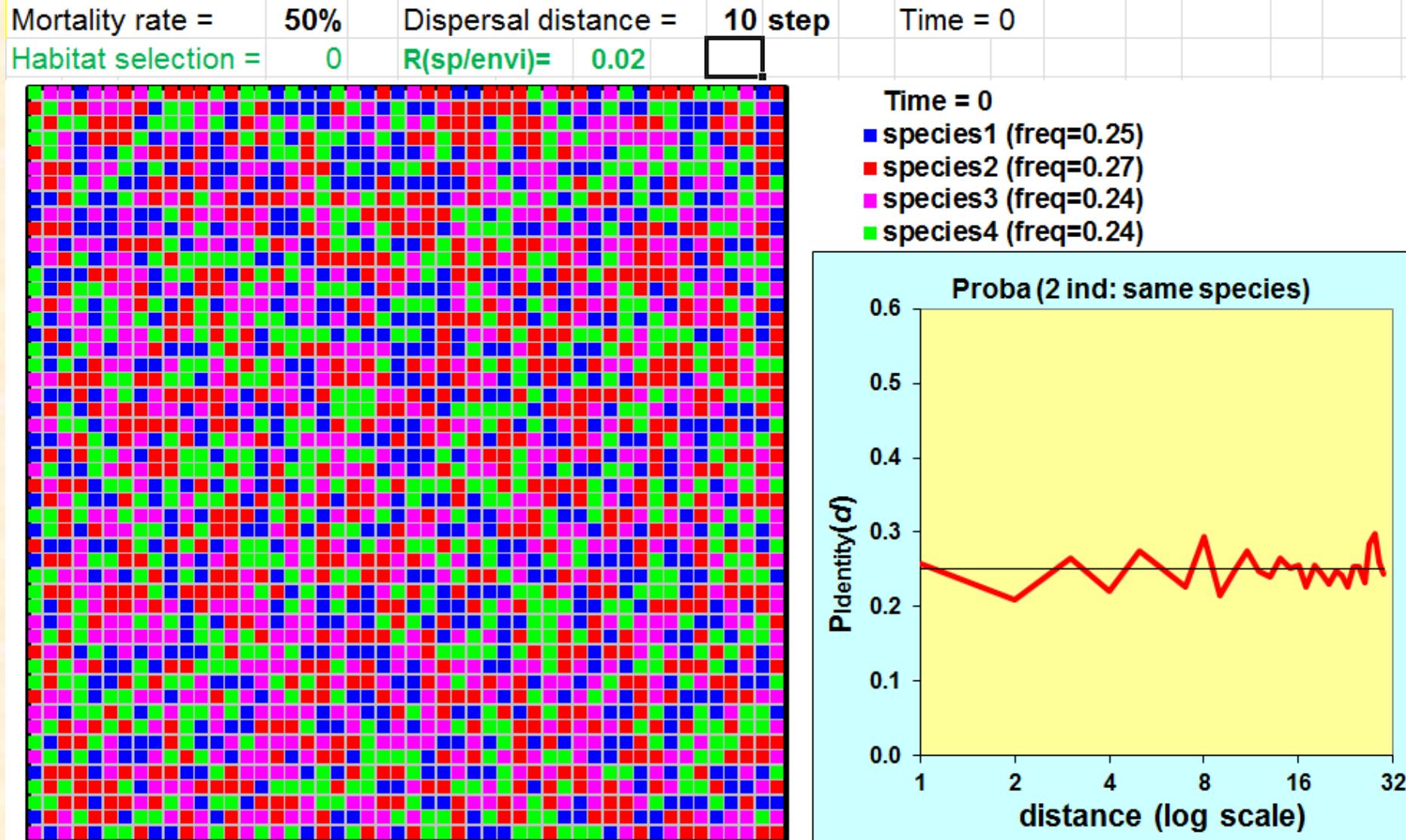
| | | | | |
|---------------------|-----|----------------------|--------|------------|
| Mortality rate = | 50% | Dispersal distance = | 1 step | Time = 153 |
| Habitat selection = | 0 | R(sp/envi)= | 0.2 | |



Time = 153
■ species1 (freq=0.33)
■ species2 (freq=0.17)
■ species3 (freq=0.37)
■ species4 (freq=0.13)



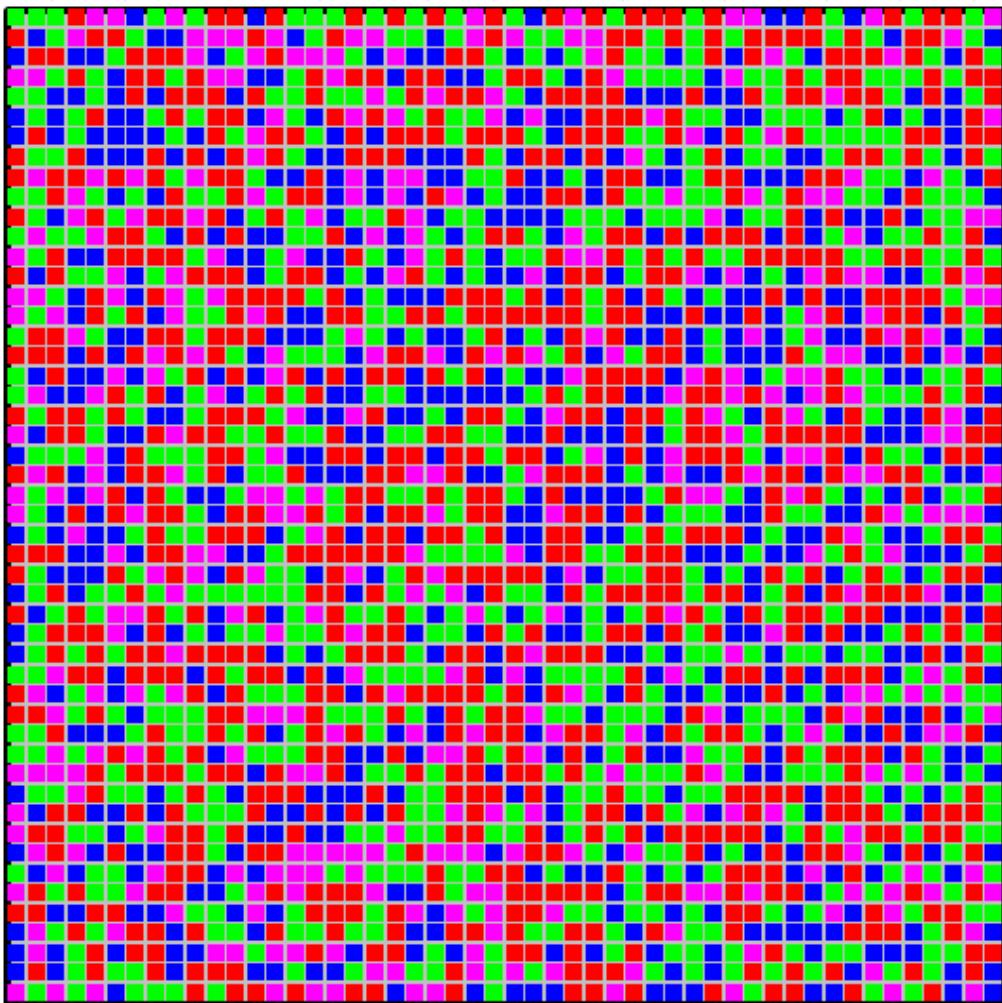
Neutre + dispersion étendue (10 pas de temps)



Neutre + dispersion étendue (100 pas de temps)

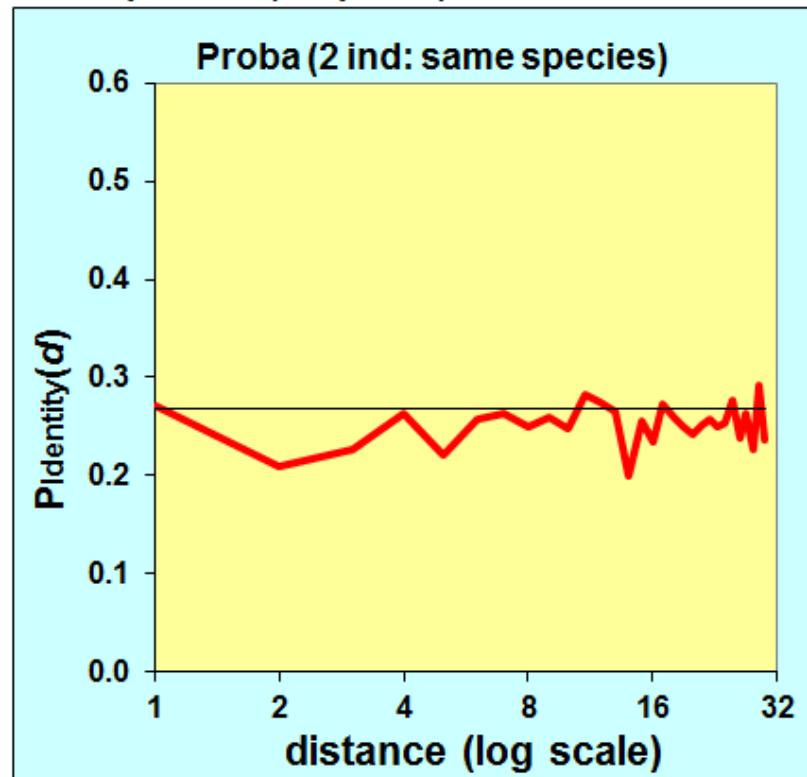
Mortality rate = 50% Dispersal distance = 10 step

Habitat selection = 0 $R(sp/envi)= 0.01$



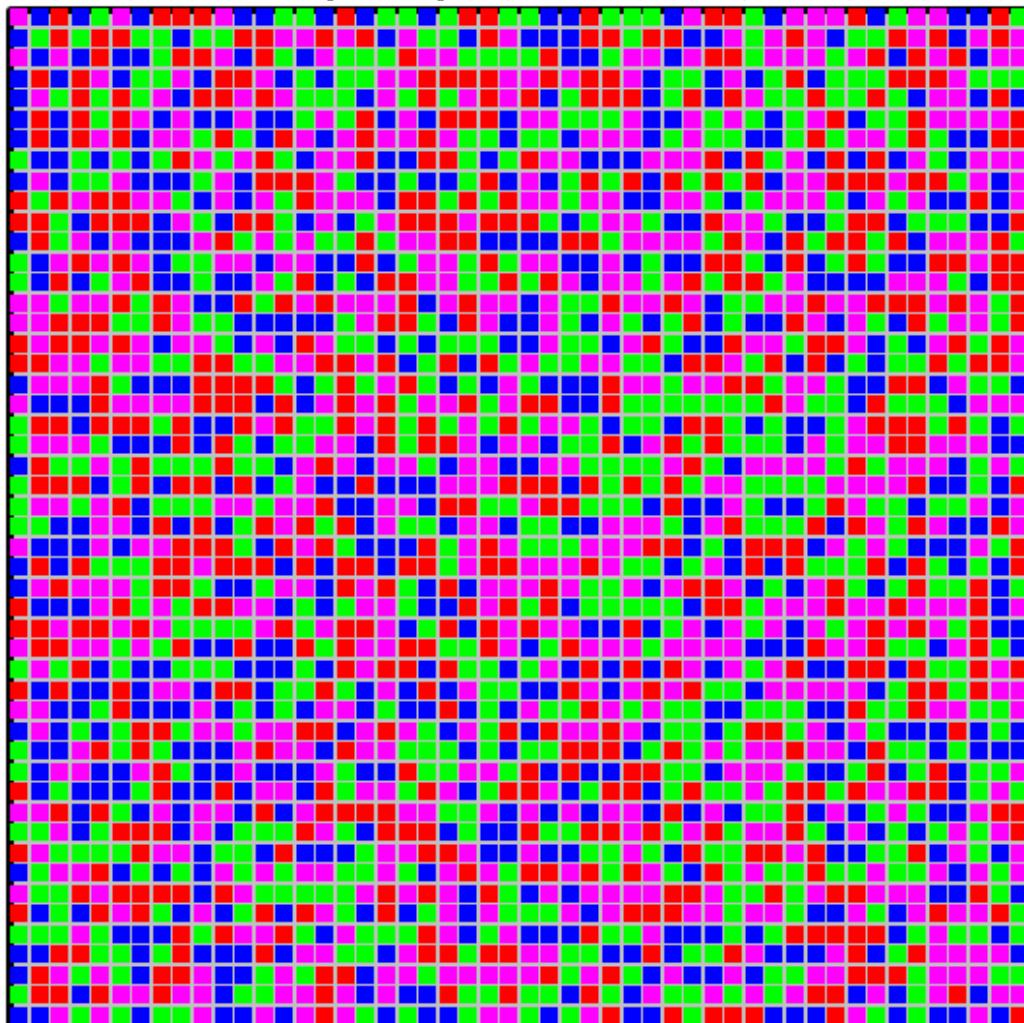
Time = 100

- Time = 100
- species1 (freq=0.25)
- species2 (freq=0.34)
- species3 (freq=0.15)
- species4 (freq=0.26)

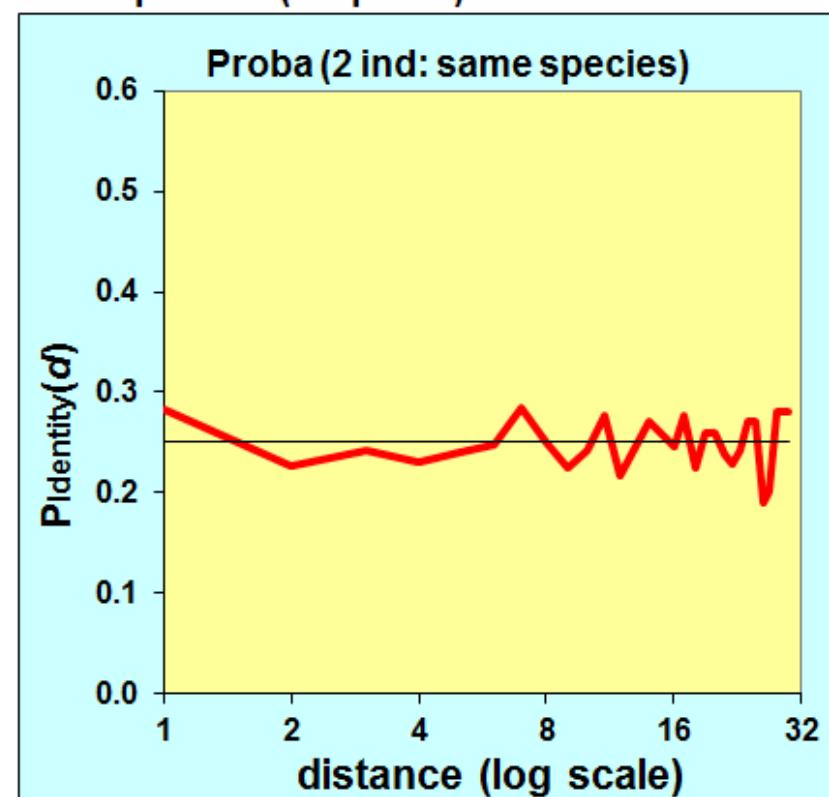


Filtre environnemental sur gradient ecol + dispersion très limitée (état initial)

Mortality rate = 50% Dispersal distance = 1 step Time = 0
Habitat selection = 0.1 $R(sp/envi)=$ 0.01

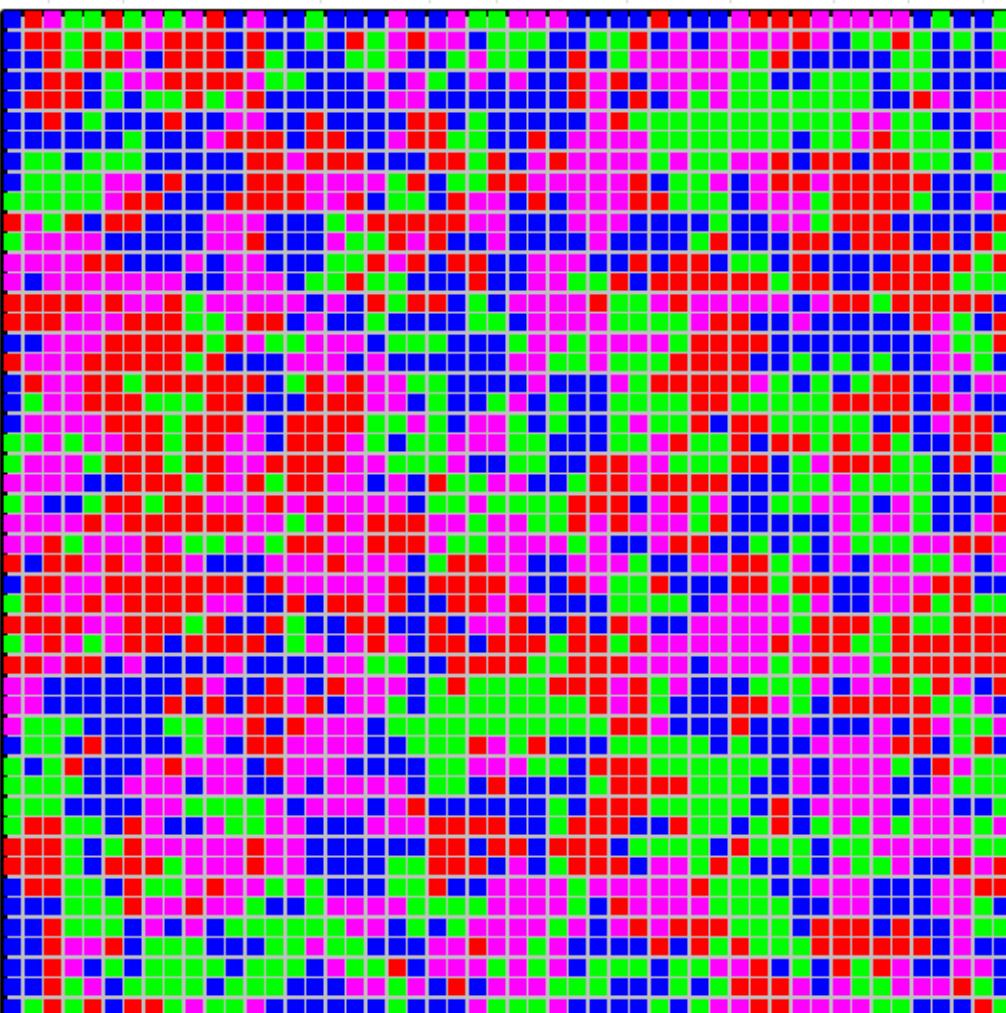


Time = 0
■ species1 (freq=0.24)
■ species2 (freq=0.24)
■ species3 (freq=0.26)
■ species4 (freq=0.25)

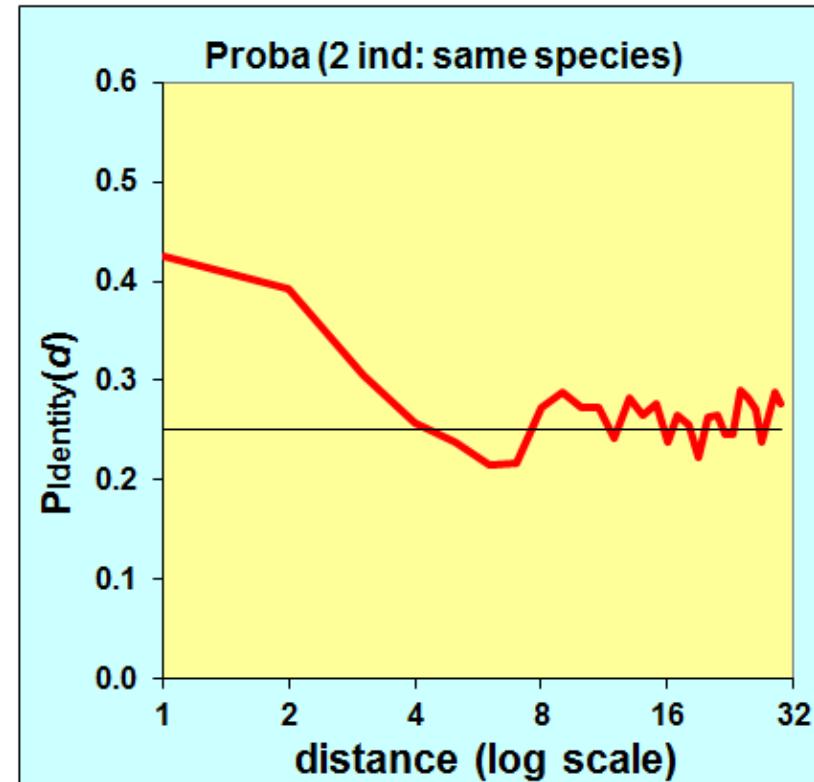


Filtre environnemental sur gradient ecol + dispersion très limitée (10 pas de temps)

Mortality rate = 50% Dispersal distance = 1 step Time = 10
Habitat selection = 0.1 R(sp/envi)= 0.07

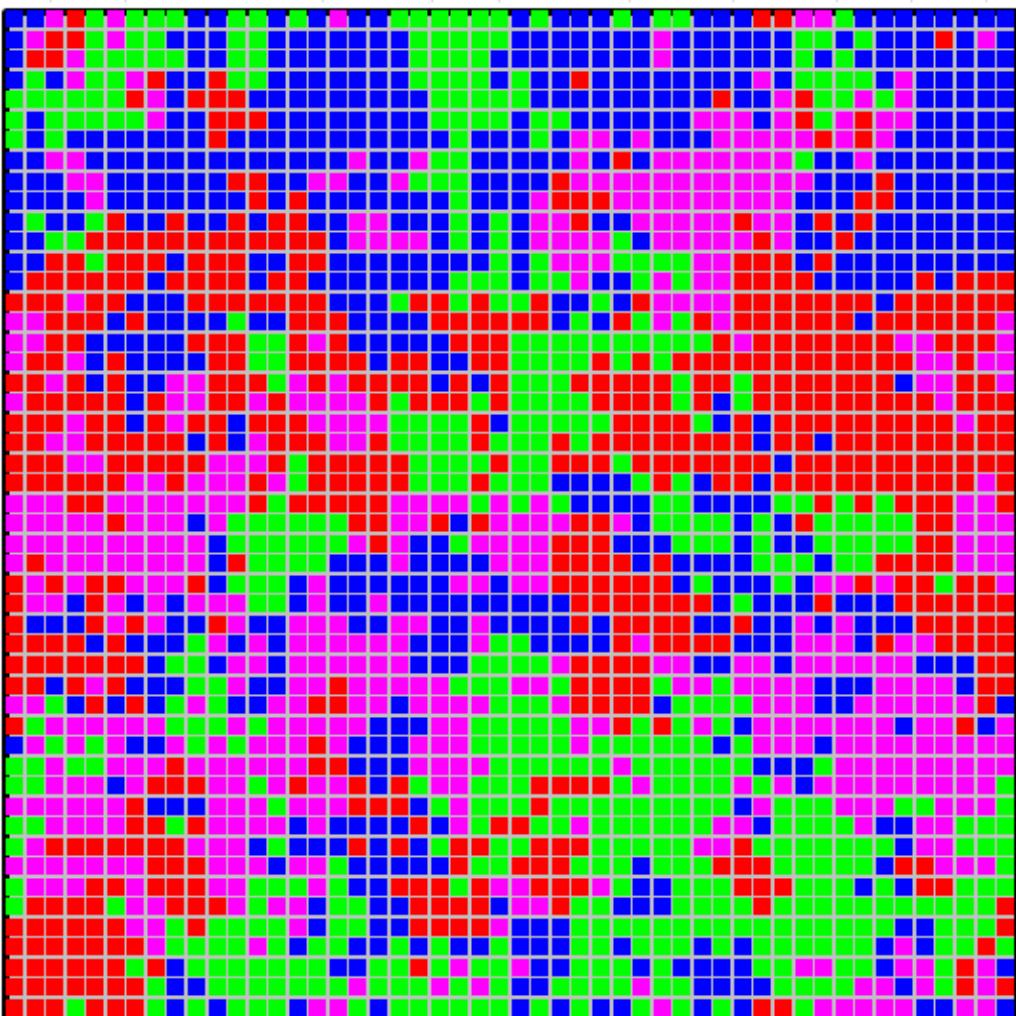


Time = 10
■ species1 (freq=0.28)
■ species2 (freq=0.24)
■ species3 (freq=0.25)
■ species4 (freq=0.23)

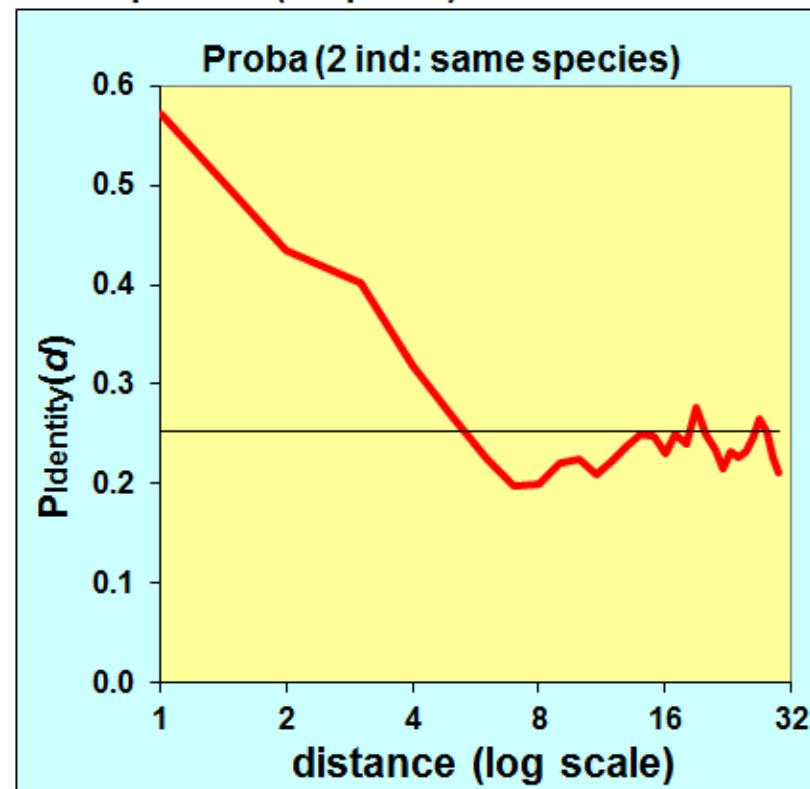


Filtre environnemental sur gradient ecol + dispersion très limitée (50 pas de temps)

Mortality rate = 50% Dispersal distance = 1 step Time = 50
Habitat selection = 0.1 R(sp/envi)= 0.29



Time = 50
■ species1 (freq=0.28)
■ species2 (freq=0.26)
■ species3 (freq=0.23)
■ species4 (freq=0.23)



Filtre environnemental sur gradient ecol + dispersion très limitée (150 pas de temps)

Mortality rate =

50%

Dispersal distance =

1 step

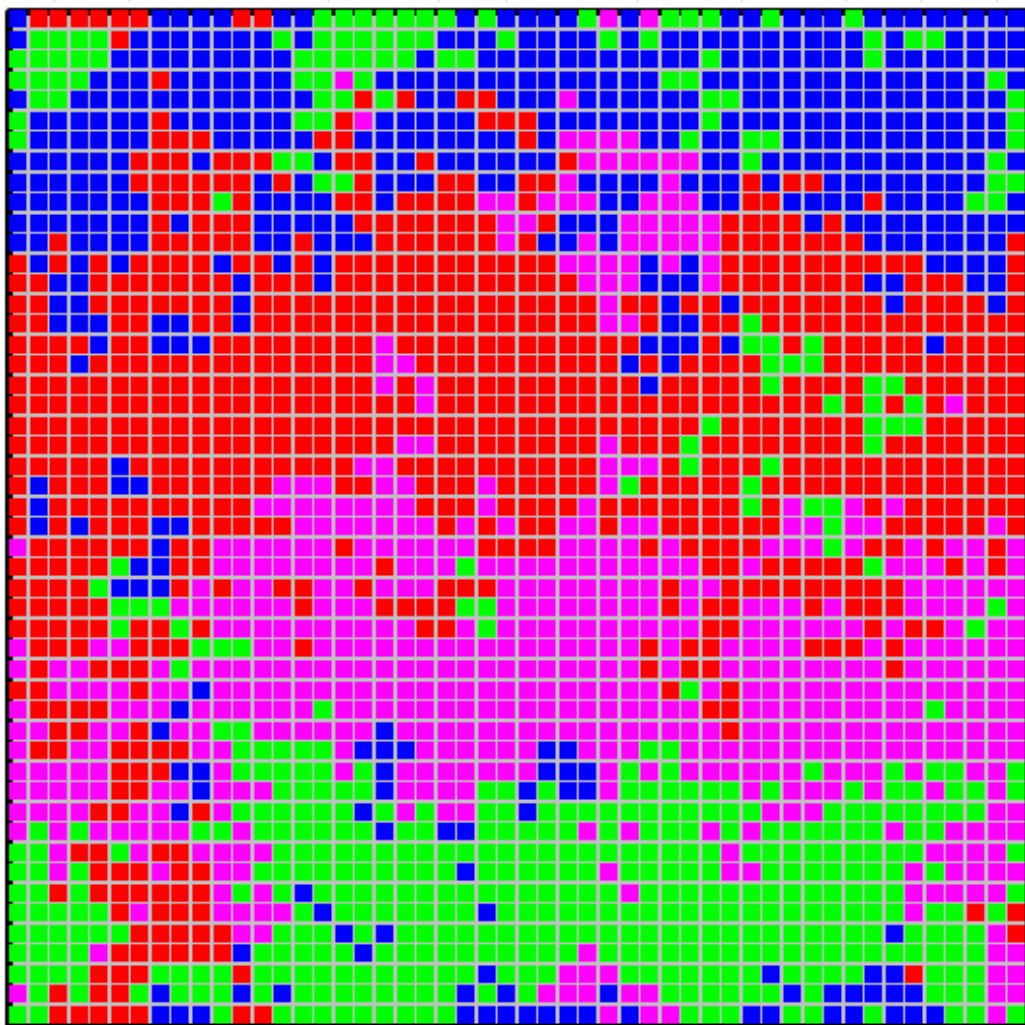
Time = 150

Habitat selection =

0.1

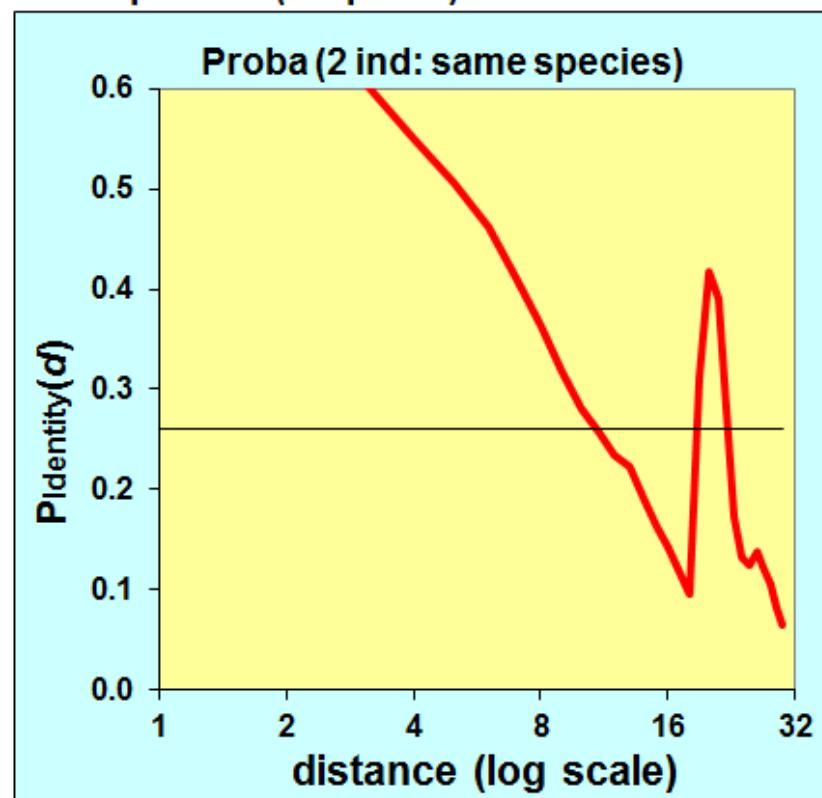
R(sp/envi)=

0.58



Time = 150

- species1 (freq=0.2)
- species2 (freq=0.34)
- species3 (freq=0.25)
- species4 (freq=0.21)



Filtre environnemental sur gradient ecol + dispersion étendue (état initial)

Mortality rate =

50%

Dispersal distance =

10 step

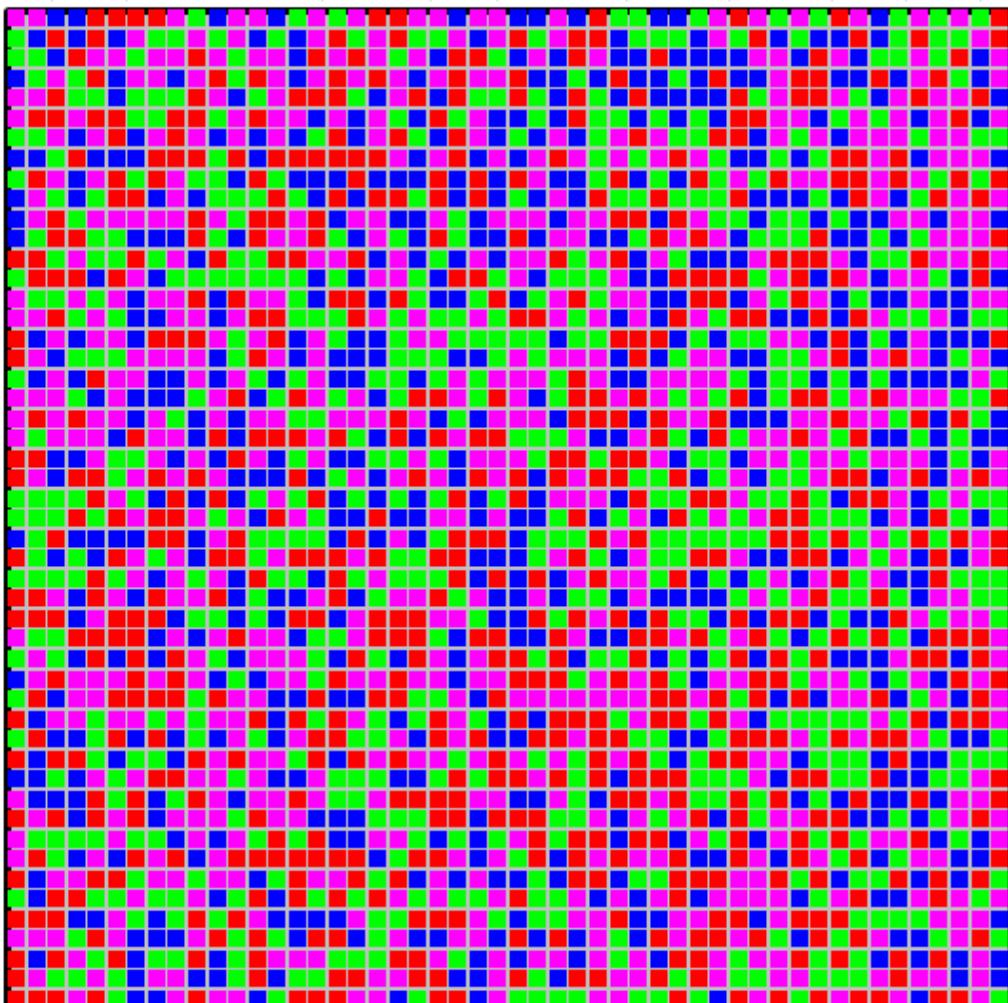
Time = 0

Habitat selection =

0.1

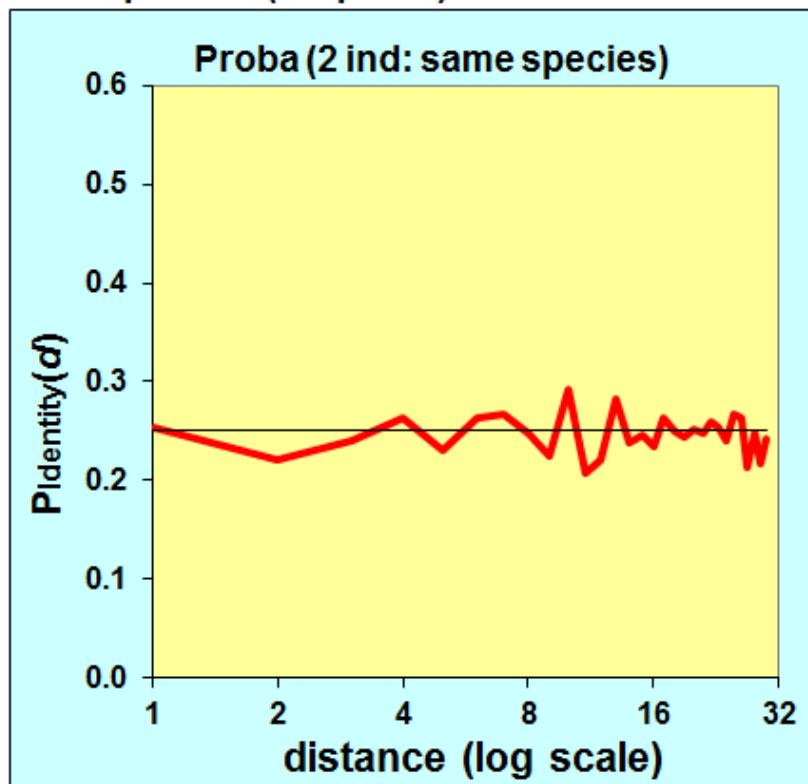
R(sp/envi)=

0.01



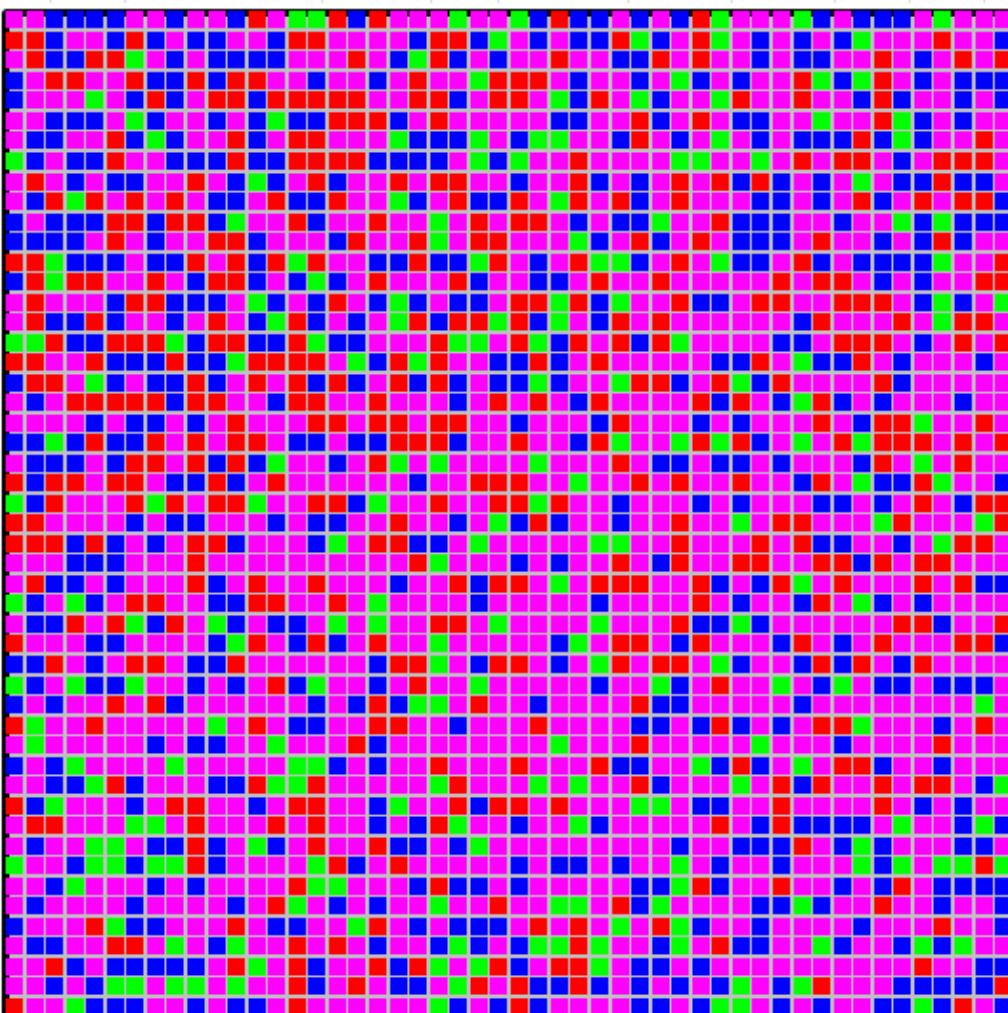
Time = 0

- species1 (freq=0.24)
- species2 (freq=0.26)
- species3 (freq=0.26)
- species4 (freq=0.24)

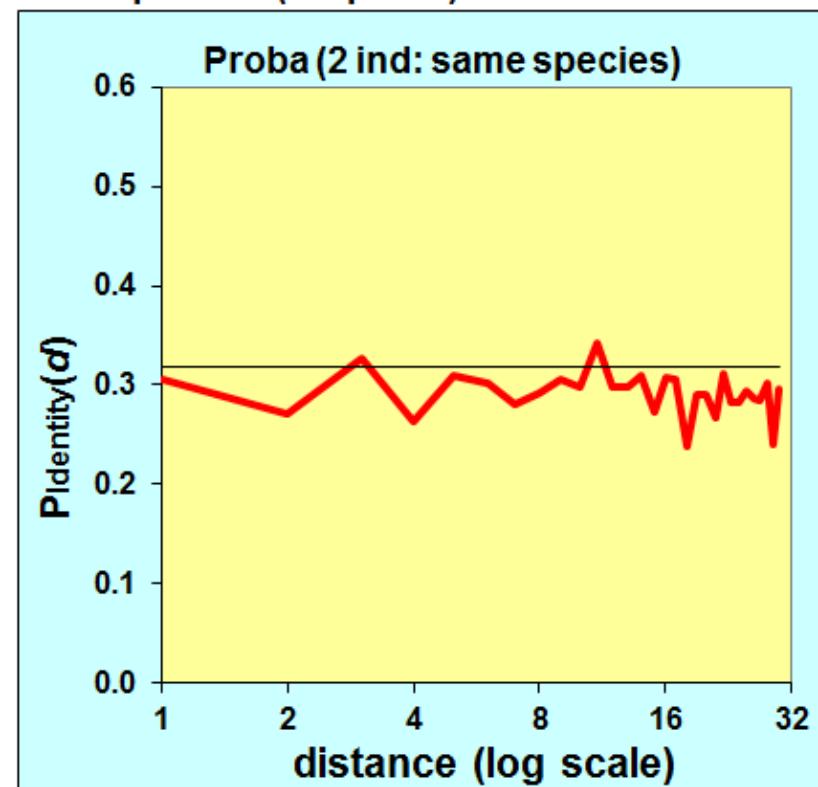


Filtre environnemental sur gradient ecol + dispersion étendue (100 pas de temps)

Mortality rate = 50% Dispersal distance = 10 step Time = 100
Habitat selection = 0.1 R(sp/envi)= 0.1



Time = 100
■ species1 (freq=0.25)
■ species2 (freq=0.2)
■ species3 (freq=0.45)
■ species4 (freq=0.09)



Properties of Similarity Indices under Niche-Based and Dispersal-Based Processes in Communities

Maxime Réjou-Méchain^{1,*} and Olivier J. Hardy²

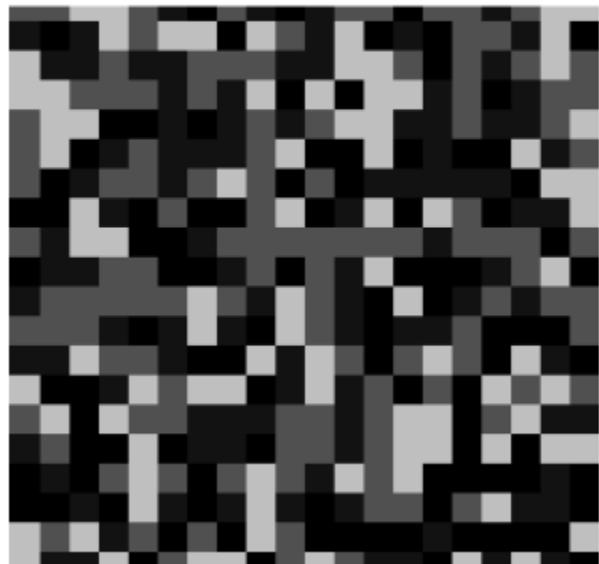
Am Nat 2011

Grid of local communities each with $N=100$
 4 habitats at different heterog. grain (H)
 0.001 immigration from constant source
 20-500 species, log-series distrib.
 Dispersal=expon. (mean dist $\delta=0.5, 1.5, \dots$)
 Neutral or selection (1 optimal hab/sp)

$H=1$

$H=4$

$H=8$



Measures of species turnover in space using pairwise (dis)similarity indices btw comm.

$$\text{Sorensen} = \frac{2a}{2a + b + c}$$

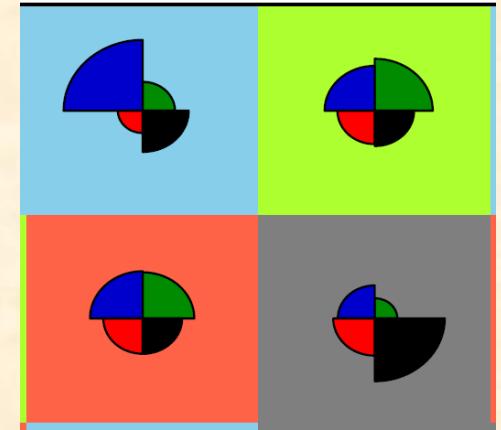
$$\text{Morisita-Horn} = \frac{\sum_s p_{is} p_{js}}{(\sum_s p_{is}^2 + \sum_s p_{js}^2)/2}$$

$$F = \sum_s p_{is} p_{js} \quad \text{within vs between hab if selection}$$

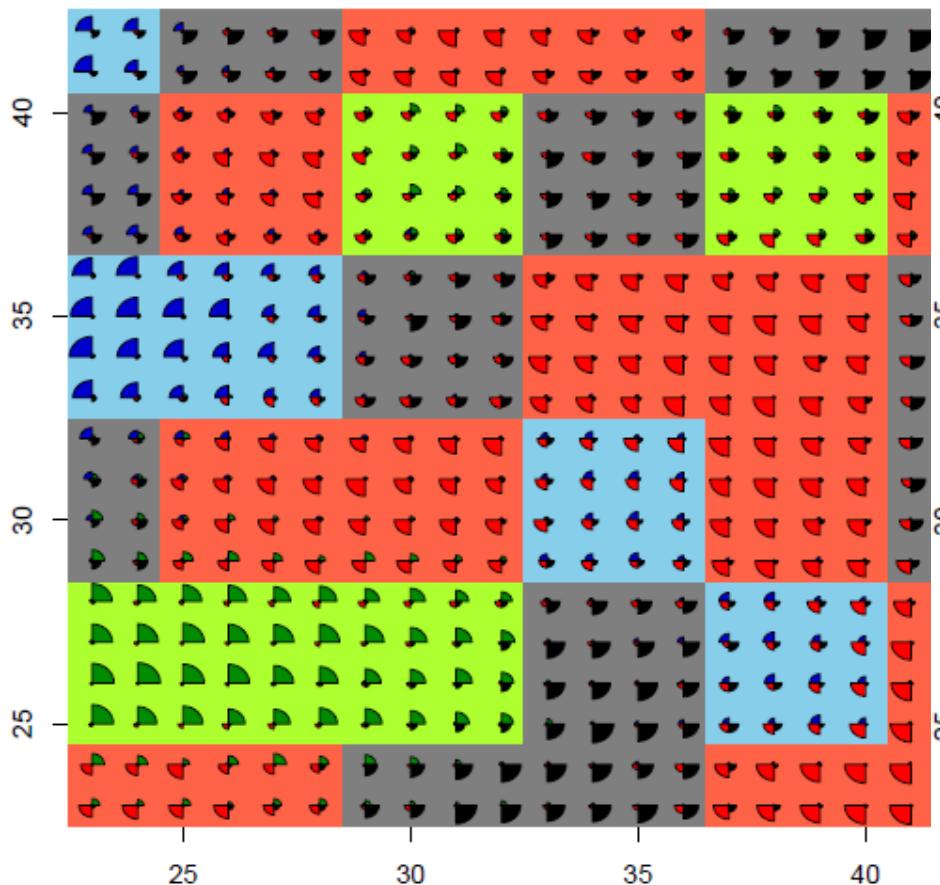
$$I_{ST} = 1 - \frac{1 - (\sum_s p_{is}^2 + \sum_s p_{js}^2)/2}{1 - (\sum_s p_{is} p_{js})}$$

Selection ($s=0.3$) : efficiency of selection

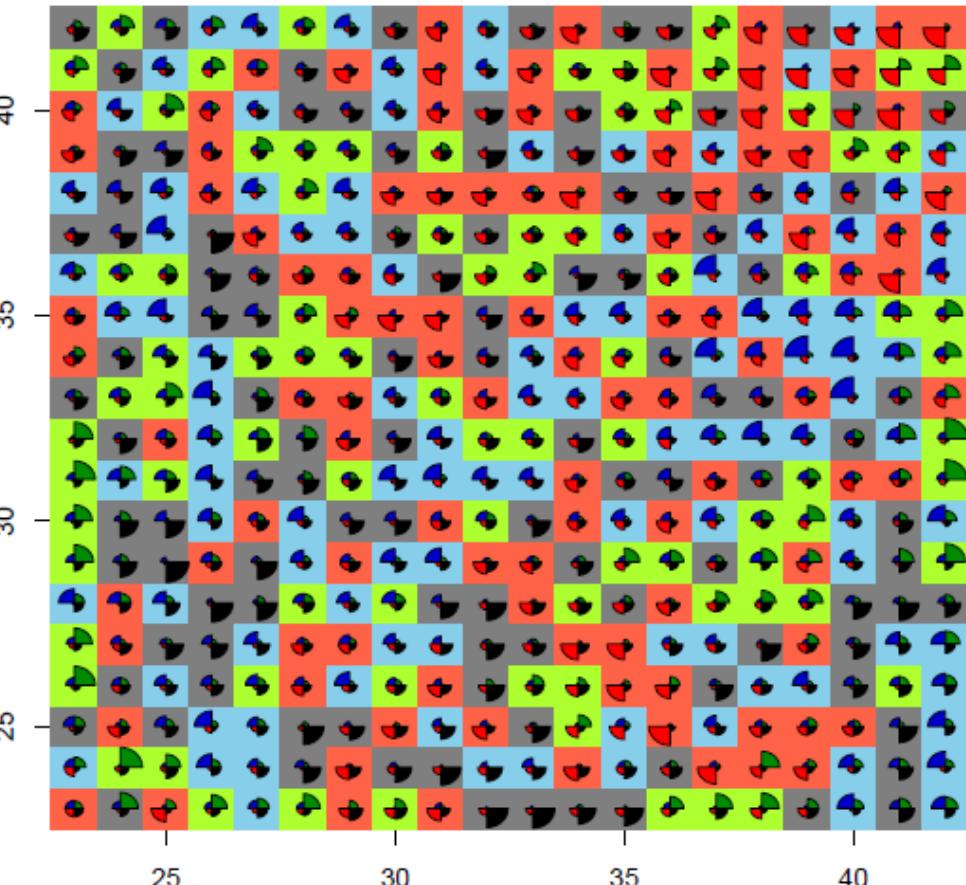
Mean dispersal distance = 1.5 grid units



Habitat H4 delta 1.5



Habitat H1 delta 1.5



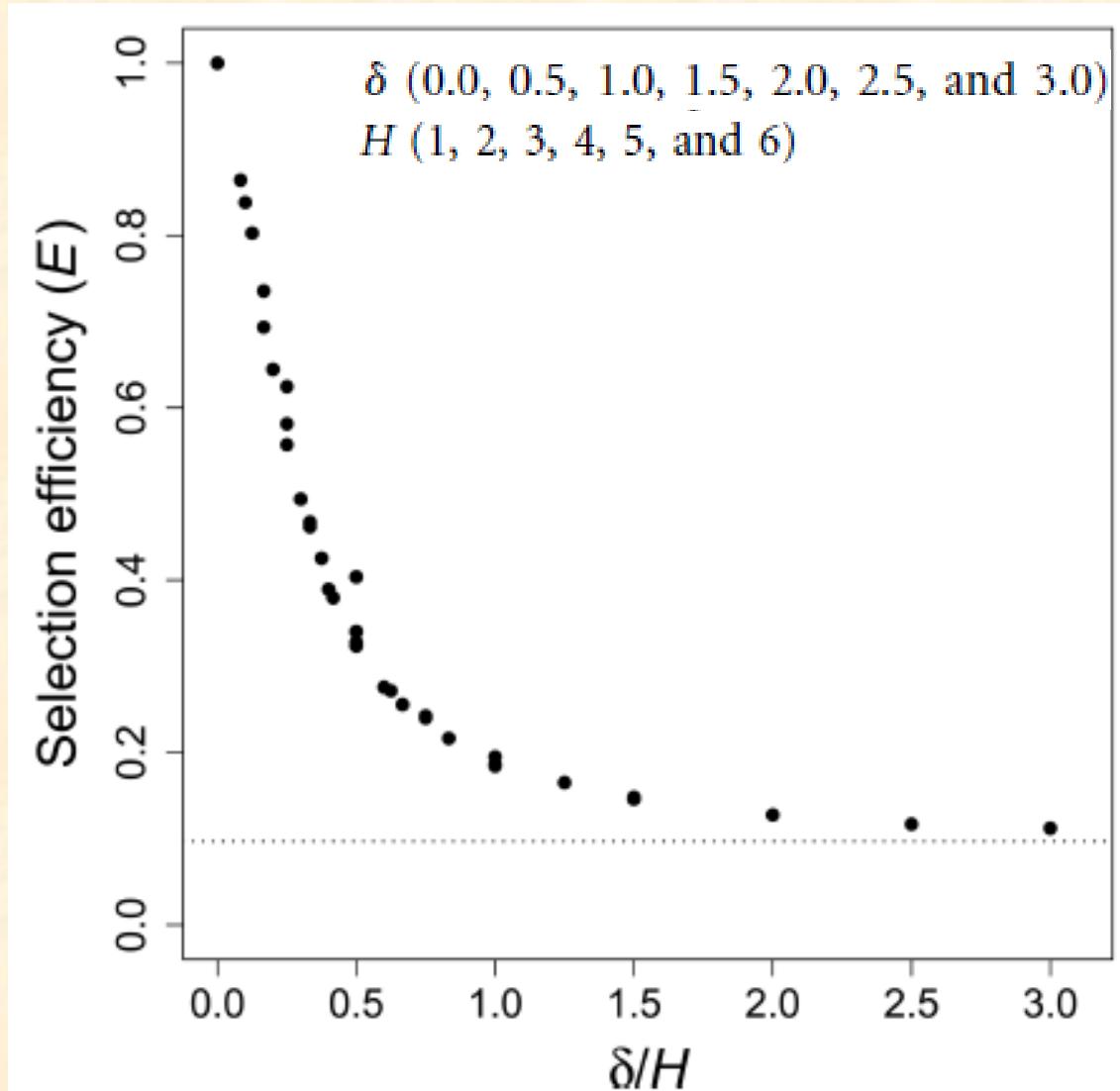
Selection : efficiency of selection

according to ratio dispersal distance / grain environm heterog

$$E = \frac{P_o - P_e}{1 - P_e}$$

P_o = observed freq. of ind.
in their optimal habitat

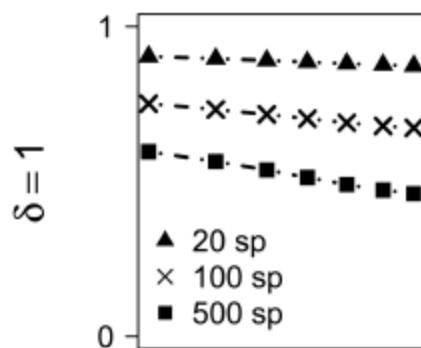
P_e = expected freq. of ind.
in their optimal habitat
under random distribution



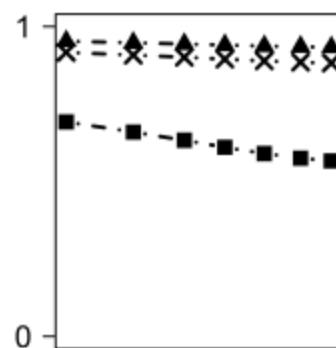
Species turnover: effect of neutral/selection + species richness on different indices

neutral
Little efficient selection
Efficient selection

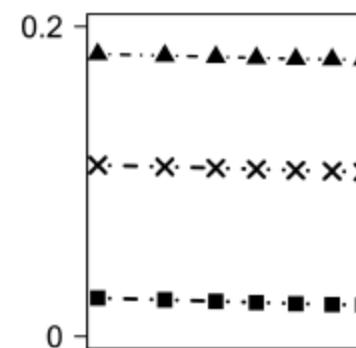
Sorensen



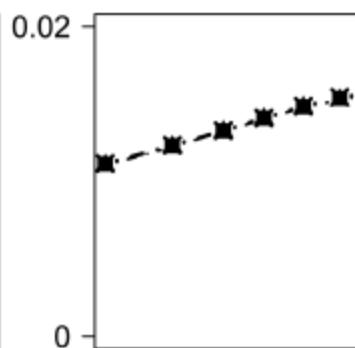
Morisita-Horn



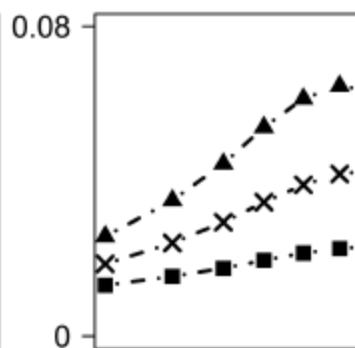
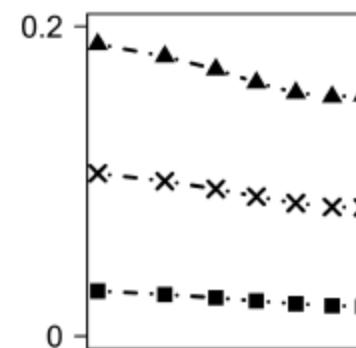
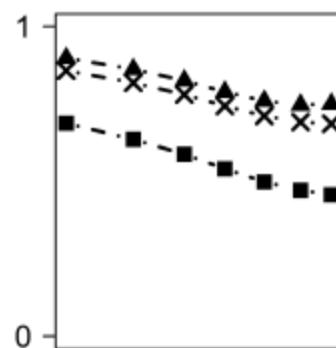
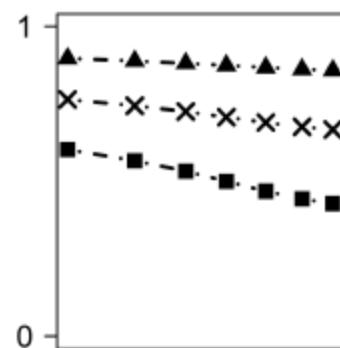
F



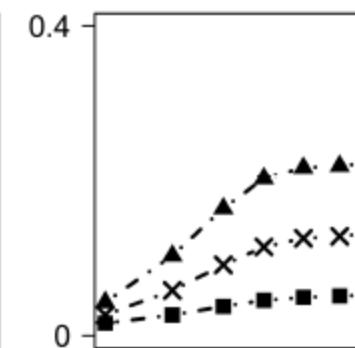
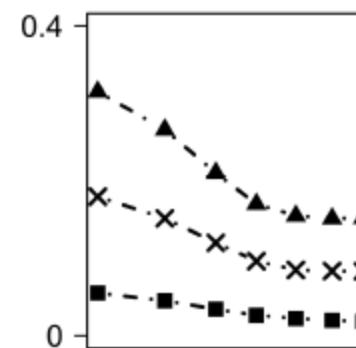
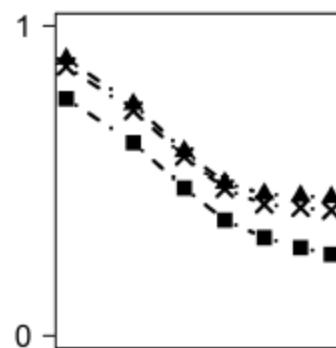
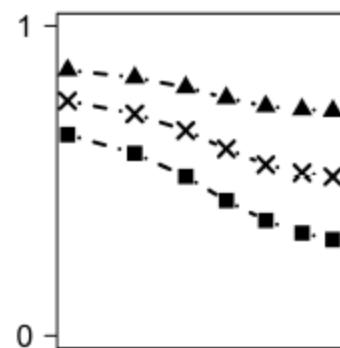
Ist



$E=0$



$E=0.19$



$E=0.58$

Geographical distance (log scale)

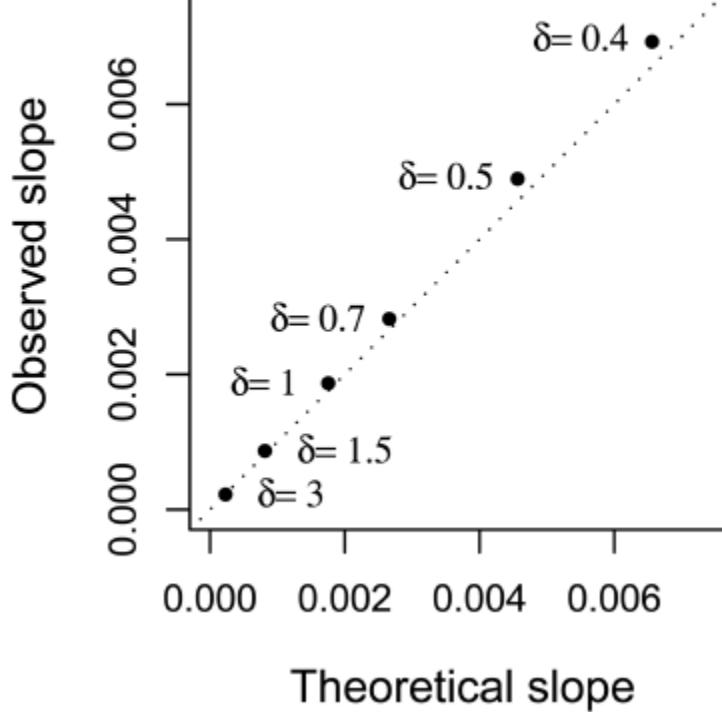
Neutral case: verification of isolation by distance prediction

Slope $Ist/(1-Ist)$ on $\ln(\text{dist}) = (2\pi D\sigma^2)^{-1}$

Rousset 1997

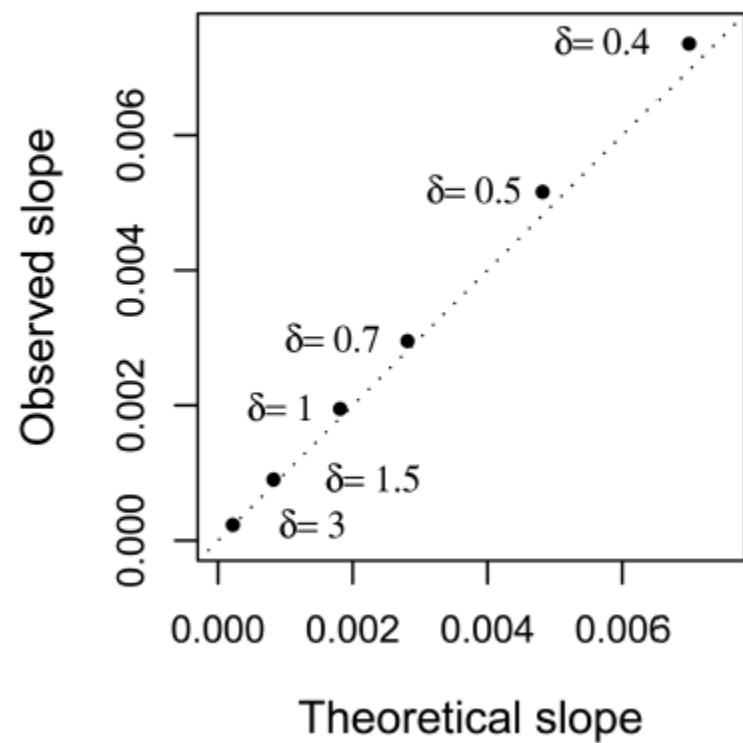
a)

F

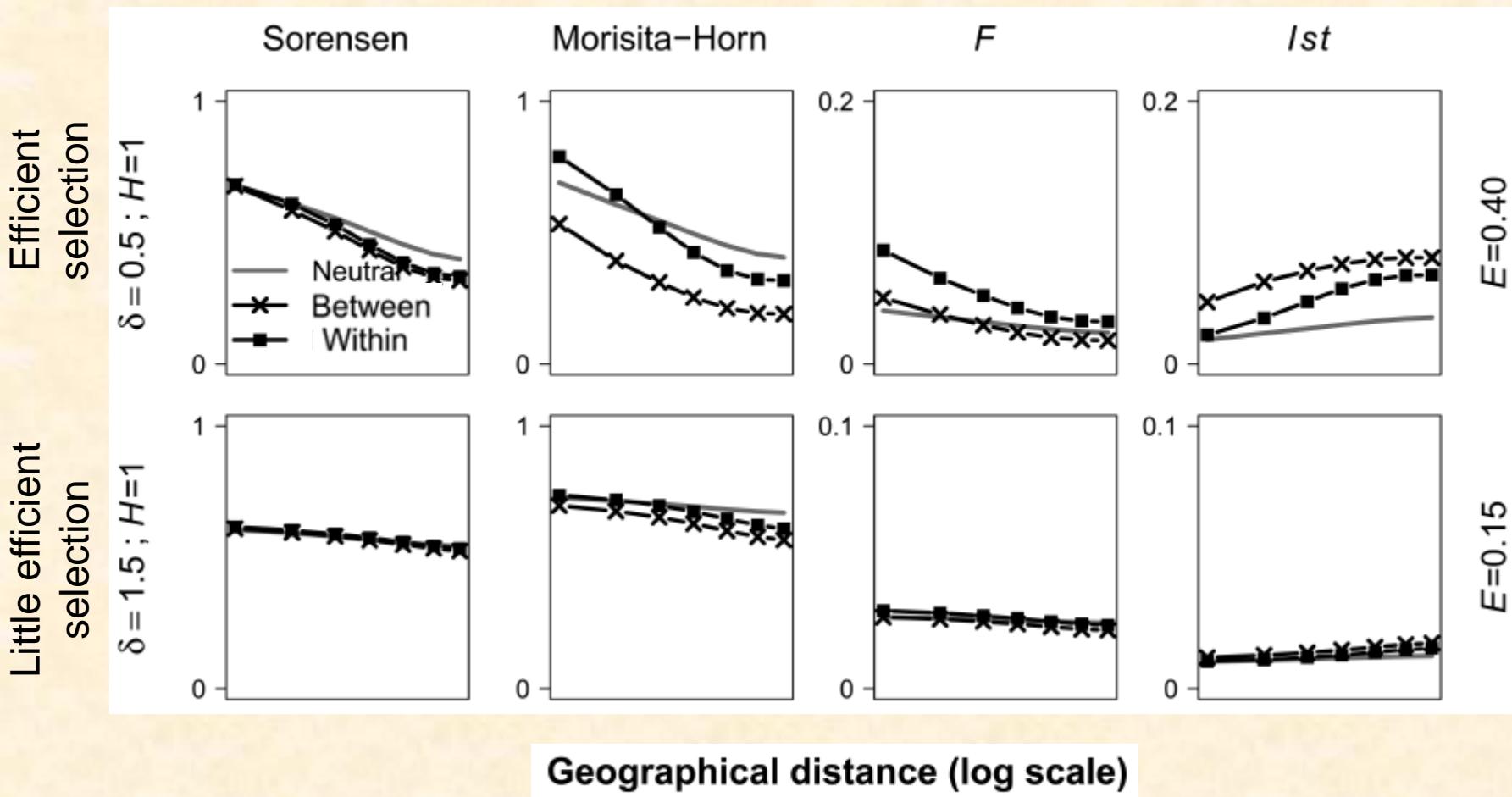


b)

Ist



Species turnover within versus between habitat types under selection



Modèle de communauté neutre (Hubbell 2001)

-> toutes les espèces sont équivalentes;
la notion d'espèce se réduit à une étiquette héritable

Modélisation d'une communauté neutre

+ capacités de dispersion limitées dans l'espace:

-> des espèces peuvent coexister longtemps malgré l'absence de différenciation de niche

-> des structures spatiales apparaissent

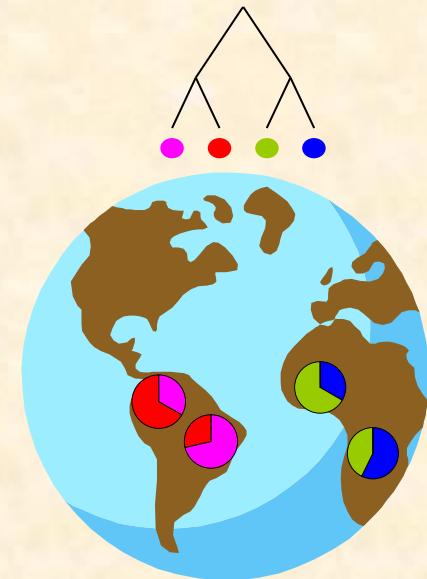
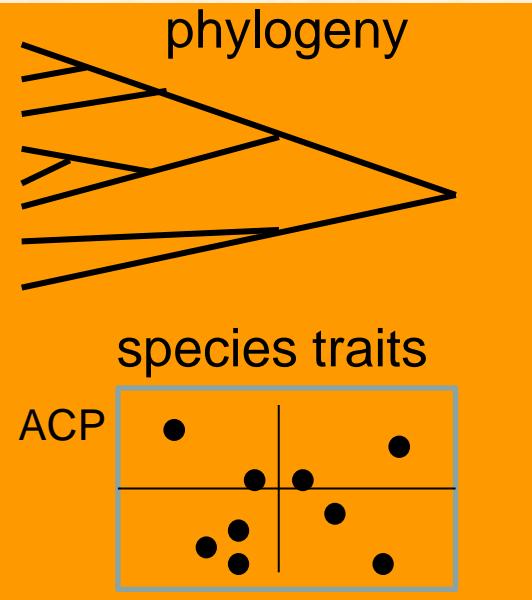
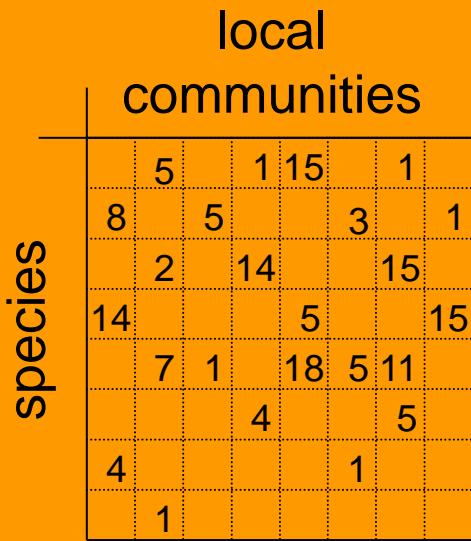
↳ effet confondant avec l'impact d'une hétérogénéité du milieu

Modélisation avec différenciation de niche:

-> la dispersion limitée peut accroître la réponse des espèces à l'hétérogénéité du milieu

Evolutionary aspects of community assembly

Insights on species assembly rules from the phylogenetic and functional structure of tropical plant communities



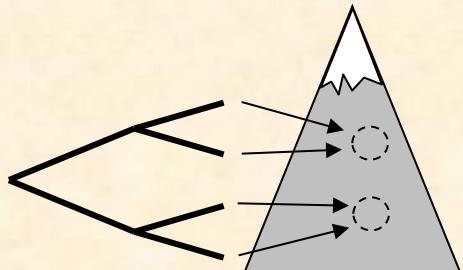
⇒ test hypotheses on mechanisms determining species assembly

Community phylogenetic structure

Example under environmental filtering (adaptation to altitude)

Phylogenetic clustering

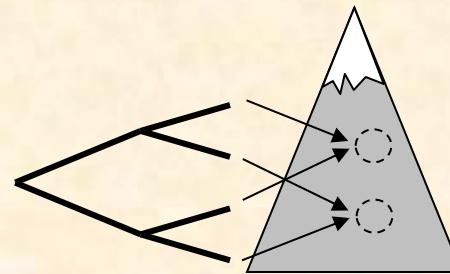
species within habitat **more** related than among habitats



habitat conservatism

Phylogenetic overdispersion

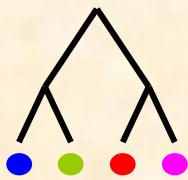
species within habitat **less** related than among habitats



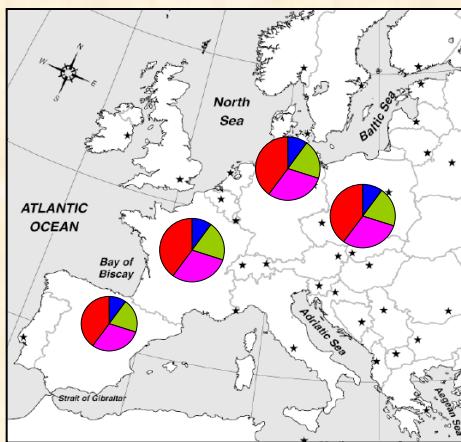
habitat convergence

e.g. *radiation of several clades in a set of new habitats*

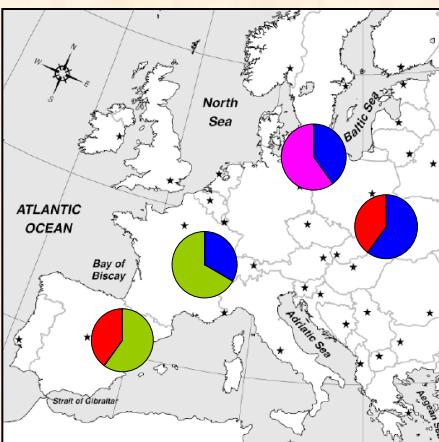
Species turnover versus phylogenetic turnover



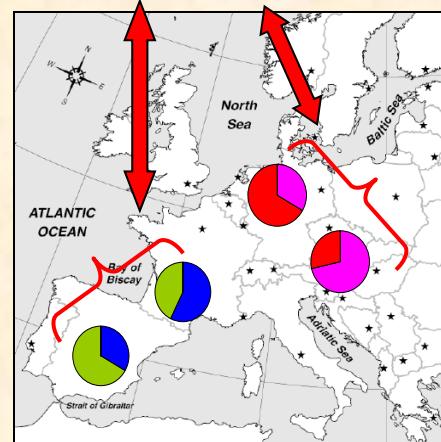
Phylogenetic tree
for 4 species



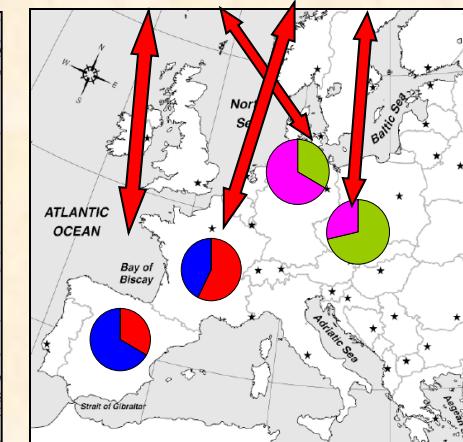
No species
turnover
&
No phylogenetic
turnover



Species turnover
but
No phylogenetic
turnover



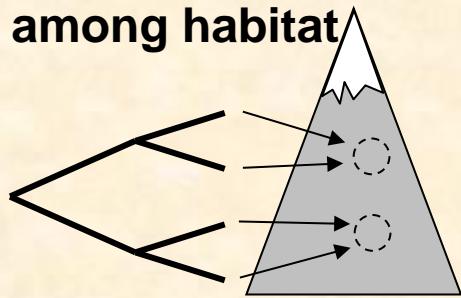
Species turnover
+
Phylogenetic
clustering
(attraction)



Species turnover
+
Phylogenetic
overdispersion
(evenness, repulsion)

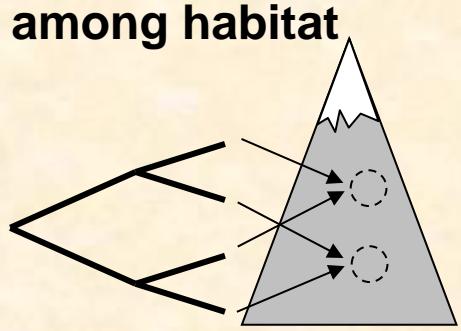
Phylogenetic turnover caused by ecological & evolutionary processes

Phylogenetic clustering



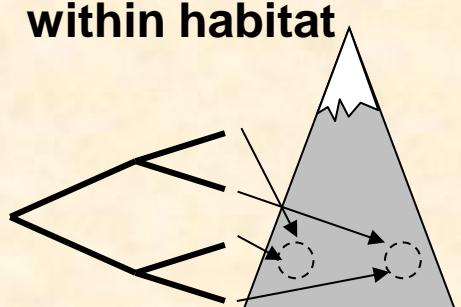
environmental filtering
adaptation to ≠ elevation
+
habitat conservatism

Phylogenetic overdispersion
(= evenness = repulsion)



environmental filtering
adaptation to ≠ elevation
+
habitat convergence
e.g. *radiation of several clades in a set of new habitats*

Phylogenetic overdispersion
(= evenness = repulsion)



competitive exclusion
+
niche conservatism
→ sister species cannot coexist because of niche overlap ('limiting similarity' principle)

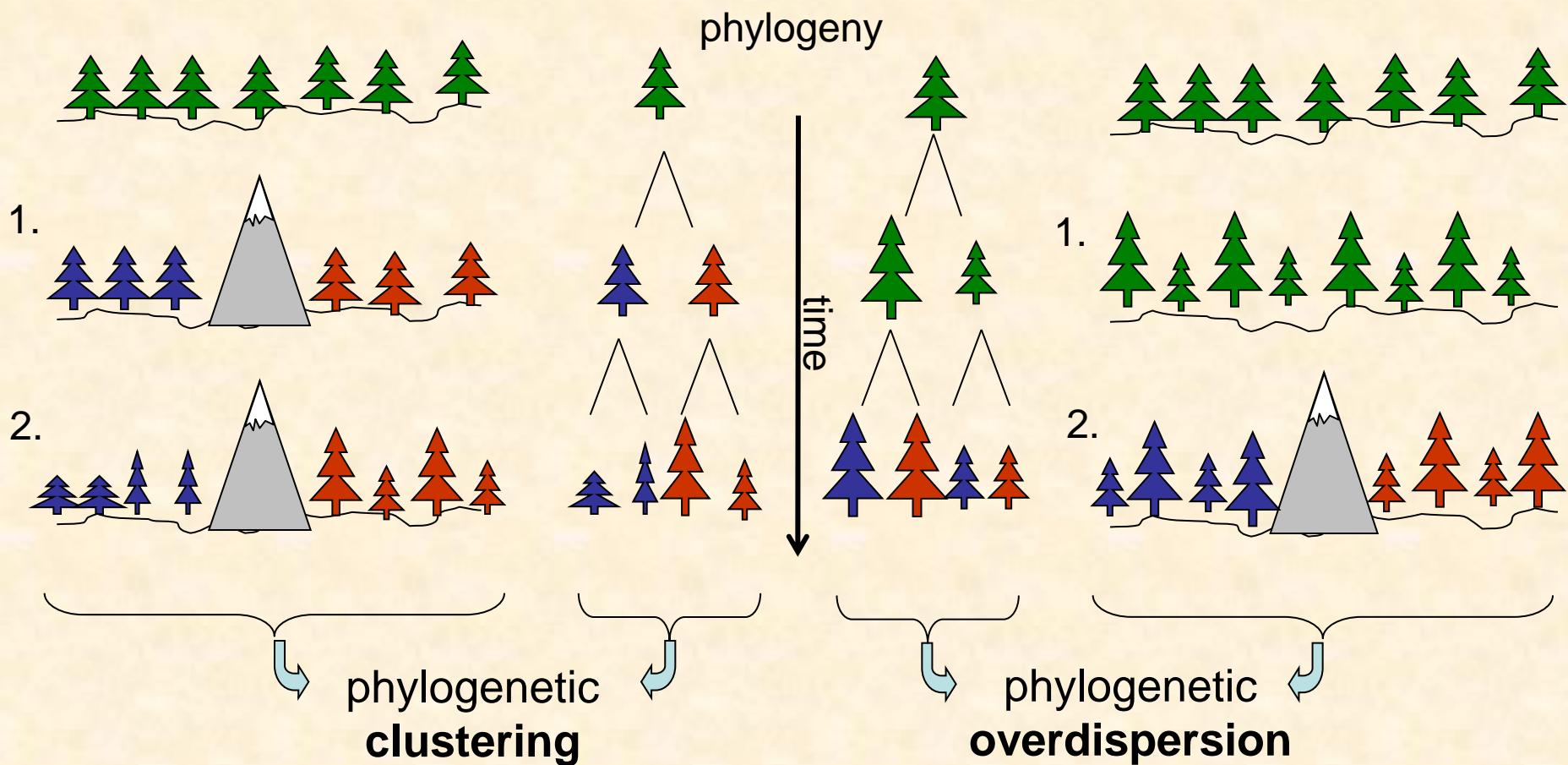
Phylogenetic turnover caused by biogeographic barrier

Scenario A:

1. barrier (allopatric speciation)
2. sympatric speciations

Scenario B:

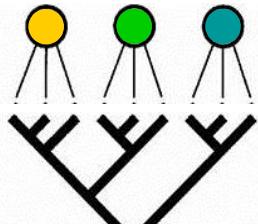
1. sympatric speciation
2. barrier (allopatric speciations)



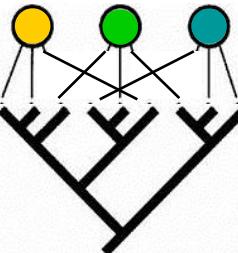
The link between Community assembly and Phylogeny: species Traits

C x P

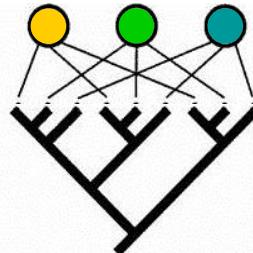
Phylogenetic clustering



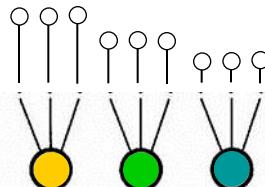
Random phylogenetic structure



Phylogenetic overdispersion

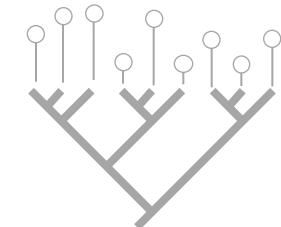


T x P

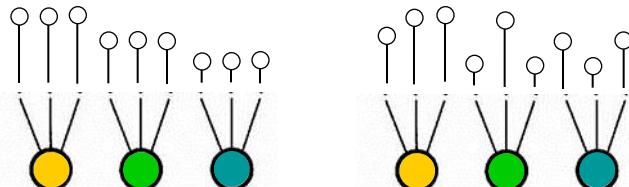


T x P

Trait conservatism



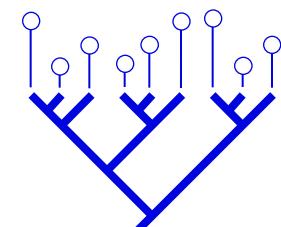
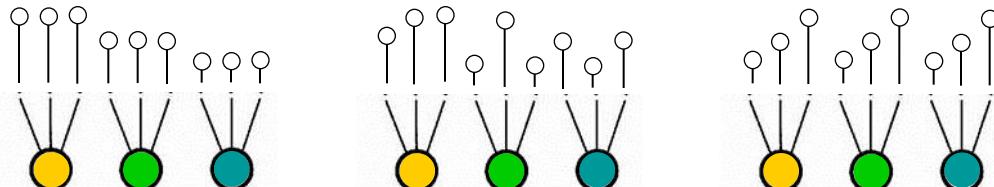
No phylogenetic signal



Random assembly (neutrality)

T x C Environmental filtering

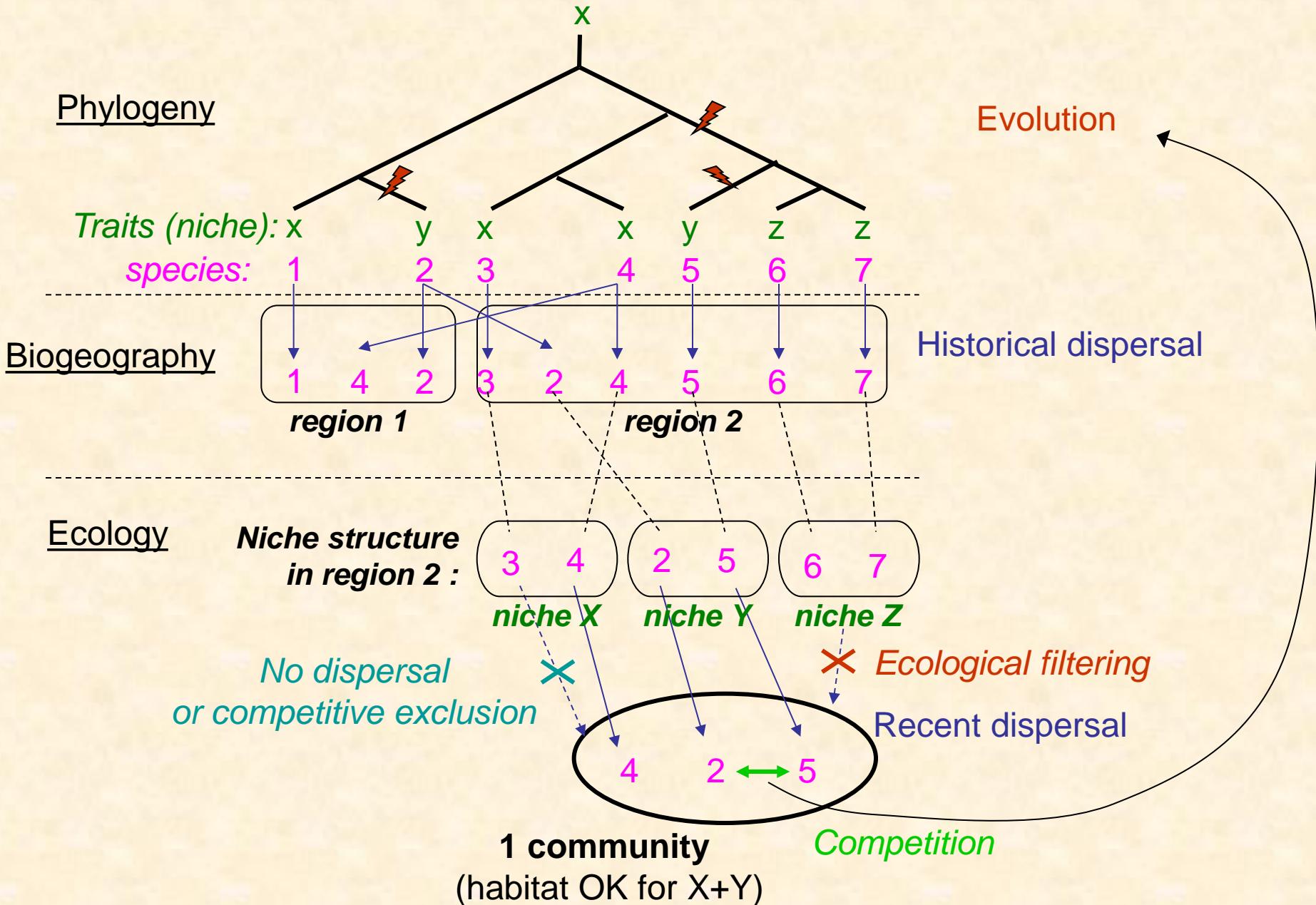
Competitive exclusion



Trait convergence

Community phylogenetic structure depends on ecological and evolutionary processes

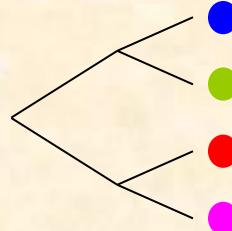
Integration communities – traits – ecology – biogeography – evolution



How to quantify phylogenetic turnover from species incidence data ?

↳ many methods developed; here just one presented

1. From a phylogeny → phyletic distances between species



$$\delta_{kl}$$

| | | | | |
|---|---|---|---|---|
| ● | 0 | 1 | 2 | 2 |
| ● | 1 | 0 | 2 | 2 |
| ● | 2 | 2 | 0 | 1 |
| ● | 2 | 2 | 1 | 0 |

→ divergence time
(molecular phylogeny)

→ (taxonomic level)

2. From community inventories → mean δ_{kl} between species

sampled

- within sites
- among sites

$$MPD_w = \bar{\delta}_{(2_species_from_same_site)}$$

$$MPD_a = \bar{\delta}_{(2_species_from_different_sites)}$$

$$\rightarrow \Pi_{ST} = 1 - MPD_w / MPD_a$$

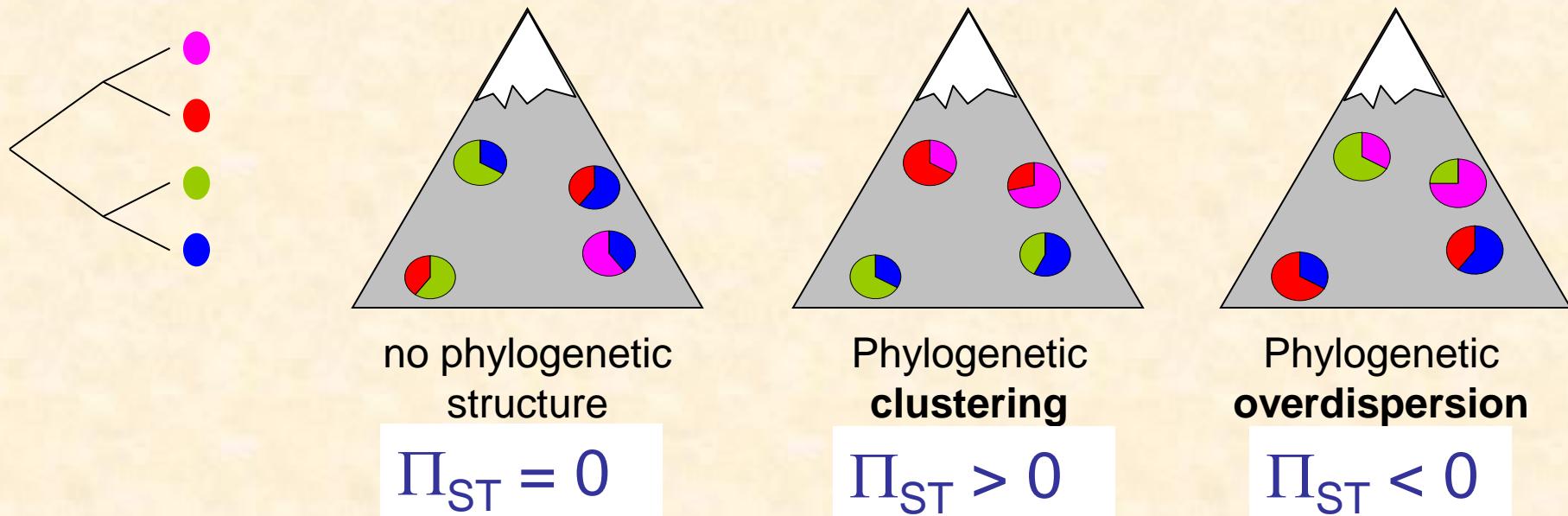
relative excess of relatedness between species co-occurring within a site (relative to species pairs sampled in distinct sites)

Hardy & Senterre 2007 (J. Ecol. 95: 493–506)

Hardy & Jost 2008 (J. Ecol. 96: 849–852)

-> phylogenetic turnover

How to test phylogenetic turnover?

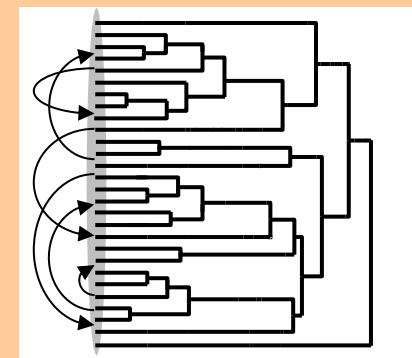


Testing $\Pi_{ST} \neq 0$

→ **randomize phylogeny**

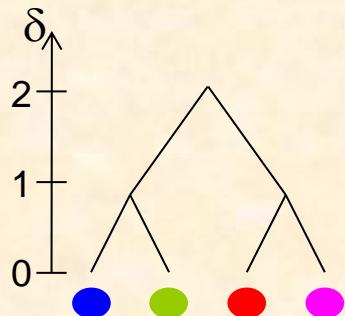
→ assess distribution of Π_{ST} for 1000 permutations

→ P-value = % of Π_{ST} after permutation > or < Π_{ST} observed



More elaborated randomization tests: Hardy 2008 (J Ecol 96: 914–926)

Exemple of computation of pairwise Π_{ST} for simple cases



$$MPD_w = \bar{\delta}_{(2 \neq \text{species_from_same_site})}$$

$$MPD_a = \bar{\delta}_{(2 \neq \text{species_from_different_sites})}$$

$$\Pi_{ST} = 1 - MPD_w / MPD_a$$

| | MPD_w | 1 | 2 |
|------------|---------|-------|-------|
| MPD_w | | | |
| 2 | | $5/3$ | $3/2$ |
| 1 | | 2 | $5/3$ |
| $5/3=1.67$ | | $5/3$ | $5/3$ |

| Π_{ST} | | |
|------------|---------------------|-----------------------|
| | $1-(3/2)/(5/3)=0.1$ | $1-2/(3/2)=-0.33$ |
| | $1-1/2=0.5$ | $1-(3/2)/(5/3)=0.1$ |
| | $1-(4/3)/(5/3)=0.2$ | $1-(11/6)/(5/3)=-0.1$ |

How to quantify species + phylogenetic turnover from species abundance data ? (using QE: Rao's quadratic entropy)

From a phylogeny → distances between species



$$\delta_{kl}$$

| | | | |
|---|---|---|---|
| ● | ● | ● | ● |
| ● | 0 | 1 | 2 |
| ● | 1 | 0 | 2 |
| ● | 2 | 2 | 0 |
| ● | 2 | 2 | 1 |

From community inventories:

> partition of QE = mean δ_{kl} between individuals

(Hardy & Senterre 2007,

sampled

within sites

among sites

$$QE_w = \bar{\delta}_{(2_individuals_from_same_site)}$$

$$QE_a = \bar{\delta}_{(2_individuals_from_different_sites)}$$

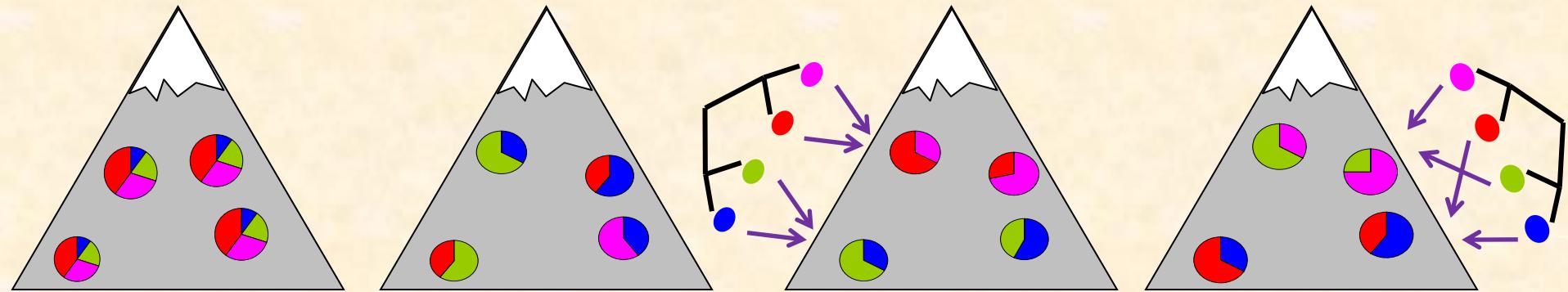
$$\rightarrow P_{ST} = 1 - QE_w / QE_a \quad \text{species + phylogenetic turnover}$$

> partition of Simpson's D = probability of species non-identity btw individuals

$$\rightarrow I_{ST} = 1 - D_w / D_a \quad \text{species turnover}$$

> partition of QE' = mean δ_{kl} between individuals of distinct species

$$\rightarrow B_{ST} = 1 - QE'_w / QE'_a \approx P_{ST} - I_{ST} \quad \text{phylogenetic turnover}$$



No species turnover
&
No phylogenetic turnover

$$P_{ST} = I_{ST} = 0 \\ B_{ST} = 0$$

Species turnover
but
No phylogenetic turnover

$$P_{ST} = I_{ST} > 0 \\ B_{ST} = 0$$

Species turnover
+
Phylogenetic clustering

$$P_{ST} > I_{ST} > 0 \\ B_{ST} > 0$$

Species turnover
+
Phylogenetic overdispersion

$$I_{ST} > P_{ST} > 0 \\ B_{ST} < 0$$

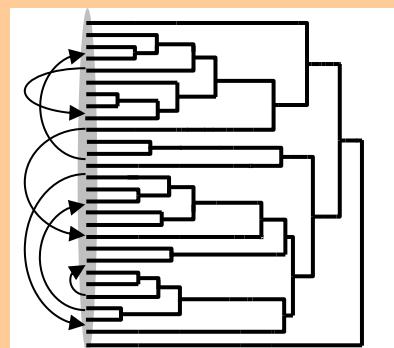
Testing $P_{ST} \leftrightarrow I_{ST}$ or $B_{ST} \leftrightarrow 0$

→ randomize phylogeny

→ assess distribution of P_{ST} for 1000 permutations

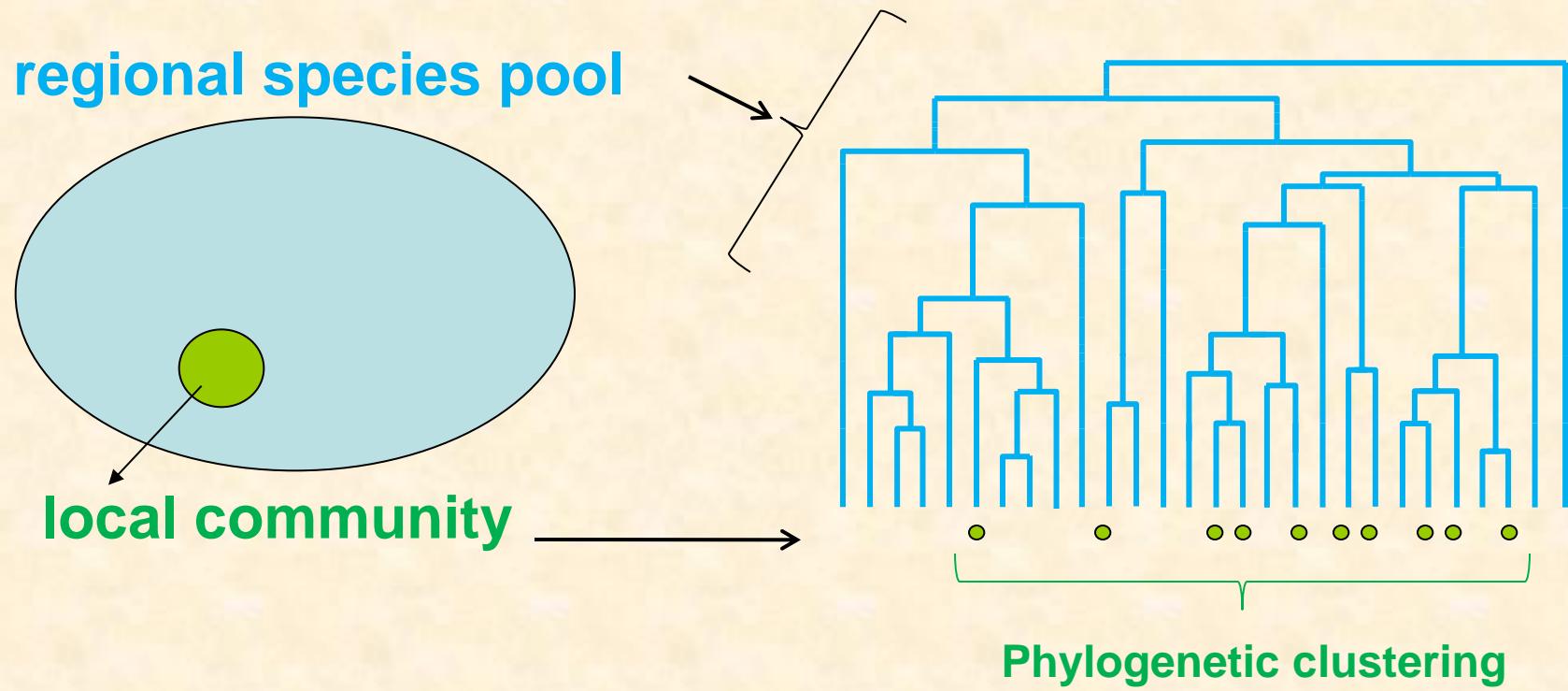
→ P-value = % of P_{ST} after permutation > or < P_{ST} observed

(for more elaborated randomization tests: Hardy 2008, J Ecol)



Community phylogenetic structure – the “species pool” approach (NRI, NTI indices)

C. Webb's approach (Webb 2000, Webb et al. 2002):



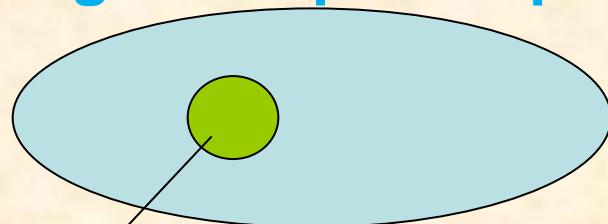
Are species in the local community
more / less related than random
species from the regional pool?

→ Compare the mean phylogenetic
distance (MPD) within local community
with distribution of MPD for ‘random’
communities generated by a null model

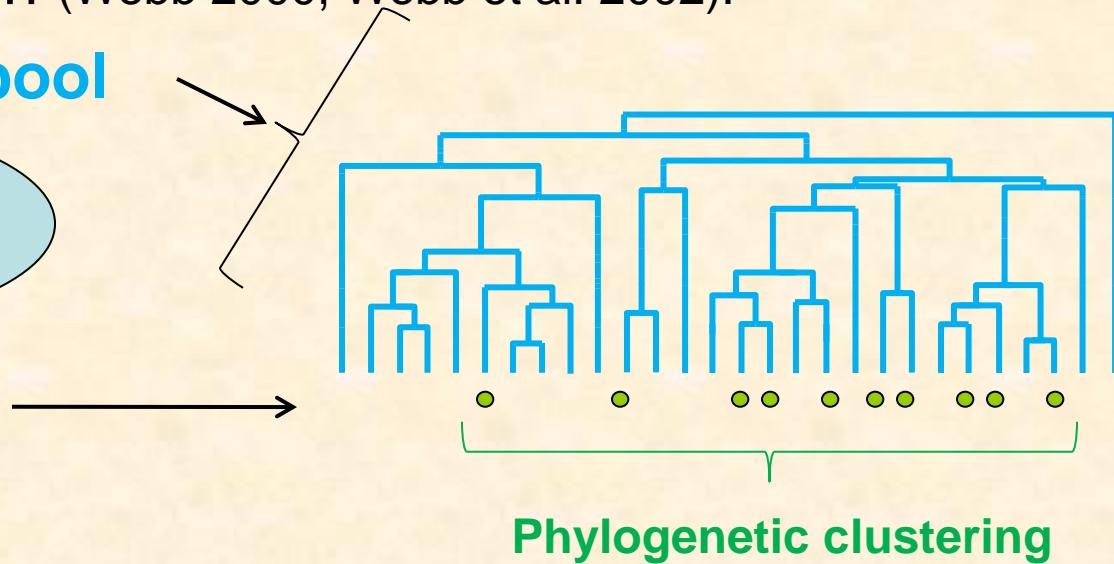
Community phylogenetic structure – the “species pool” approach (NRI, NTI indices)

C. Webb's approach (Webb 2000, Webb et al. 2002):

regional species pool



local community



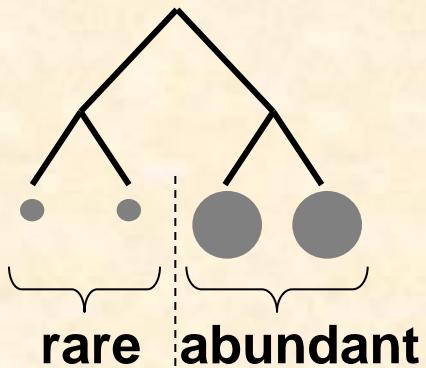
Net Relatedness Index:

$$NRI = -\frac{MPDw - \text{mean}(MPDrandomized)}{SD(MPDrandomized)}$$

Randomization: different null models are possible, most often it consists in random sampling from the species pool or in random permutations of the species across the phylogeny (both are equivalent when applied on a single local community).

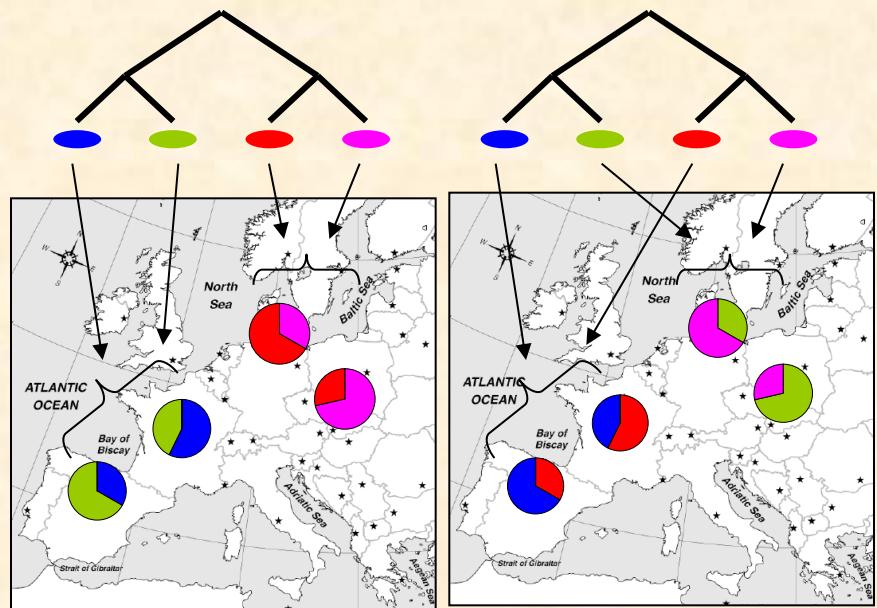
2 aspects of community phylogenetic structure (CPS)

species abundance CPS



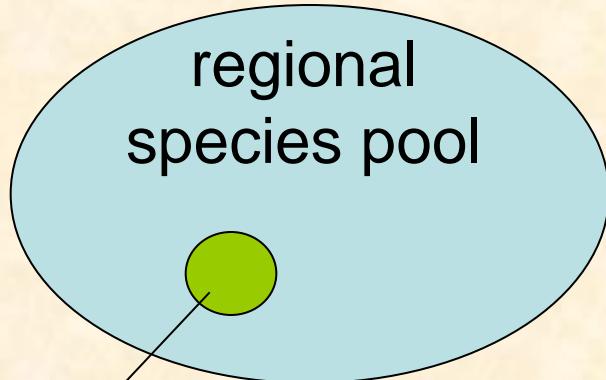
phylogenetic clustering

spatial CPS
= phylogenetic turnover



2 ways of investigating CPS

Species pool based



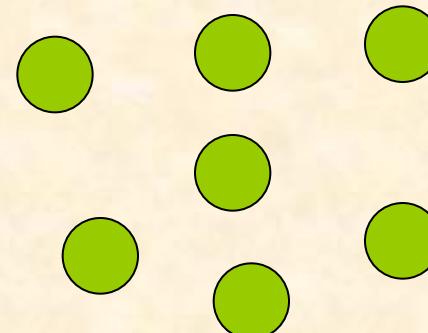
local community

Are species in the local community more / less related than random species from the regional pool?

Difficulties:

- How to delimit a species pool?
- How to account for differences in species abundances?
- How to distinguish species abundance from spatial CPS?

Sample based



set of local communities

Are species within local communities more / less related than species from distinct local communities?

- Describe only spatial CPS (phylogenetic turnover).
- Pattern specific for a given sample of local communities.
- Pairwise measures to regress on explanatory variables

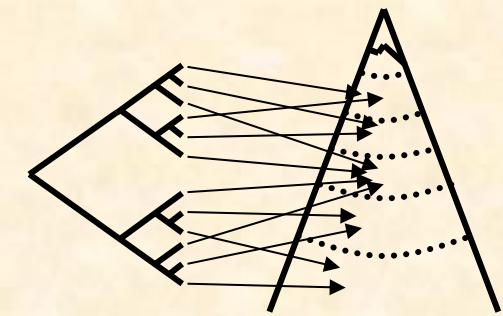
Data exploration

1. Pairwise analyses (between plots):

→ compute P_{ST} or Π_{ST} for each pair of plots

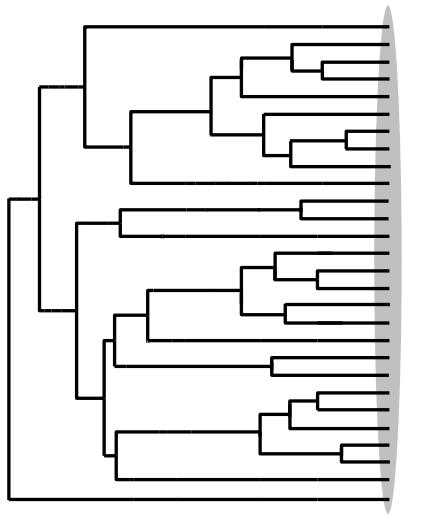
→ check correlations between P_{ST} or Π_{ST} and

{
geographic distance
ecological distance}

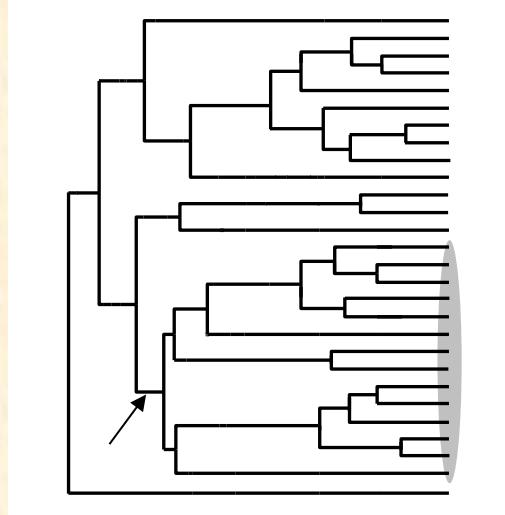


2. Partial tests

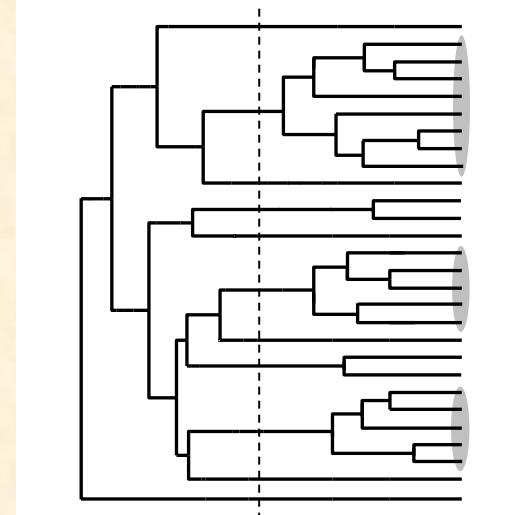
→ randomize only some parts of the phylogenetic tree



Complete
randomization



Randomiz.
within one clade

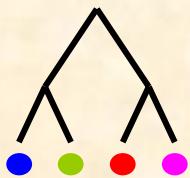


Randomiz. within clades
younger than age threshold

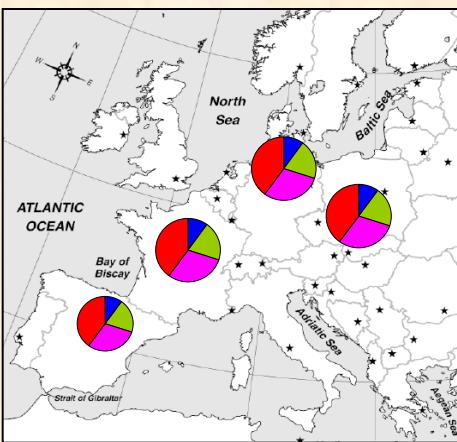
Case studies in tropical forest plant communities

1. Impact of **dispersal limitation versus ecological gradients** → species and phylogenetic turnover in Central African inselberg vegetation
2. Impact of **ecological gradients versus biogeography** → phylogenetic turnover along rainfall gradients in Western Ghats and Panama
3. Integrating **phylogeny** and **traits** → **phylogenetic versus functional** structures of rainforests from French Guiana
4. Phylogenetic **overdispersion** due to multiple recent radiations → forest communities from Reunion island
5. Phylogenetic structure at the scale of **inter-individual competition** → fine-scale spatial phylogenetic turnover in 50-ha plots

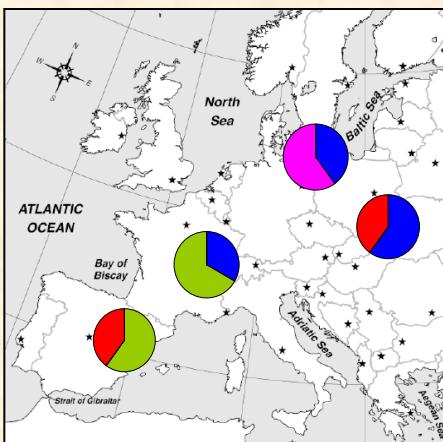
1. Impact of dispersal limitation vs ecological gradients on species turnover and phylogenetic turnover



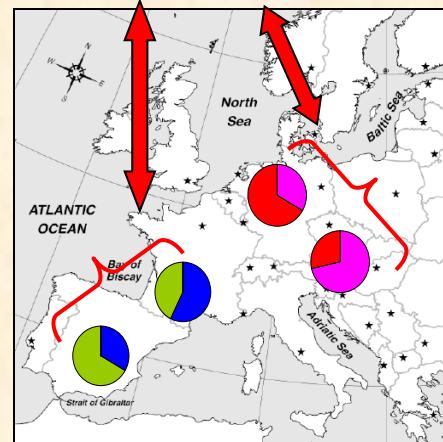
Phylogenetic tree
for 4 species



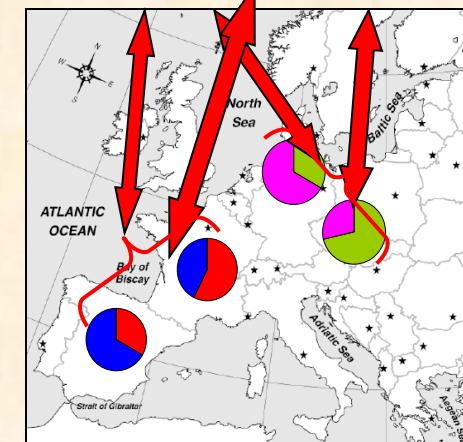
No species turnover
&
No phylo. turnover



Species turnover
but
No phylo. turnover



Species turnover
+
Phylo. clustering

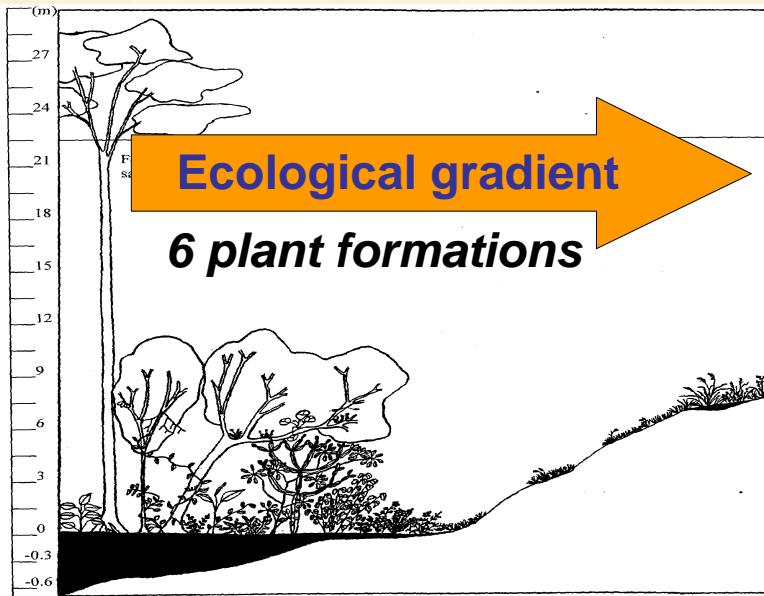


Species turnover
+
Phylo. overdisp.

Species turnover → Jaccard index of species similarity or I_{ST}

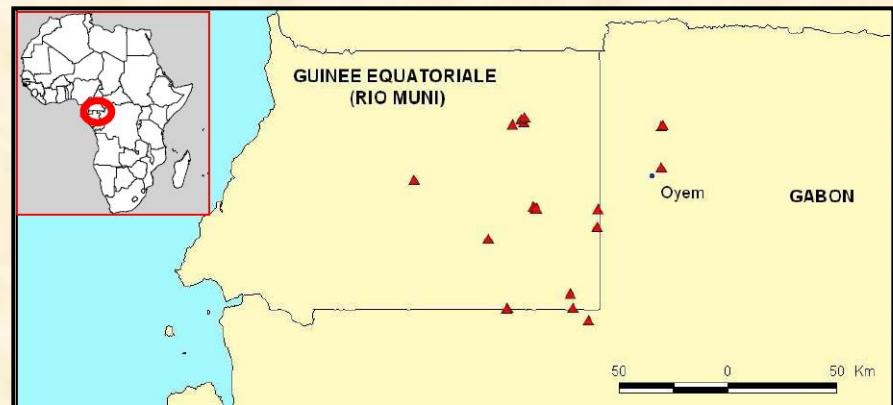
Phylogenetic turnover → Π_{ST} or P_{ST}

Inselberg vegetation in Equatorial Guinea



Geographical isolation

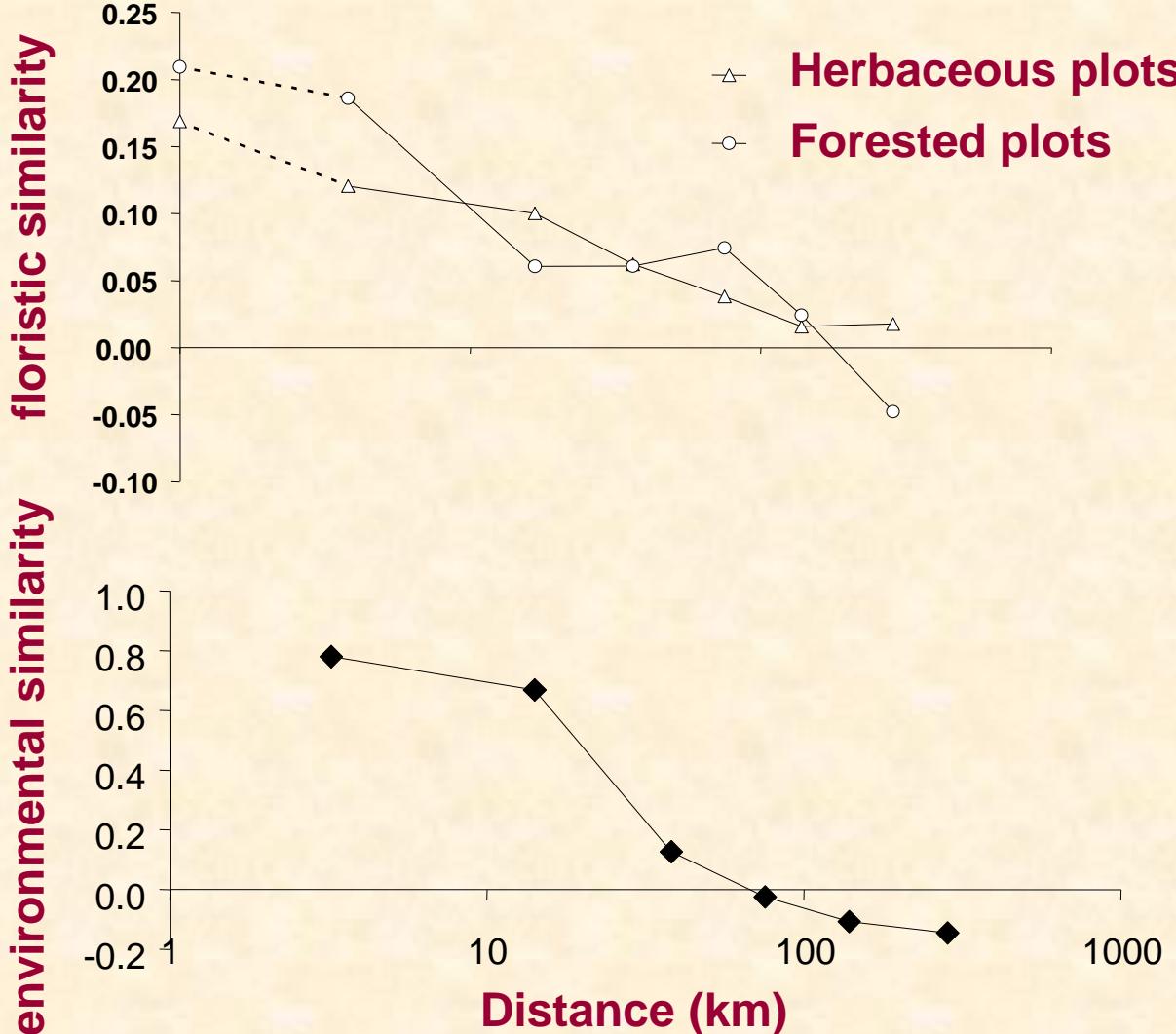
inselberg ↔ island



(ca. 300 floristic plots on 21 inselbergs)

Regional scale: subregional species pools

Spatial autocorrelation



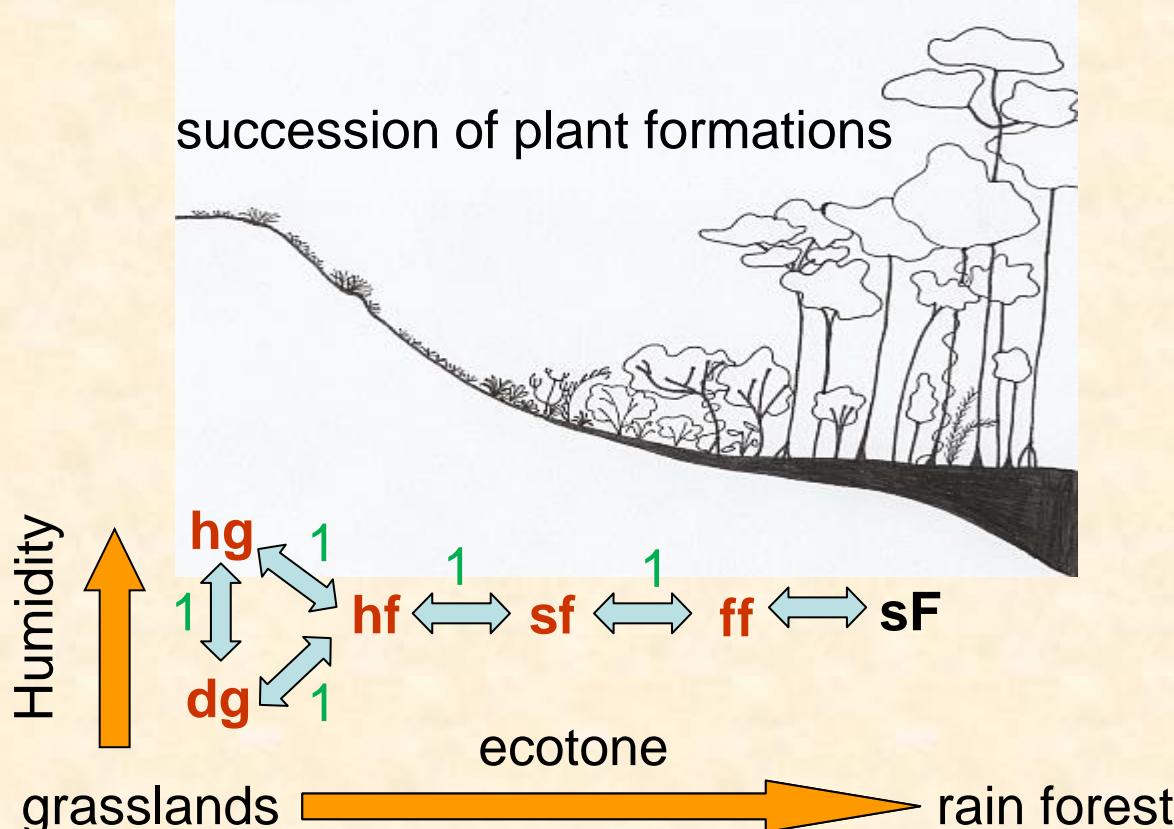
Species turnover due to:
Regional scale
deterministic factors
(climate...) ?
or
Dispersal limitation
(and/or colonization
history) ?

Parmentier *et al.* 2005
J Biogeography



Local scale: environmental constraints

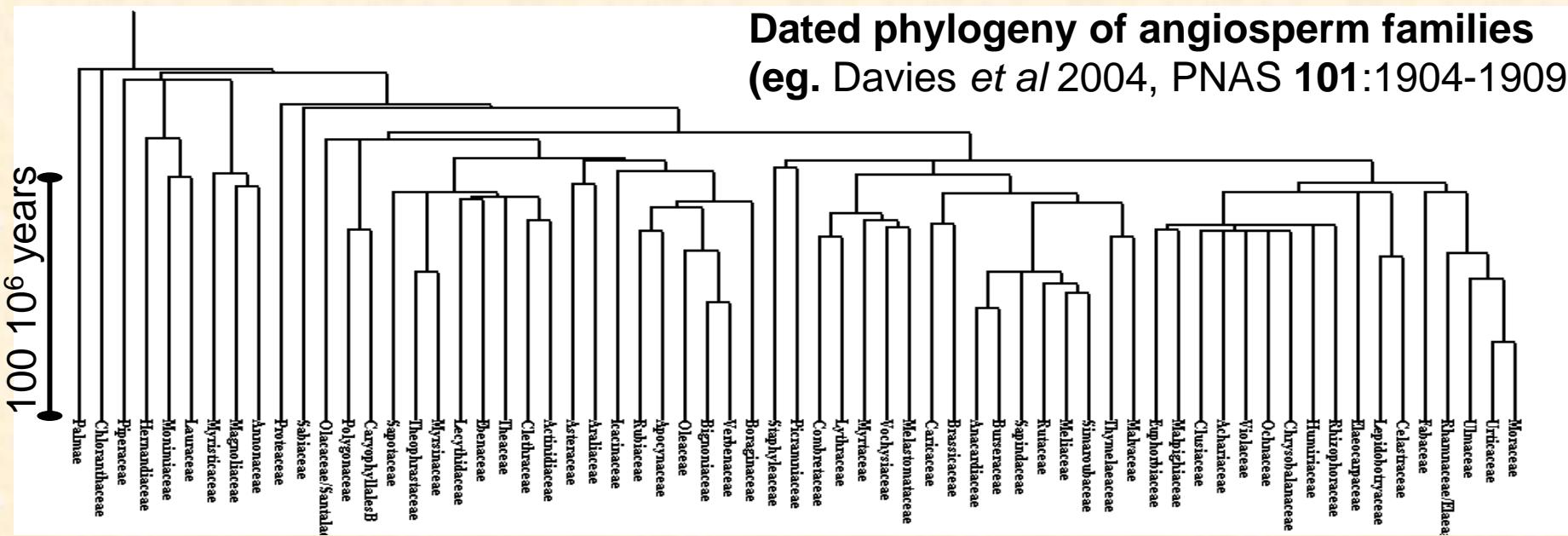
Ecotone : soil depth, luminosity, humidity...



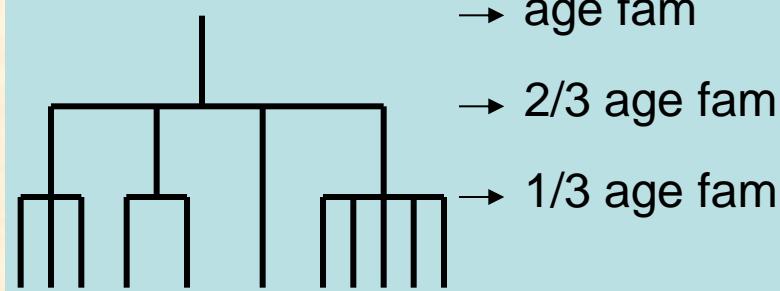
Ecological distance =
number of steps along the ecotone

Species phylogeny

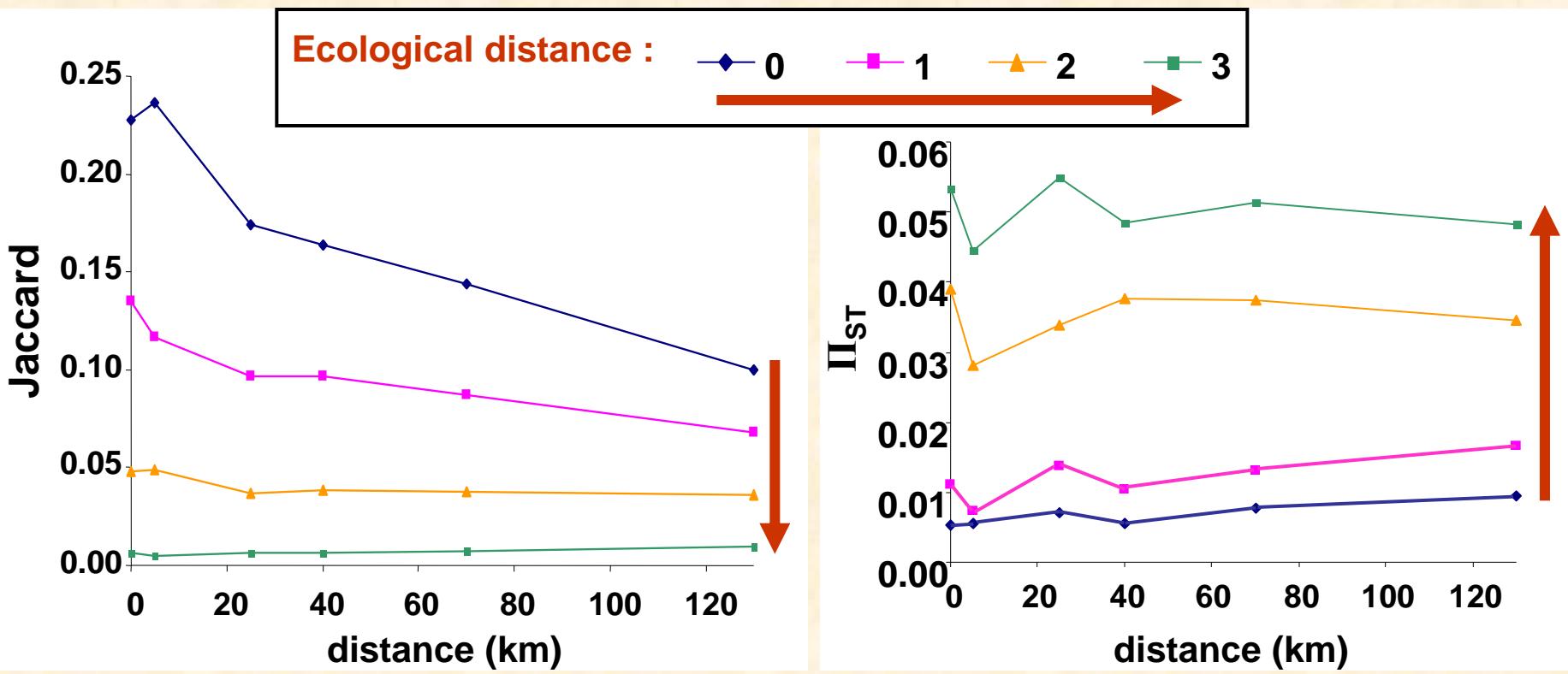
-> software Phylomatics (<http://www.phylodiversity.net/phylomatic/phylomatic.html>)
⇒ requires a list of species with their families (APG)



Within families, if no phylogeny available
→ polytomies of genera
→ polytomies of species
within genus



Species and phylogenetic turnover among inselbergs



Species turnover (J) depends on spatial distance + ecological gradient

Phylogenetic turnover (Π_{ST}) depends on ecological gradient only

Species turnover among inselbergs is probably due to dispersal limitation

Conclusion 1

At deep phylogenetic level (e.g. within angiosperms)

- phylogenetic clustering \Leftrightarrow ecological differentiation
 ~~$\not\Rightarrow$~~ dispersal limitation at regional scale
- ⇒ disentangle dispersal limitation *versus* ecological filtering

Parmentier & Hardy 2009 (Ecography 32: 613-622)

2. Impact of ecological gradients versus biogeography

Patterns of tree species phylogenetic turnover for 1ha (semi-)evergreen forest plots (DBH>10cm) on different continents

O Hardy, P Couteron, F Munoz, BR Ramesh, R Pélissier



data CTFS

Panama canal watershed

50 plots of 1ha
Area 50km x 60km

Rainfall:

1500-3300mm

Elevation :

50-400m

data French Institute
Pondicherry

Western Ghats

50 plots of 1ha
Area 50km x 200km

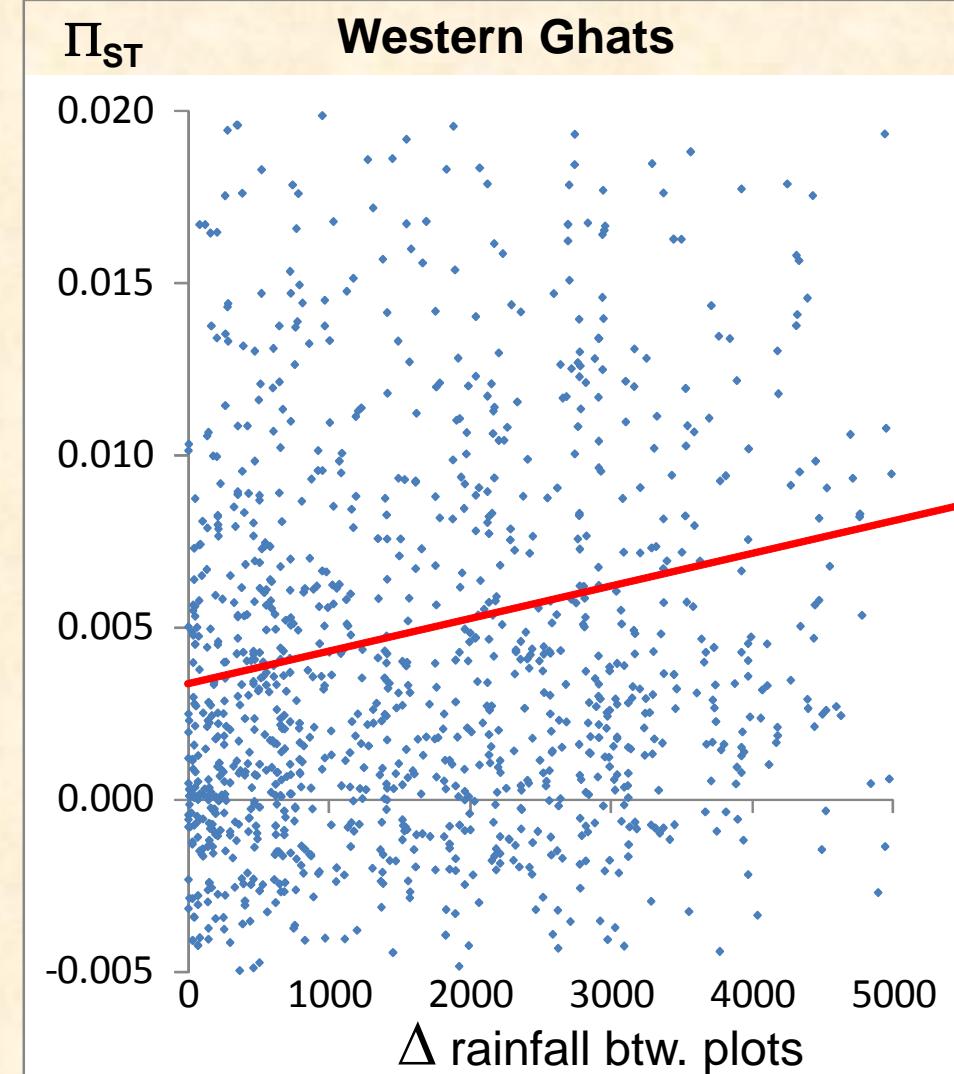
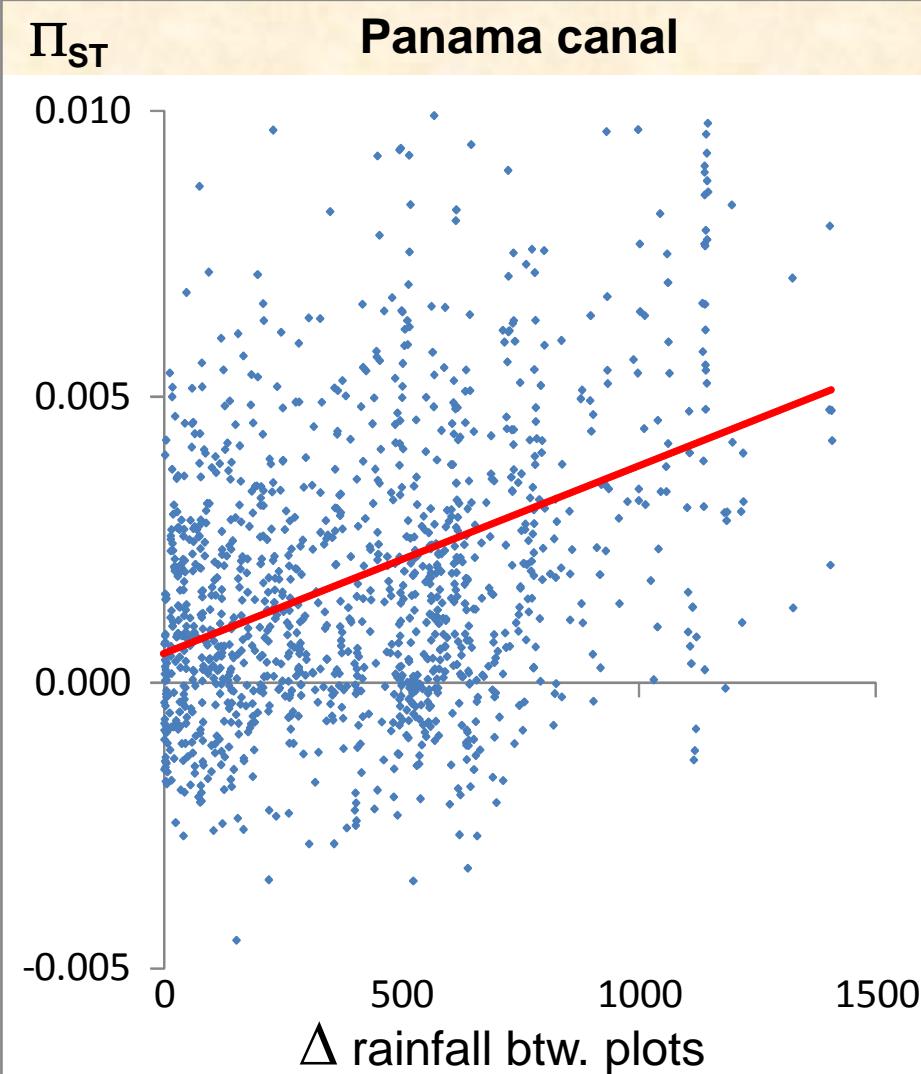
Rainfall:

1400-6000mm

Elevation:

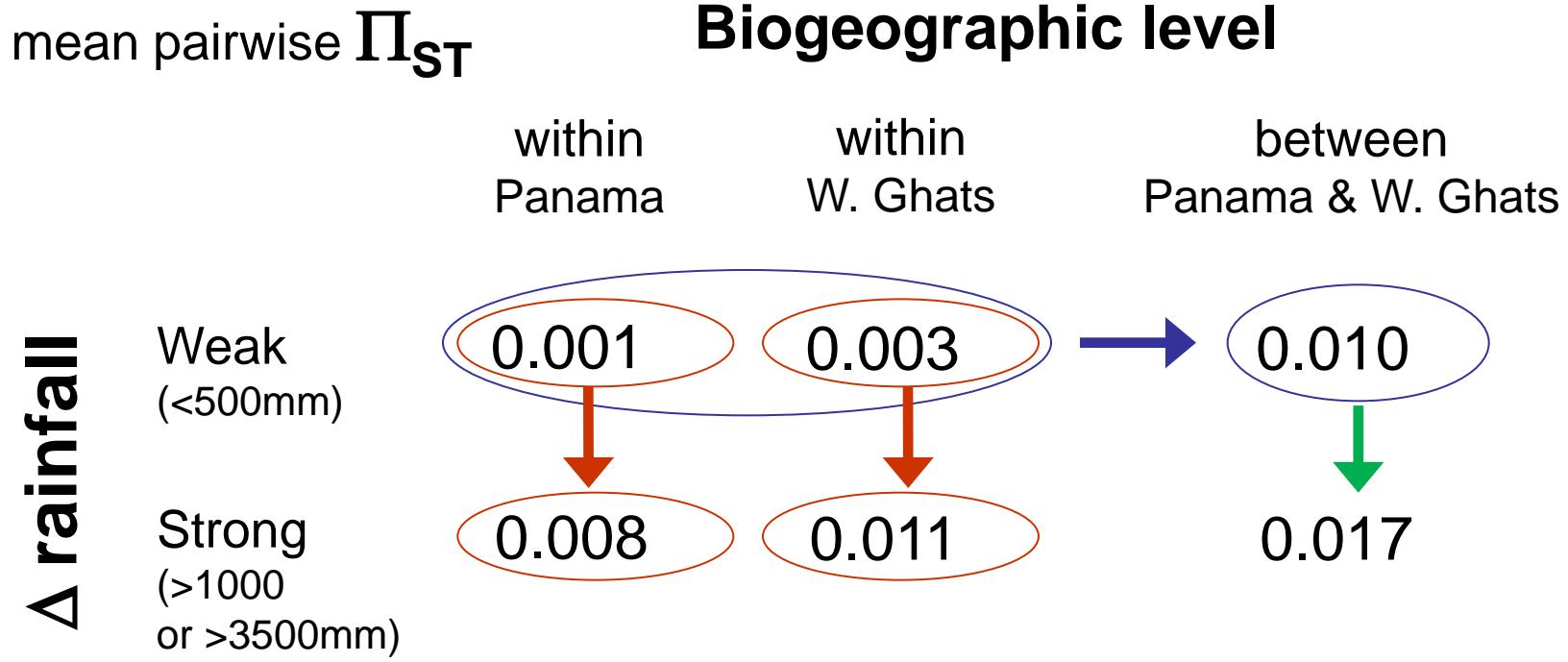
50-1050m

Phylogenetic turnover at regional scale



👉 Phylogenetic clustering within plots mostly explained by rainfall differences

Phylogenetic turnover impact of ecological gradients vs biogeography

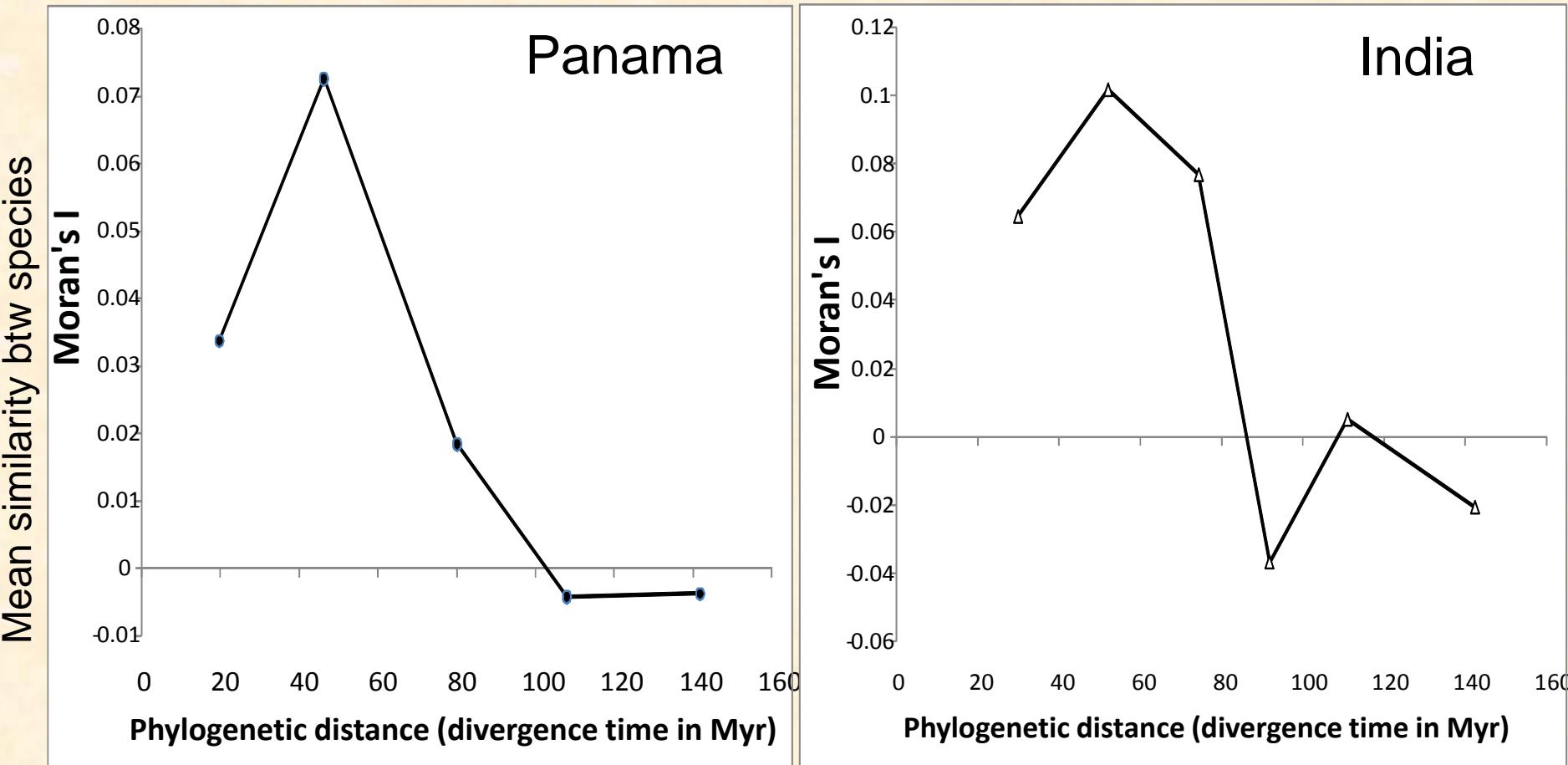


For forest tree communities, **at the scale of angiosperms**:

Impacts on Π_{ST} of { biogeography (inter-continental)
meso-climatic gradient } are comparable
+ seem cumulative

Is there a phylogenetic signal in species ‘adaptations’ towards the rainfall gradient?

Adaptation quantified by correlation between species abundance and rainfall/yr



Species having diverged <80 Myrs ago tend to share similar ‘adaptations’

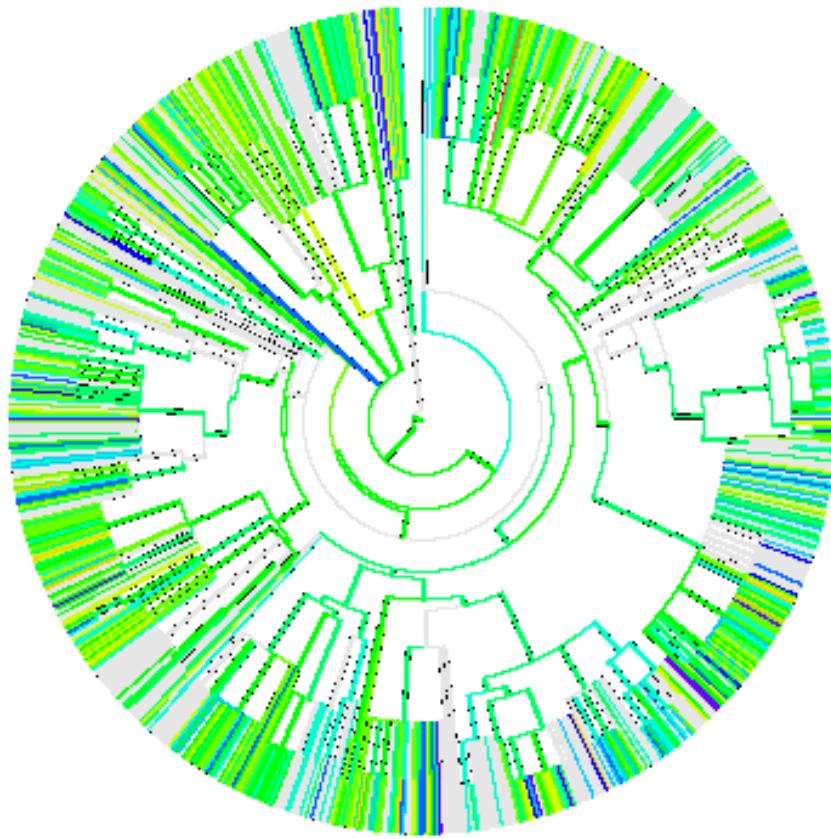
Reconstruction of ancestral niche along the phylogenies

↳ Mesquite software

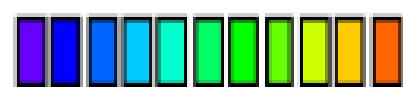
Panama

India

No shared species but
many shared clades



Moister



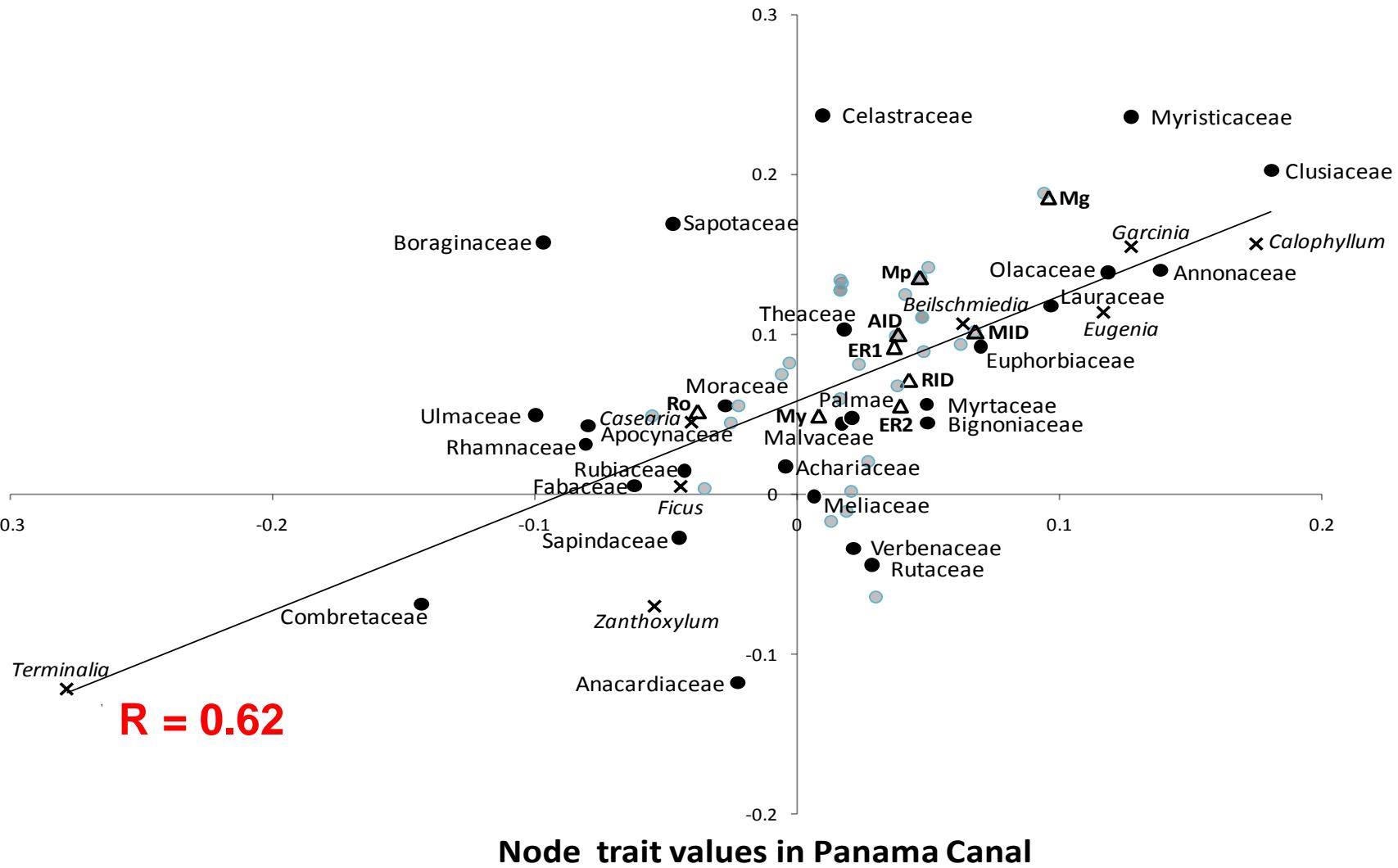
Drier

Drier

Drier

Inter-continental correlation of clade ‘adaptations’ with respect to rainfall gradient

Node trait values in Western Ghats



-> (pre)adaptations can persist over long evolutionary time (ca. 100 Myrs)

Conclusion 2

In rainforest tree communities, at deep phylogenetic level (e.g. within angiosperms):

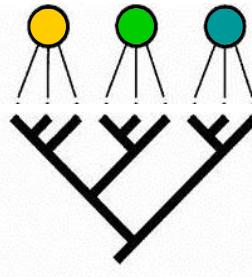
- relative impacts of biogeography and meso-climatic gradients on phylogenetic clustering are comparable (on Π_{ST} metric)
- deep (worldwide) phylogenetic conservatism of climate adaptations

Hardy et al. 2012 (Glob. Ecol. Biogeogr. 21:1007–1016)

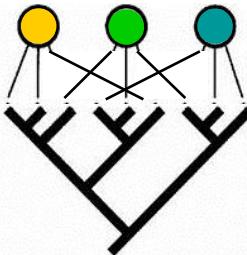
3. Integrating phylogeny and traits

C x P

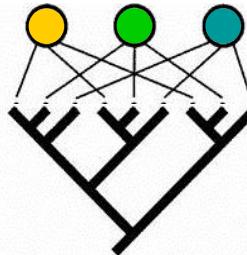
Phylogenetic clustering



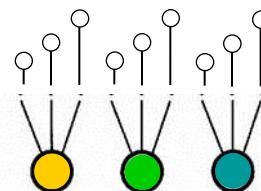
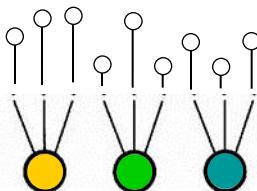
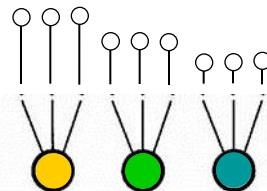
Random phylogenetic structure



Phylogenetic overdispersion



T x P



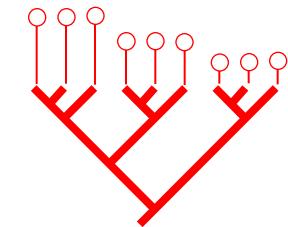
T x C

Environmental filtering

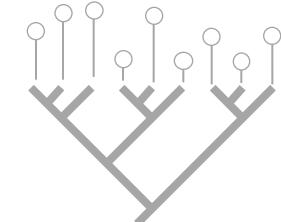
Random assembly (neutrality)

Competitive exclusion

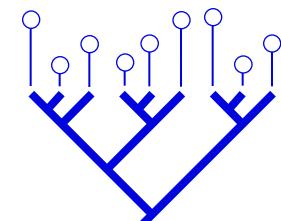
T x P



Trait conservatism



No phylogenetic signal



Trait convergence

Ideally, all levels should be taken into account

3. The BRIDGE project : phylogenetic and functional structures of rainforests from French Guiana

Jérôme Chave, Chris Baraloto, Kyle Dexter, Timothy Paine, Olivier Hardy, Vincent Savolainen



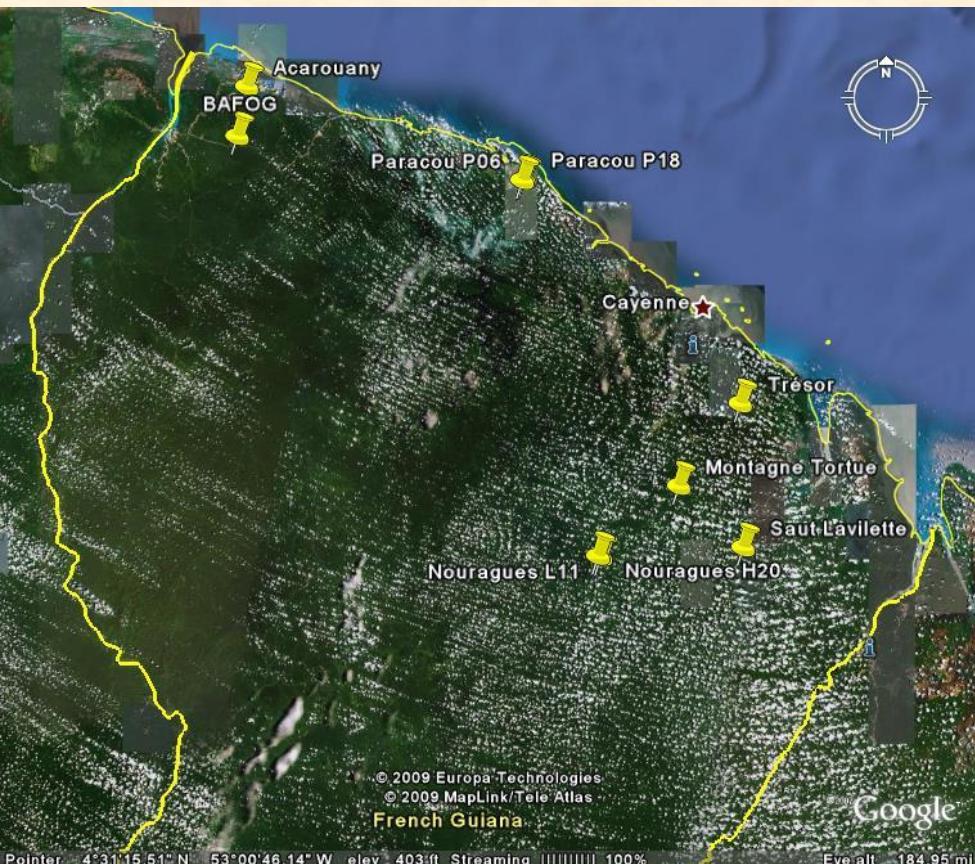
Guyane field teams (Jean-Yves Serein, Jean-Yves Goret, Elodie Courtois, Pierre-Alain Blandinières, Pascal Pétronelli, Mailyn Gonzalez, Odile Poncy, Julien Engel...)

Toulouse lab team (Celine Vicedos, Mailyn Gonzalez, Julien Vieu...)

Funding:
ANR (programme Biodiversité)
CNRS (programme AMAZONIE)
Génoscope
Région Guyane, DIREN, DRRT

BRIDGE Project: 9 one-ha plots sampled

Trees DBH > 10cm



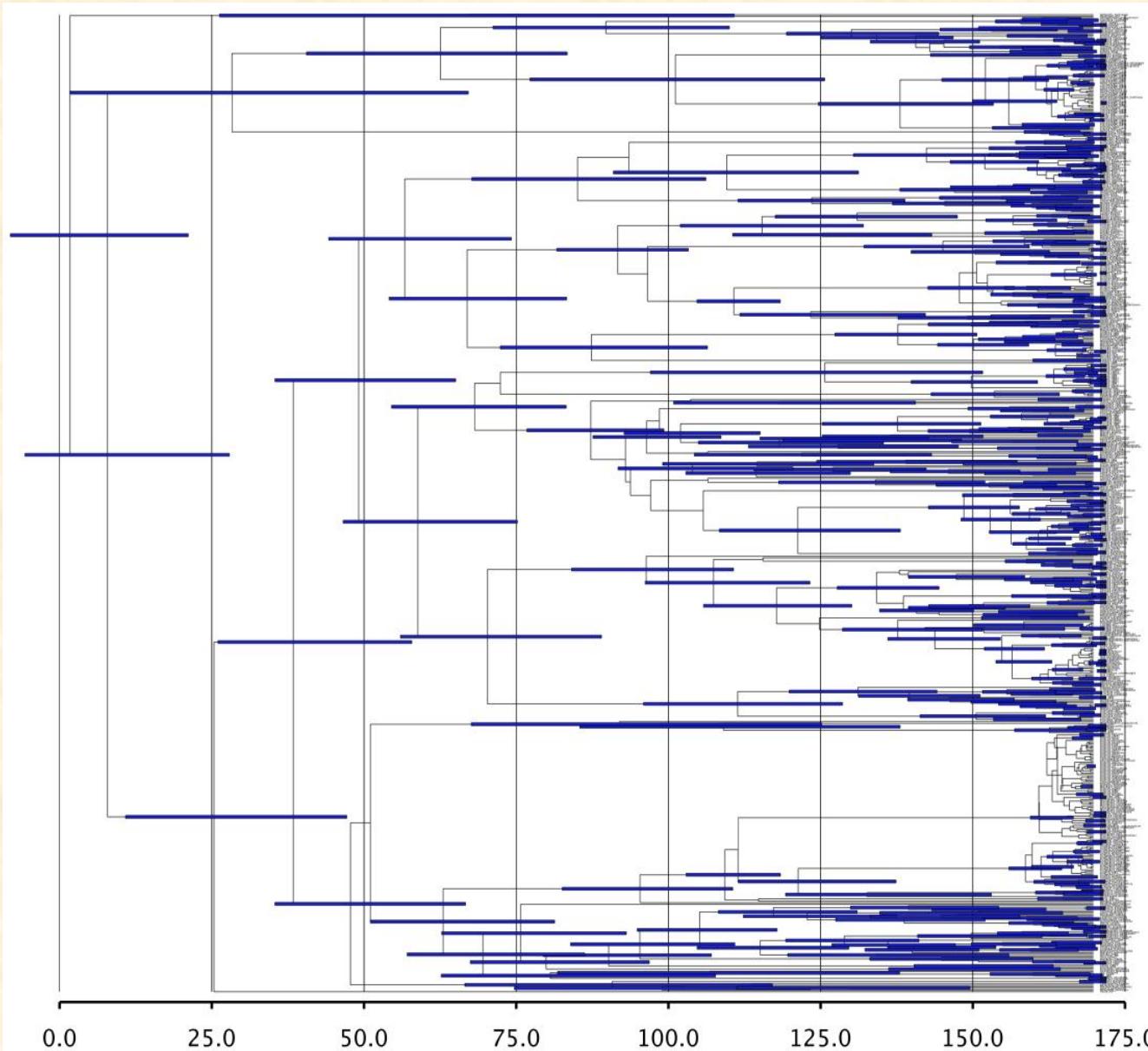
| Plot | # ind | # sp | Fisher's α |
|-----------------|-------------|------------|-------------------|
| Bafog | 576 | 153 | 67.4 |
| Acarouany | 442 | 147 | 75.3 |
| Paracou 1 | 632 | 149 | 60.8 |
| Paracou 2 | 478 | 151 | 74.6 |
| Nouragues 1 | 544 | 196 | 108.2 |
| Nouragues 2 | 507 | 177 | 95.6 |
| Trésor-Kaw | 418 | 156 | 89.2 |
| Montagne Tortue | 501 | 201 | 122.3 |
| Saut Lavilette | 582 | 208 | 114.8 |
| TOTAL | 4680 | 688 | |



Traits Measured

- Leaf toughness
 - Leaf tissue density
 - Leaf area
 - Specific leaf area
 - N content
 - C:N ratio
 - NH_4 proportion
 - P content
 - K content
 - leaf ^{13}C
 - Chlorophyll content
- Sapwood density
- Trunk moisture
- Bark thickness
- Twig wood density
- Twig bark thickness
- Leaf structure traits**
- Leaf chemistry traits**
- Trunk and wood traits**
- Twig traits**

**Phylogeny generated for >600 species (*rbcL* and *matK*)
+ temporal calibration using relaxed molecular clock**



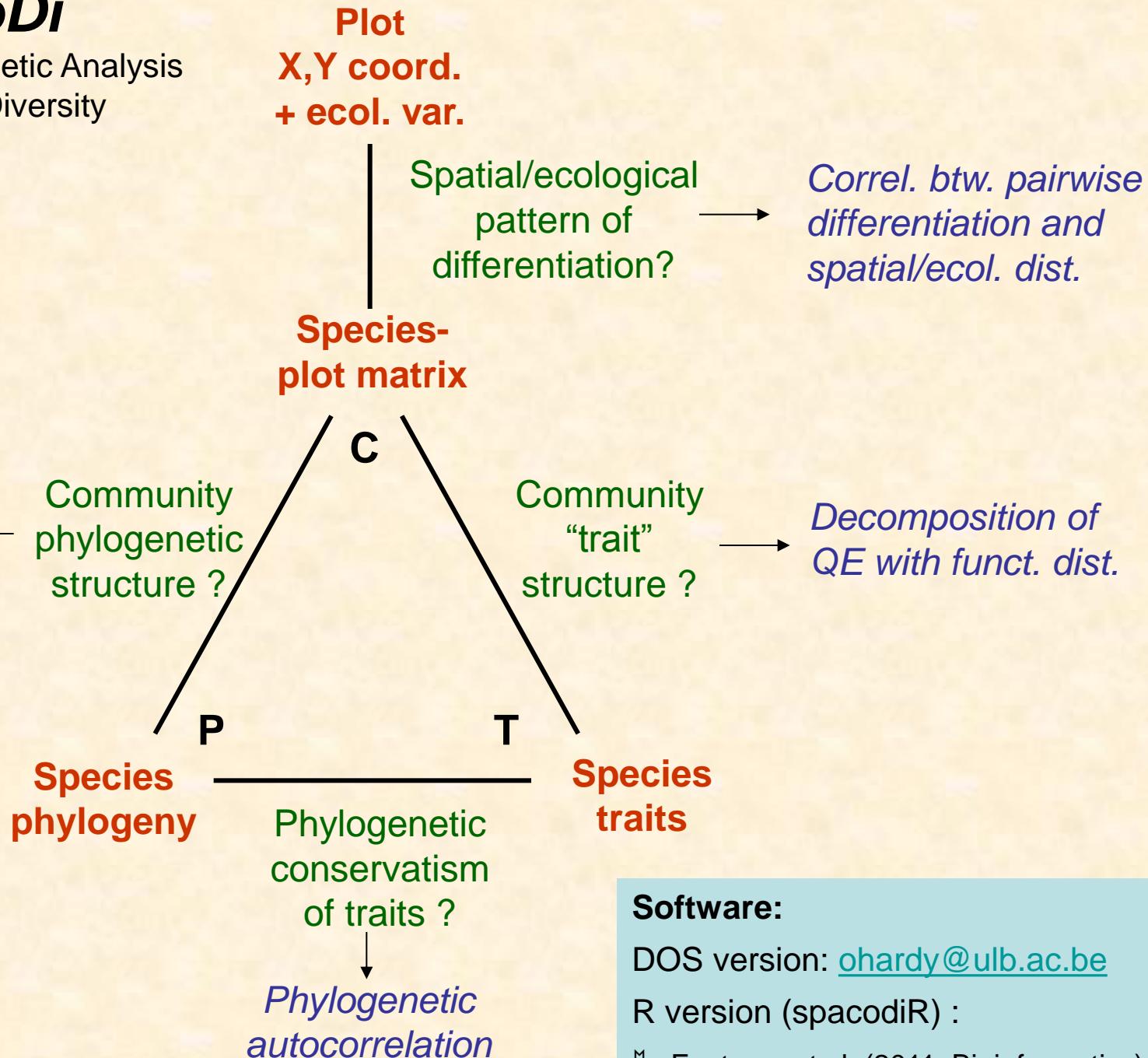
BEAST v1.53
Drummond and
Rambaut 2009

SPACoDi

Spatial and Phylogenetic Analysis
of Community Diversity

{
Data
Questions
Methods

Decomposition
of QE with phylo
distance
Whole or partial
phylogeny
random. tests



Comparing phylogenetic and functional community structures

From a phylogeny

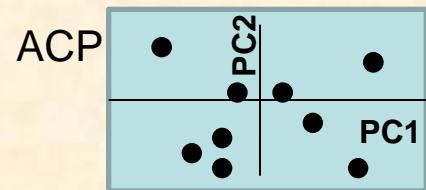


phylogenetic
or
functional

δ_{kl}

| | 0 | 1 | 2 | 3 |
|---|---|---|---|---|
| 0 | 0 | 1 | 2 | 3 |
| 1 | 1 | 0 | 2 | 2 |
| 2 | 2 | 2 | 0 | 1 |
| 3 | 3 | 2 | 1 | 0 |

From species traits



distances between species

From community inventories:

⇒ partition of QE = mean δ_{kl} between individuals

sampled

within sites

$$QE_w = \bar{\delta}_{(2_individuals_from_same_site)}$$

among sites

$$QE_a = \bar{\delta}_{(2_individuals_from_different_sites)}$$

$$\rightarrow P_{ST} = 1 - QE_w / QE_a$$

species + phylogenetic turnover

$$\rightarrow T_{ST} = 1 - QE_w / QE_a$$

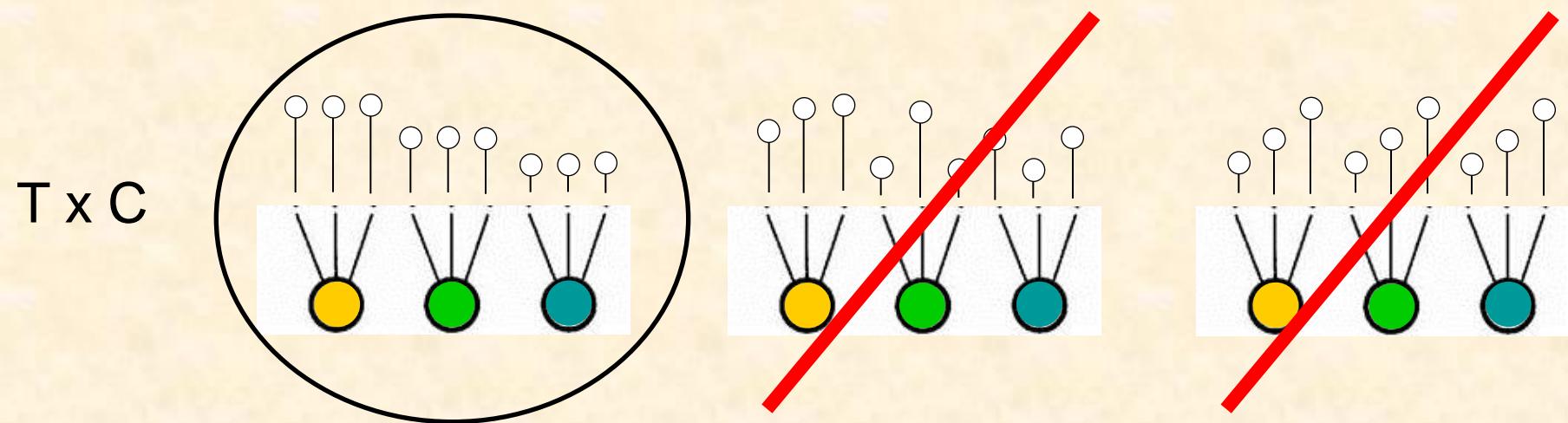
species + functional turnover

or

CxT : Functional structure of communities

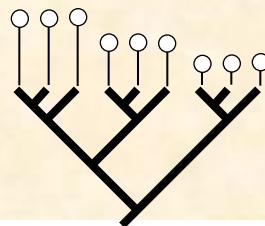
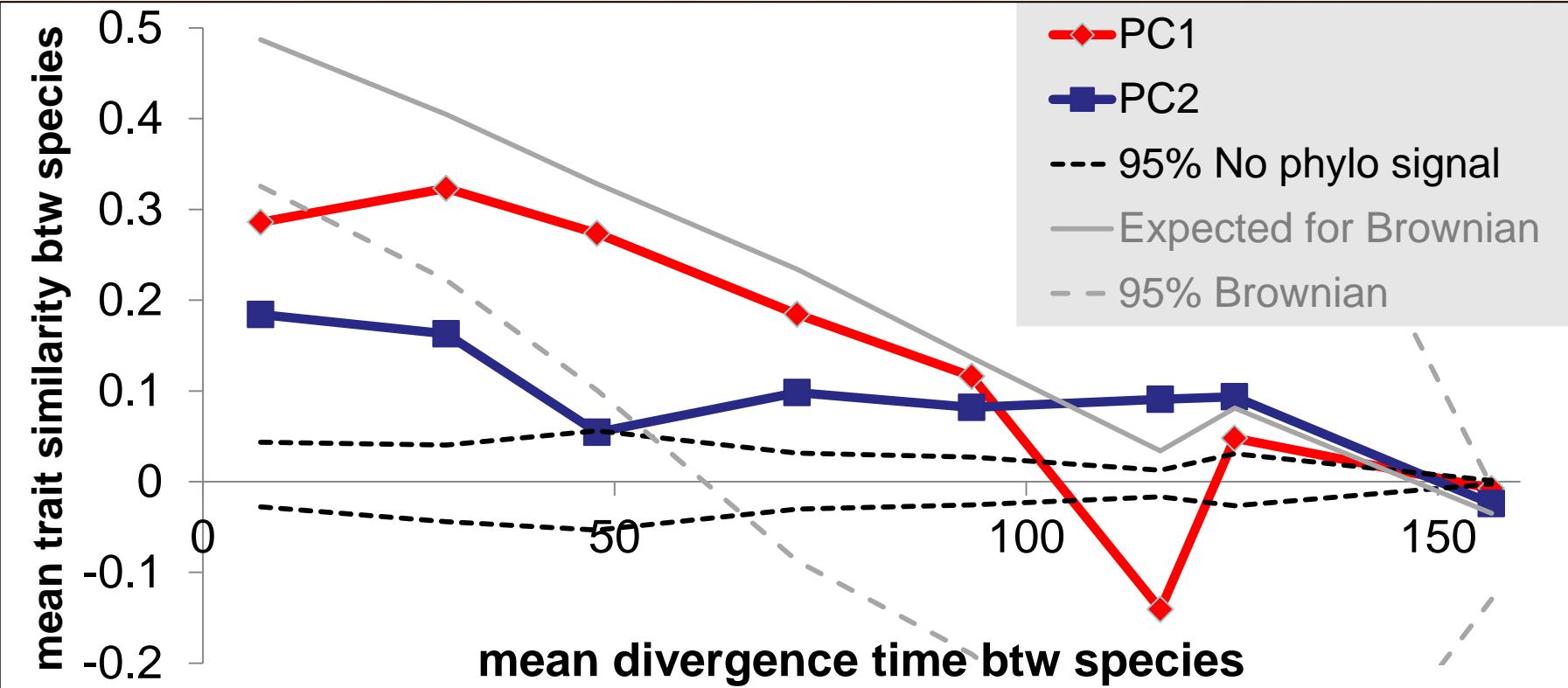
- $T_{ST} = \mathbf{0.0313} > [0.0076, 0.0172]$ under H_0
- $\tau_{ST} = \mathbf{0.0119} > [-0.0010, 0.0014]$ under H_0

→ **Functional clustering**

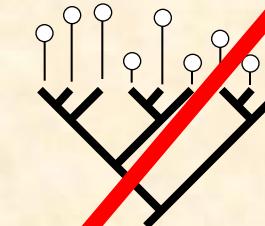


↳ interpretation: habitat filtering dominates

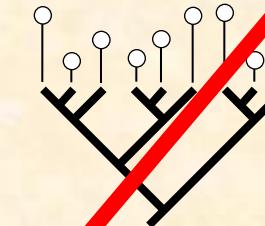
TxP : Phylogenetic structure of traits



T x P
Trait conservatism



No phylogenetic
signal



Trait convergence

PC1 conforms to Brownian motion model

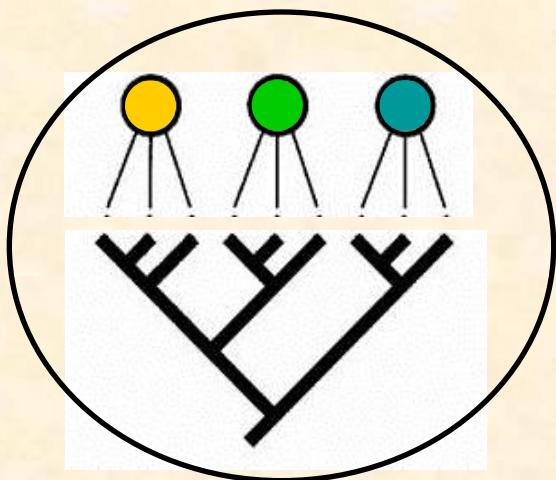
PC2 less conserved than Brownian (conform to Ornstein-Uhlenbeck model)

PxC : Phylogenetic structure of communities

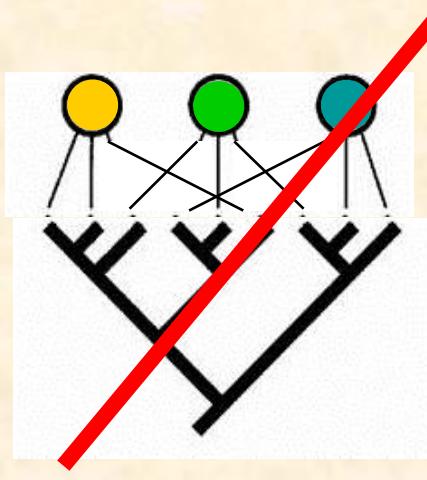
- $P_{ST} = \mathbf{0.0246} > [0.0098, 0.0161]$ under Ho
- $\Pi_{ST} = \mathbf{0.0033} > [-0.0006, 0.0008]$ under Ho

→ Phylogenetic clustering

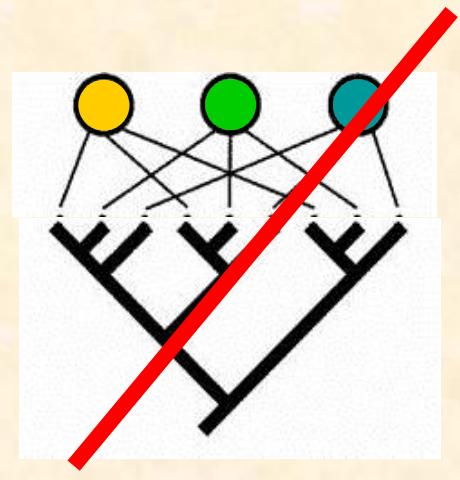
PxC



Phylogenetic
clustering



Random
phylogenetic
structure

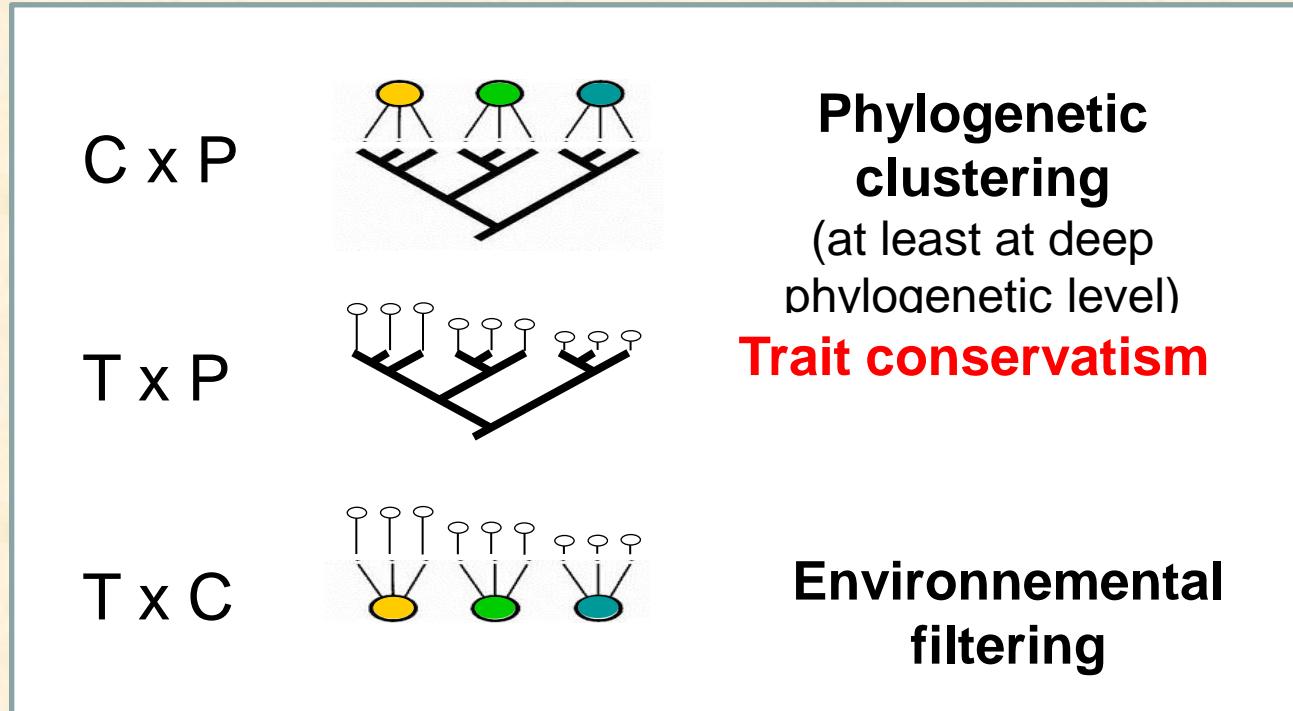


Phylogenetic
overdispersion

Conclusion 3

In continental tropical rainforests

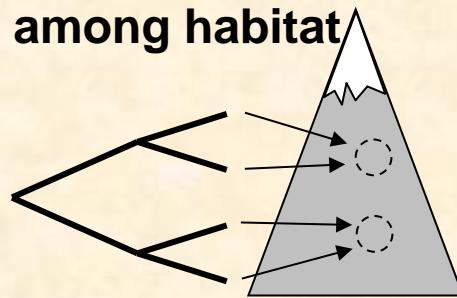
- trait conservatism seems to hold in general
- phylogenetic clustering reflects functional clustering due to environmental filtering



Does phylogenetic overdispersion sometimes occur ?

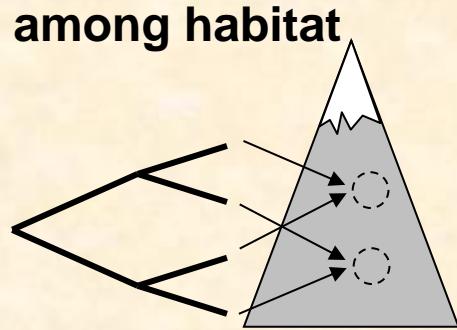
Phylogenetic clustering

↳ case studies 1, 2, 3



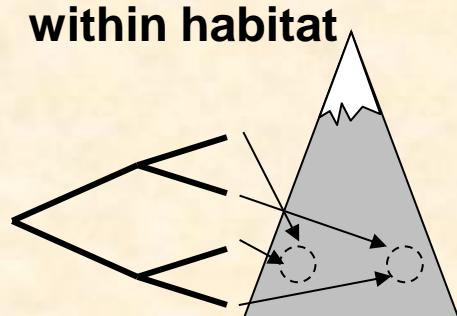
Phylogenetic overdispersion

↳ case study 4 ?



Phylogenetic overdispersion

↳ case study 5 ?



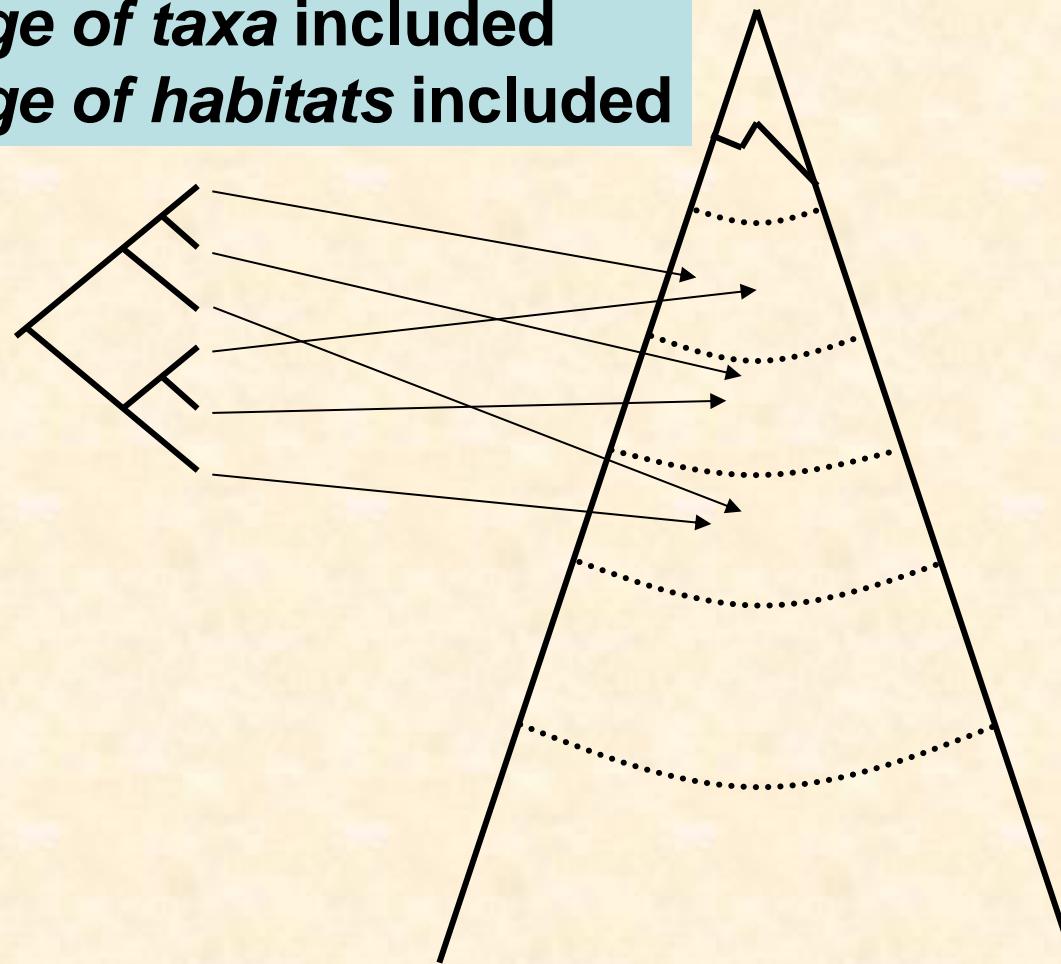
environmental filtering
adaptation to ≠ elevation
+
habitat conservatism

environmental filtering
adaptation to ≠ elevation
+
habitat convergence
e.g. *radiation of several clades in a set of new habitats*

competitive exclusion
+
niche conservatism
→ sister species cannot coexist because of niche overlap

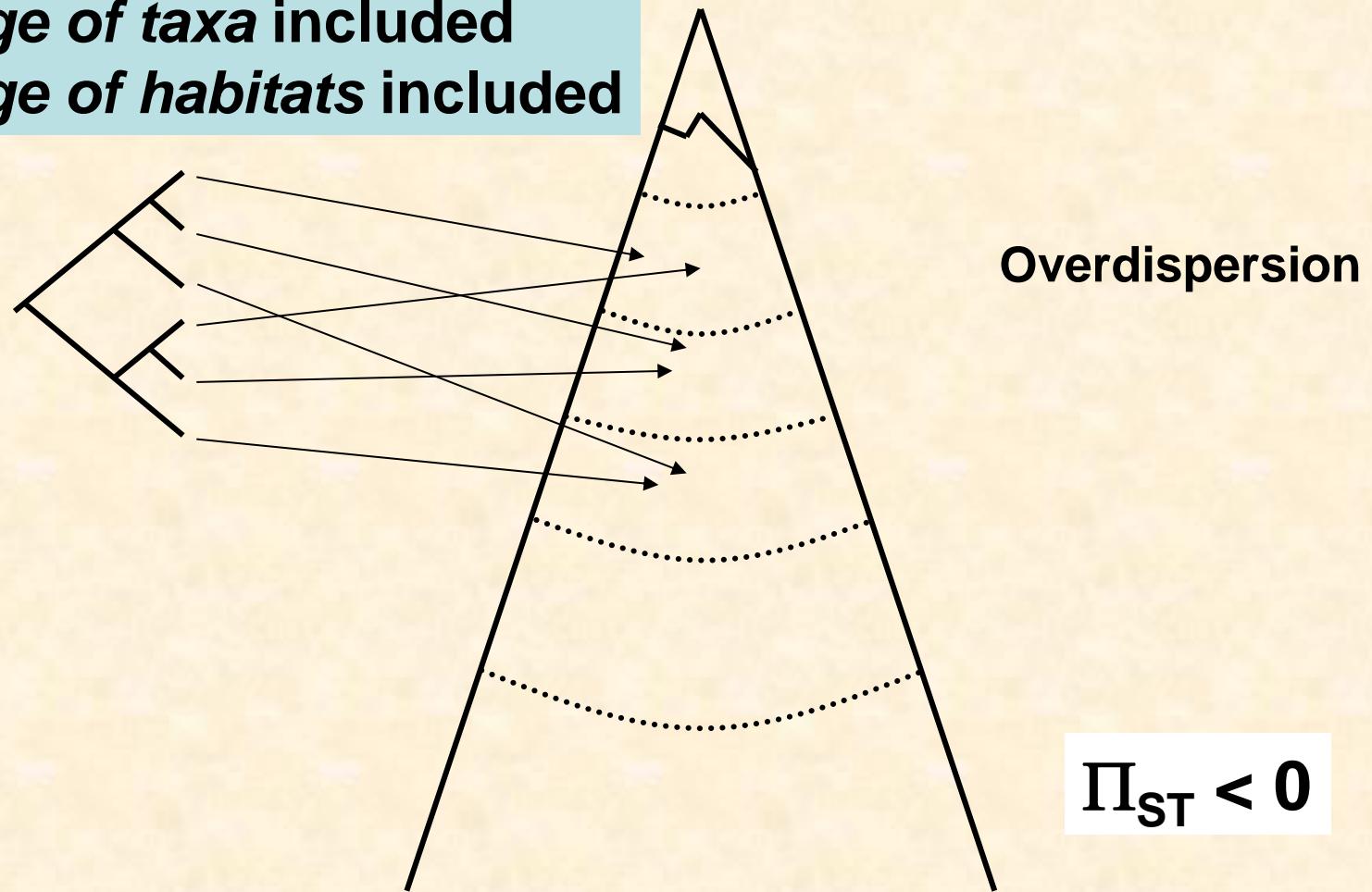
Phylogenetic clustering / overdispersion is relative to

{ the *range of taxa included*
the *range of habitats included*



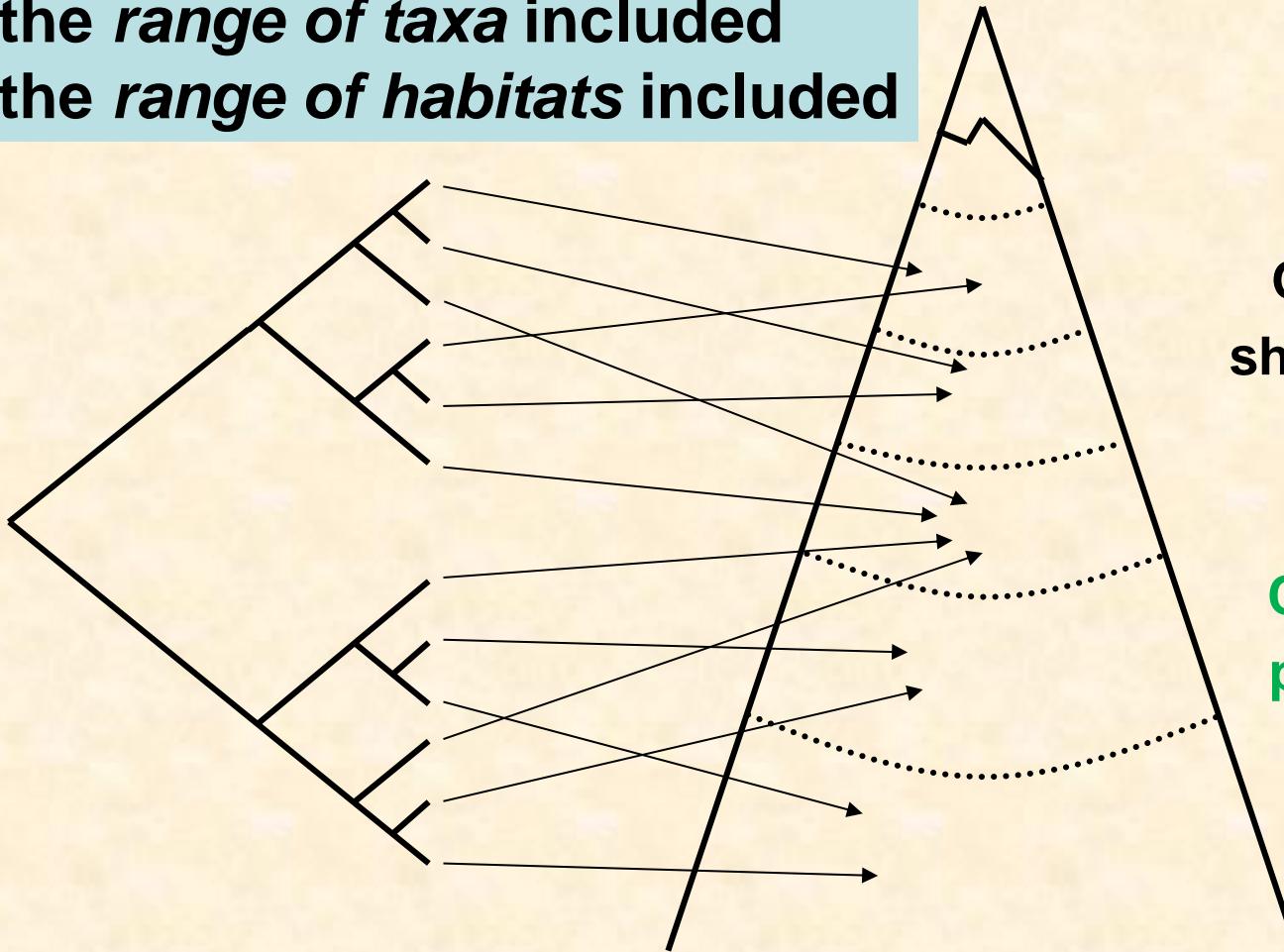
Phylogenetic clustering / overdispersion is relative to

{ the *range of taxa included*
the *range of habitats included*



Phylogenetic clustering / overdispersion is relative to

{ the *range of taxa included*
the *range of habitats included*



Overdispersion at
shallow phylogenetic
level

+

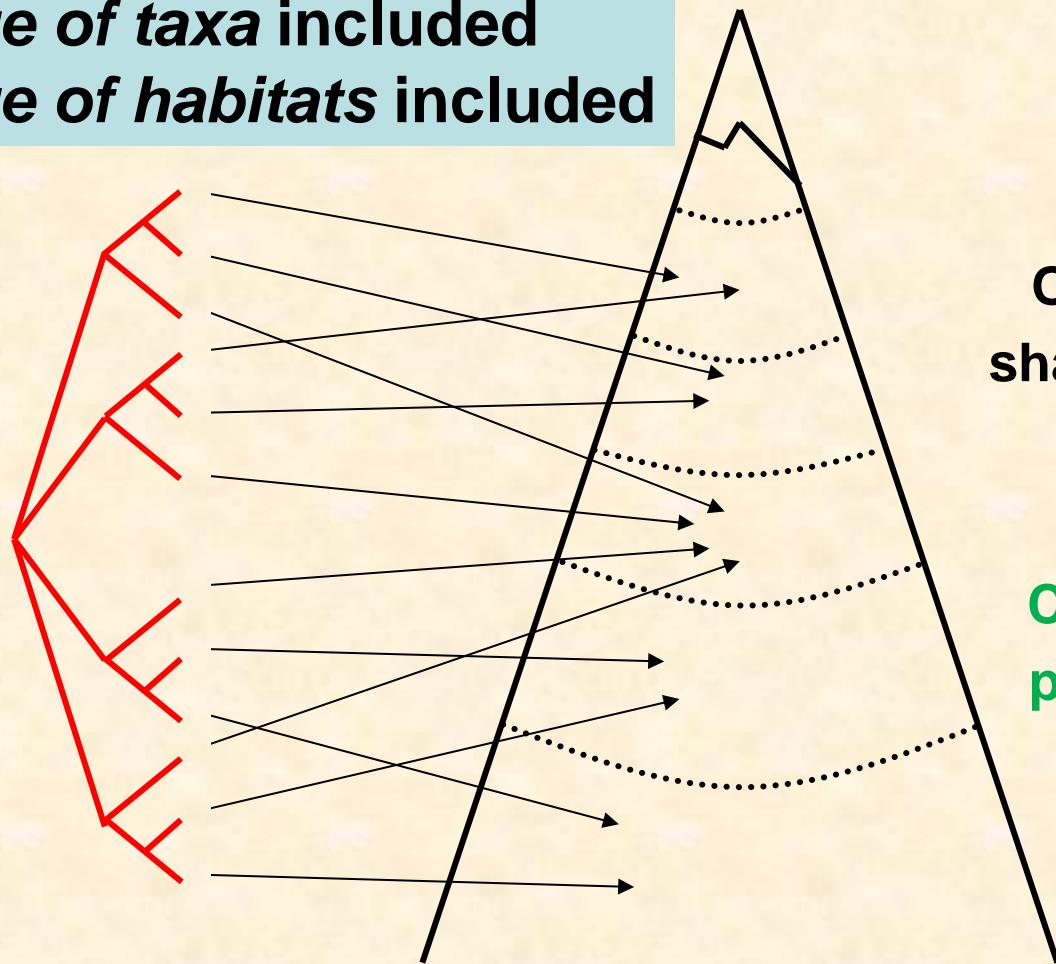
Clustering at deep
phylogenetic level

$$\Pi_{ST} > 0$$

How to test this pattern ?

Phylogenetic clustering / overdispersion is relative to

{ the *range of taxa included*
the *range of habitats included*



Overdispersion at
shallow phylogenetic
level

+

Clustering at deep
phylogenetic level

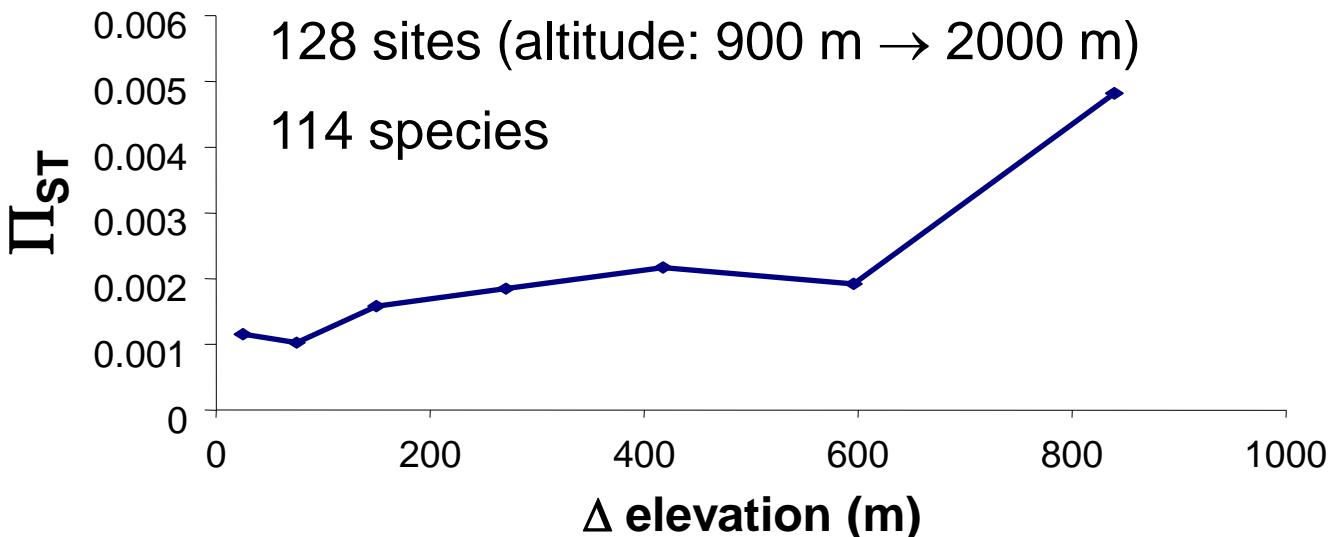
$$\Pi_{ST} < 0$$

How to test this pattern ?

👉 transform the phylogenetic tree to remove deep phylogenetic information

4. Phylogenetic structure under recent multiple radiations: Forest communities from Reunion island

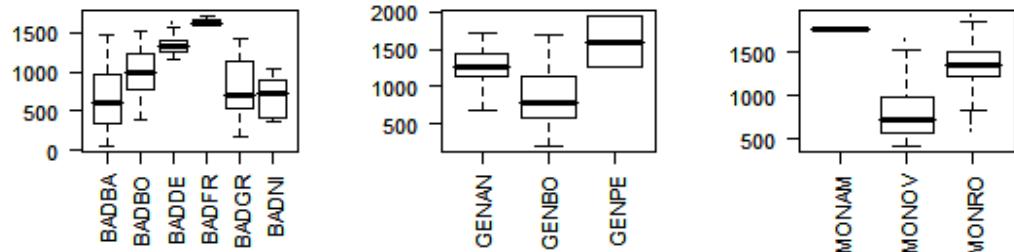
J. Eyraud, D. Stasberg, C. Thebaud, O. Hardy, J. Chave (unpublished)



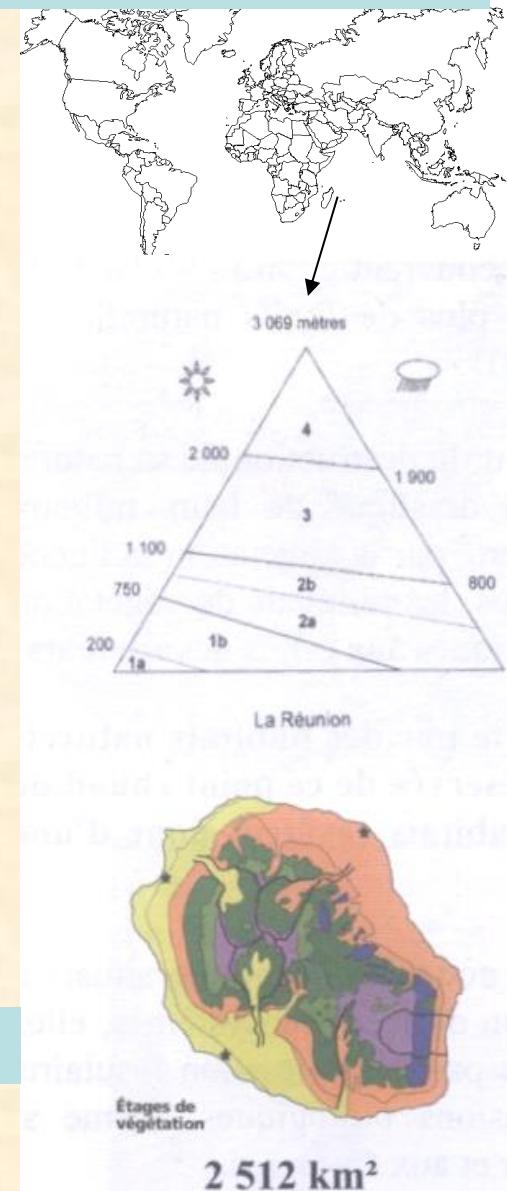
→ phylogenetic clustering ↑ with Δ elevation

However, 9 genera have diversified on the island
→ why no phylogenetic overdispersion ?

Elevation ranges of endemic species in 3 genera



PARADOX?



Phylogenetic turnover at different time depths

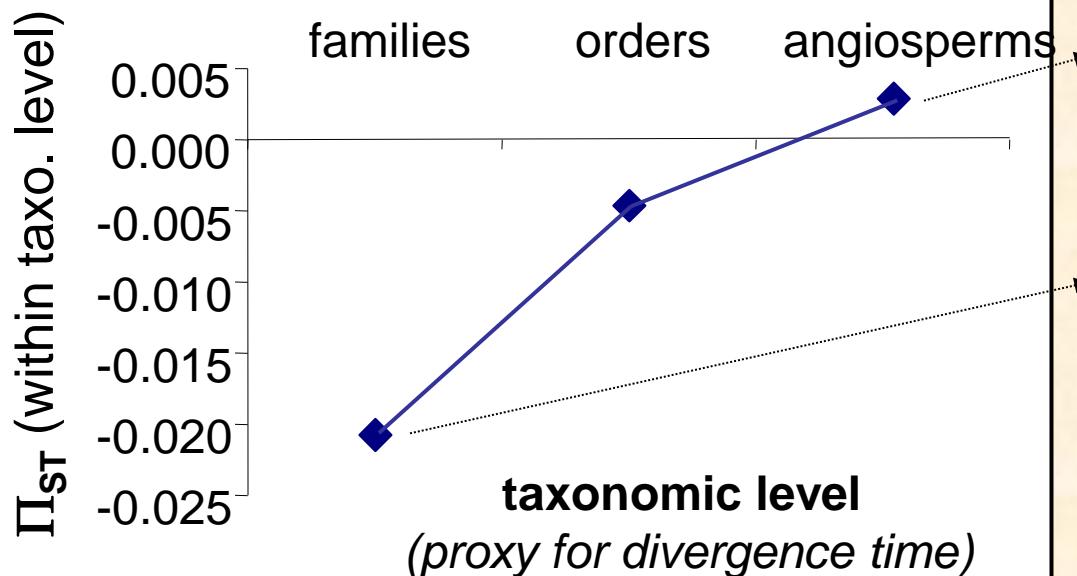
$$MPD^{\delta < T}$$
 mean divergence time (δ) between species for $\delta < T$

→ compare pairs of species sampled *within* versus *among* sites :

$$\Pi_{ST}^T = 1 - MPD_w^{\delta < T} / MPD_a^{\delta < T}$$

→ characterize phylogenetic turnover at different time depths

Example: Reunion island



for all angiosperms

→ phylogenetic clustering

↳ overall habitat conservatism

within families

→ phylogenetic overdispersion

↳ diversification of several genera
(congeneric species do not co-occur)

Conclusion 4

At shallow phylogenetic levels (e.g. within a genus or a family)

- phylogenetic overdispersion may occur due to recent local diversification (e.g. on island)

⇒ impact of time depth detectable using $\Pi_{ST}^{\delta < T}$

Emerging patterns in the comparative analysis of phylogenetic community structure

S. M. VAMOSI,* S. B. HEARD,† J. C. VAMOSI* and C. O. WEBB‡

Molecular Ecology (2009) 18, 572–592

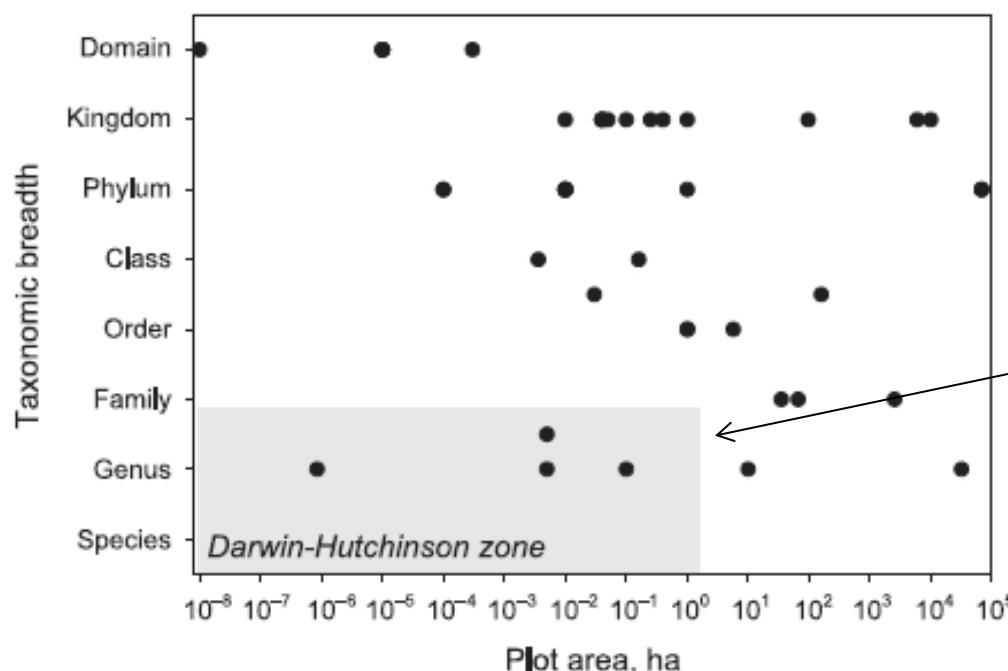


Fig. 4 Taxonomic and geographical scales of contemporary studies in phylogenetic community structure. The x-axis refers to the geographical extent of individual study plots, and the y-axis the highest taxonomic level considered in the study. The 'Darwin–Hutchinson zone' marks an approximate region of the plane for which we might be particularly interested in community phylogenetic structure: plots small enough for individuals to interact, and species closely enough related that competition is a plausible expectation (see Taxonomic and geographical scales for further discussion).

=> Phylogenetic clustering
most often observed, but
competitive exclusion may
be important in the
'Darwin–Hutchinson zone'



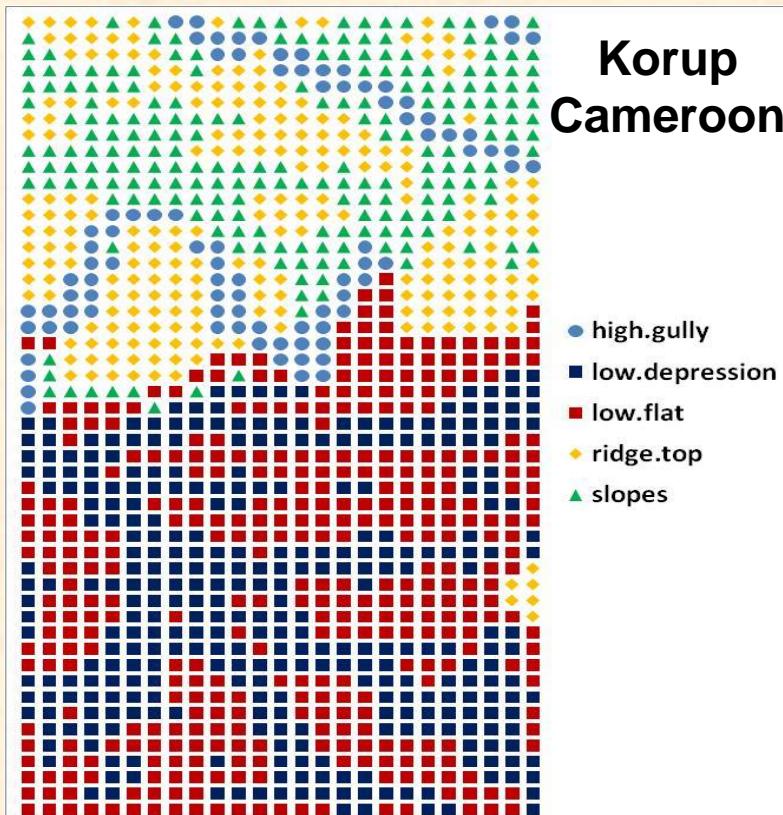
Investigate patterns at:
- scale of inter-individual
competition
- shallow phylogenetic depth

5. Phylogenetic structure under inter-individual competition : fine-scale spatial phylogenetic turnover in 50-ha plots

O Hardy, I Parmentier, M Réjou-Méchain, J Duminil, M Kuzmina, DW Thomas, D Kenfack, GB Chuyong

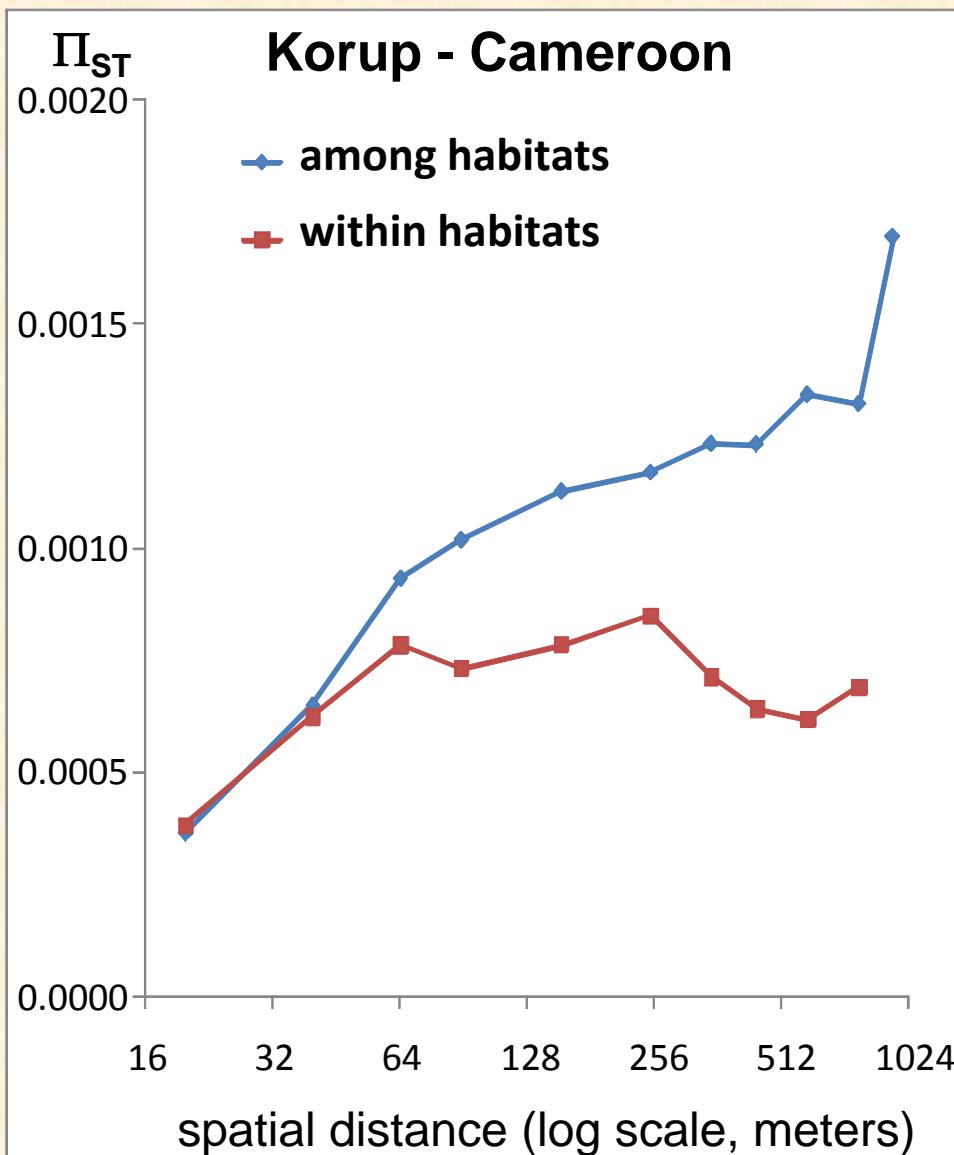
500m x 1000m evergreen forest plot (330 000 stems DBH>1cm inventoried)

↳ subdivided into 20m x 20m subplots categorized into **5 habitats**



**Molecular phylogenetic trees
based on *rbcL* + *matK***
(for 272 species represented by >50
individuals and making 97% of all
stems of 494 species)

Phylogenetic turnover: impact of spatial distance vs topographic habitat types



Habitat categories explain part but not all the phylogenetic turnover

Phylogenetic clustering ($\Pi_{ST}>0$) occurs at all scales (no overdispersion)

Phylogenetic turnover at shallow phylogenetic depth: Does phylogenetic overdispersion occur?

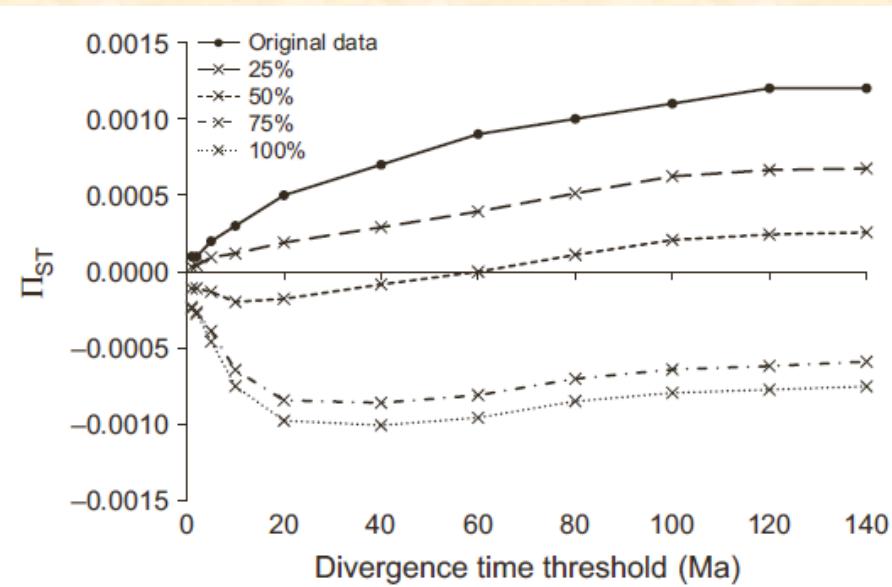
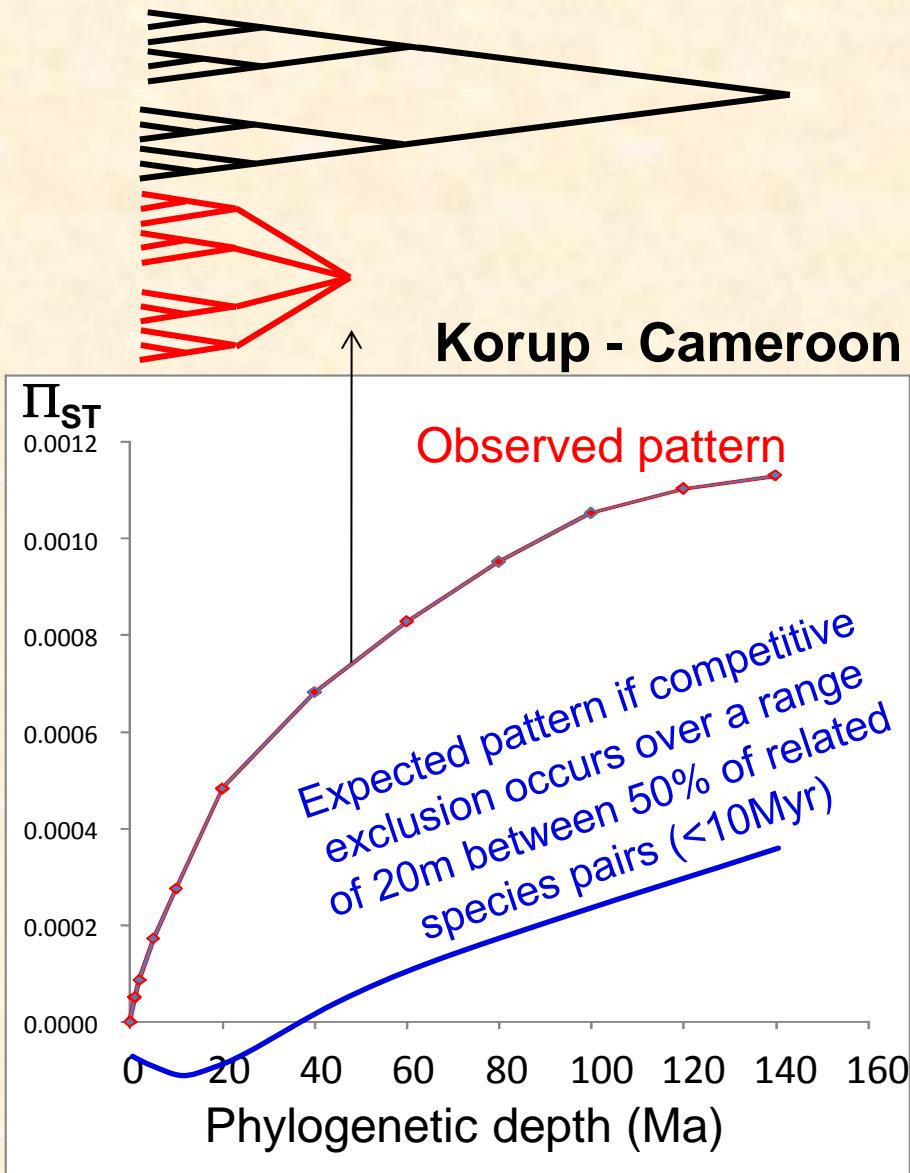


Fig. 4. Overall signal of phylogenetic turnover among quadrats (mean Π_{ST} , solid line) depending on phylogenetic depth (in Ma). To eliminate signal occurring above a given phylogenetic depth, the mean Π_{ST} is computed using truncated phylogenetic trees (see Fig. 1) for a range of divergence time thresholds. Simulated data help compare the observed pattern (i.e. original data) with a situation where competitive exclusion would have occurred between closely related species over a distance range of 20 meters (dashed lines) for 25% to 100% of the species pairs having diverged <10 Ma ago.

👉 no evidence of phylogenetic overdispersion

Does phylogenetic overdispersion occur at smaller scales?

↳ individual-based analyses

Computing a ‘co-occurrence excess’ of hetero-specific pairs

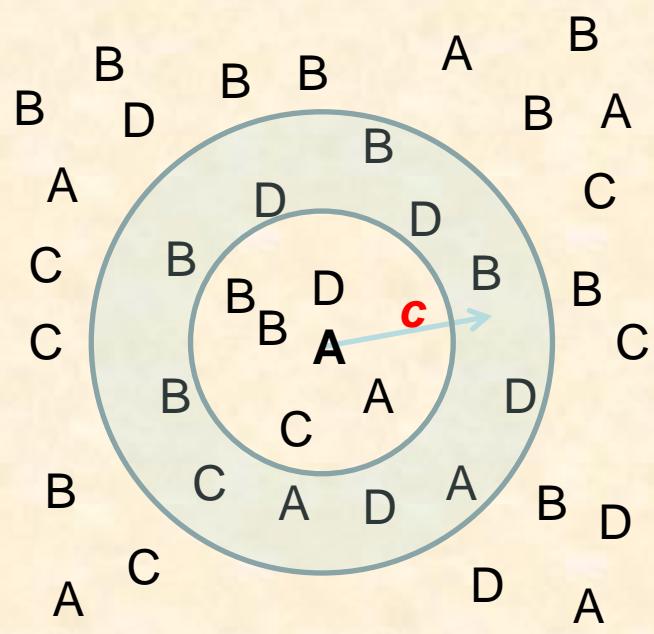
$\varepsilon_{A,i} = 1$ if ind i belongs to species A, otherwise 0

$\omega_{i,j,c} = 1$ if distance i-j belongs to interval c, otherwise 0

$$f_{A,B,c} = \frac{\sum_{i,j} \varepsilon_{A,i} \varepsilon_{B,j} \omega_{i,j,c}}{\sum_{i,j} \varepsilon_{A,i} \omega_{i,j,c} (1 - \varepsilon_{A,j})} - \frac{\sum_i \varepsilon_{B,i}}{\sum_i (1 - \varepsilon_{A,i})}$$

frequency of B at
distance c from A
(excluding A)

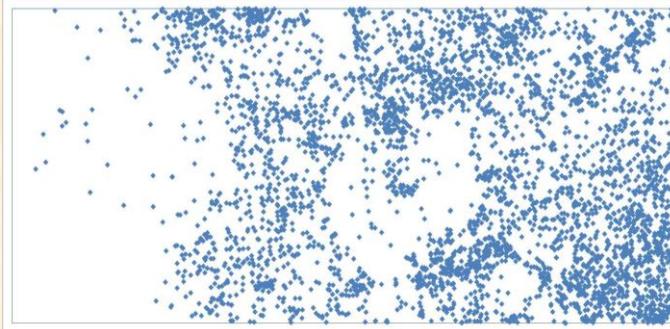
frequency of B in the
plot (excluding A)



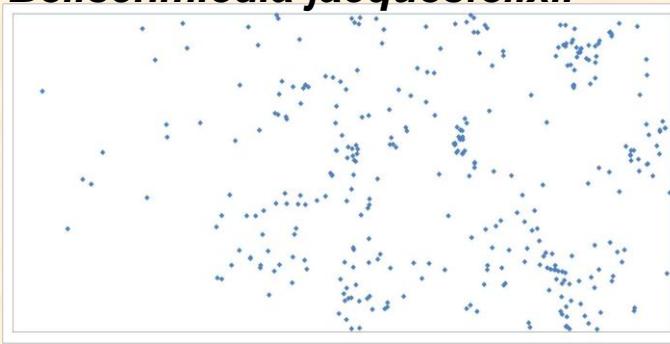
Does phylogenetic overdispersion occur at smaller scales?

↳ individual-based analyses

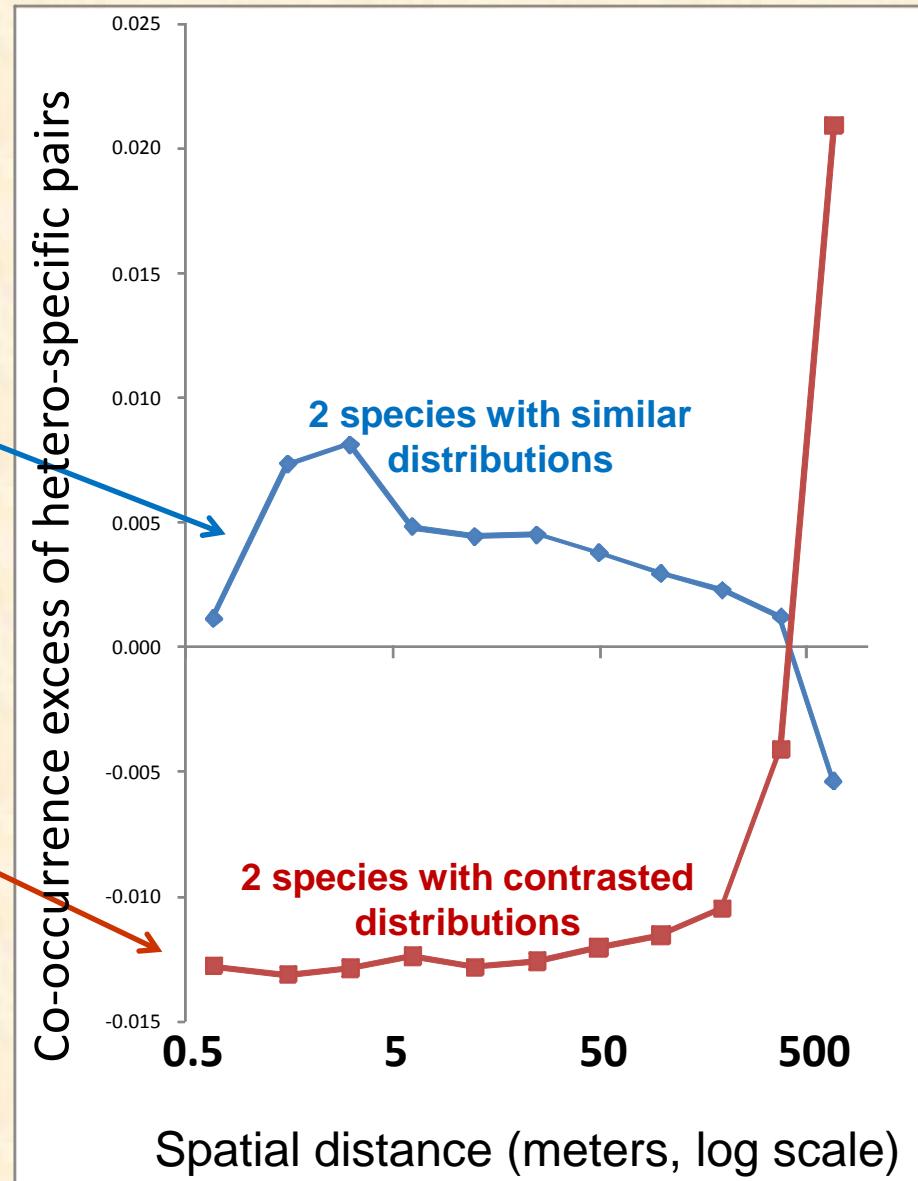
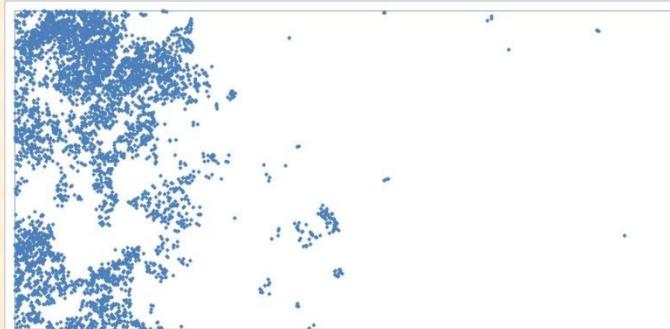
Cola cauliflora



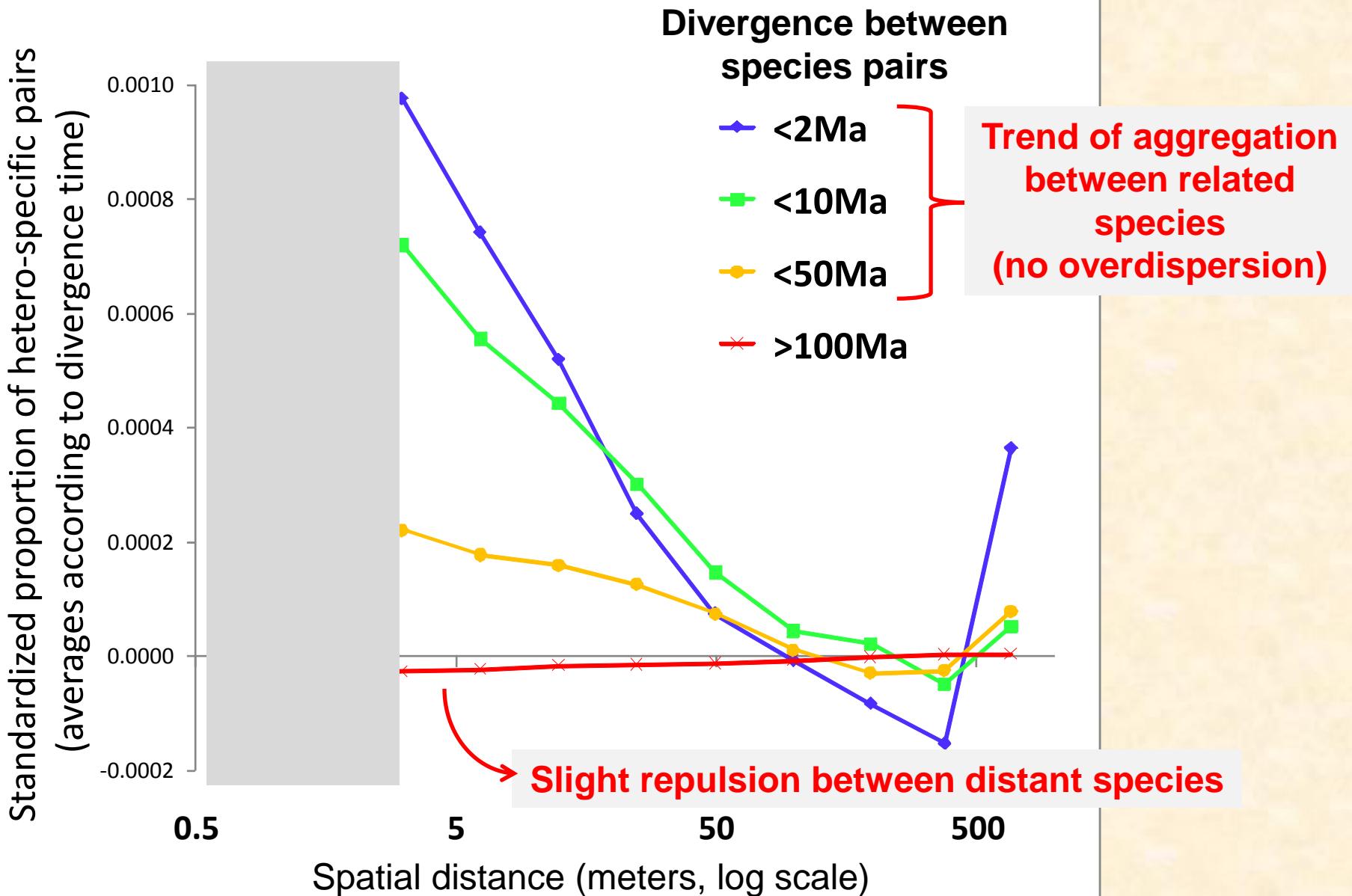
Beilschmiedia jacquesfelixii



Rinorea gabunensis



Korup - Cameroon



Korup - Cameroon

Repulsion between related species at <1m

Divergence between
species pairs

<2Ma

<10Ma

<50Ma

>100Ma

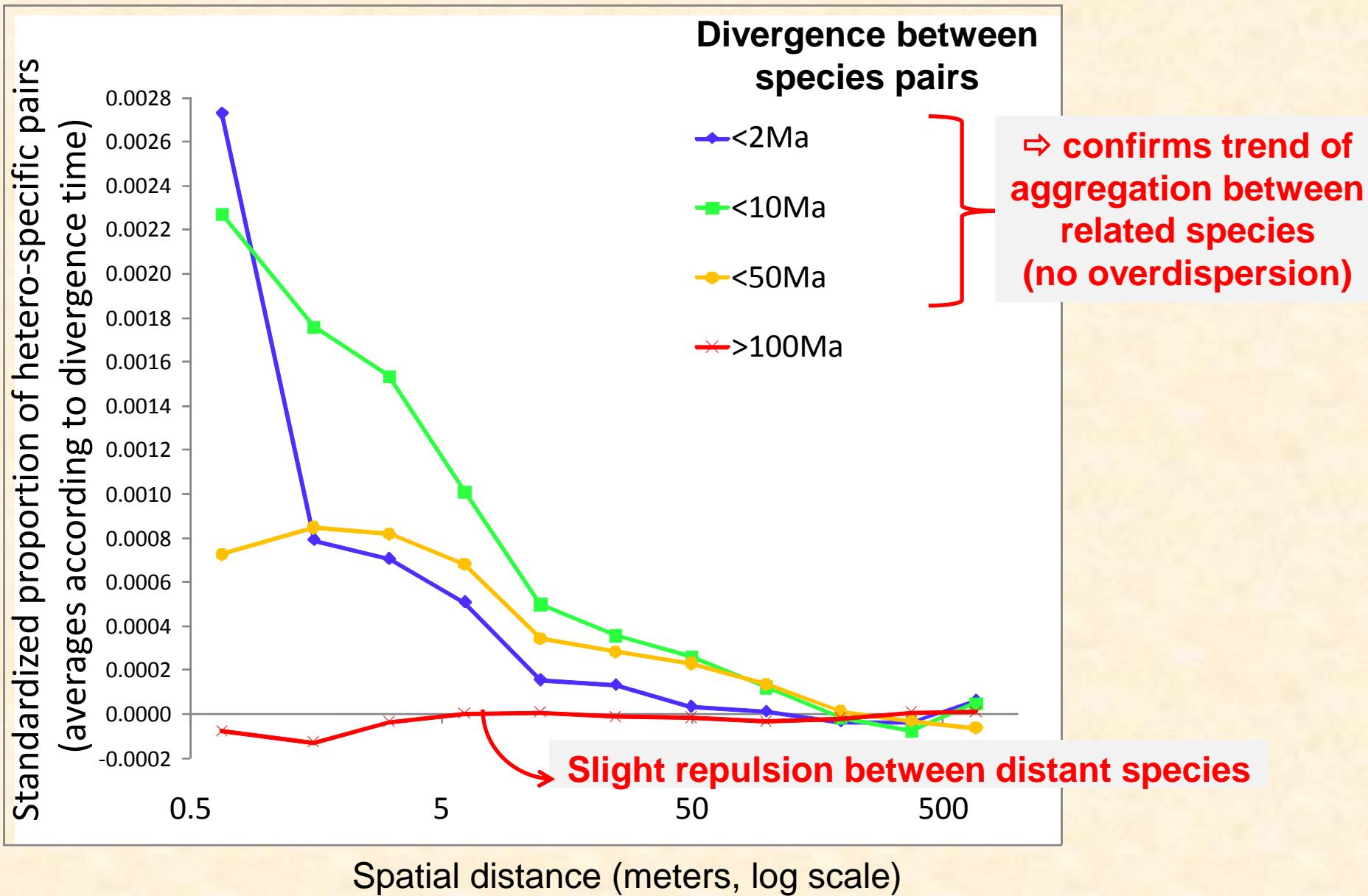
Trend of aggregation
between related
species
(no overdispersion)

Standardized proportion of hetero-specific pairs
(averages according to divergence time)

0.5 5 50 500

Spatial distance (meters, log scale)

Slight repulsion between distant species



Conclusion 5

In rainforest tree communities, **at a local scale**

- phylogenetic clustering due to topographic habitats
- no phylogenetic overdispersion due to competitive exclusion between related species, except may be at very short distance (<1m)

Perspective for modelling

- many metrics/methods used to characterize the phylogenetic or functional structures of communities
 - their interpretation in terms of biological processes essentially based on “verbal” arguments (lack of mathematical formality)
 - neutral processes can generate non-random patterns (Hardy 2008, J Ecol)
- ↳ must evaluate robustness and power (type I and II error rates) of methods against simulated data sets (cf. Münkemüller et al. 2012 Ecography)

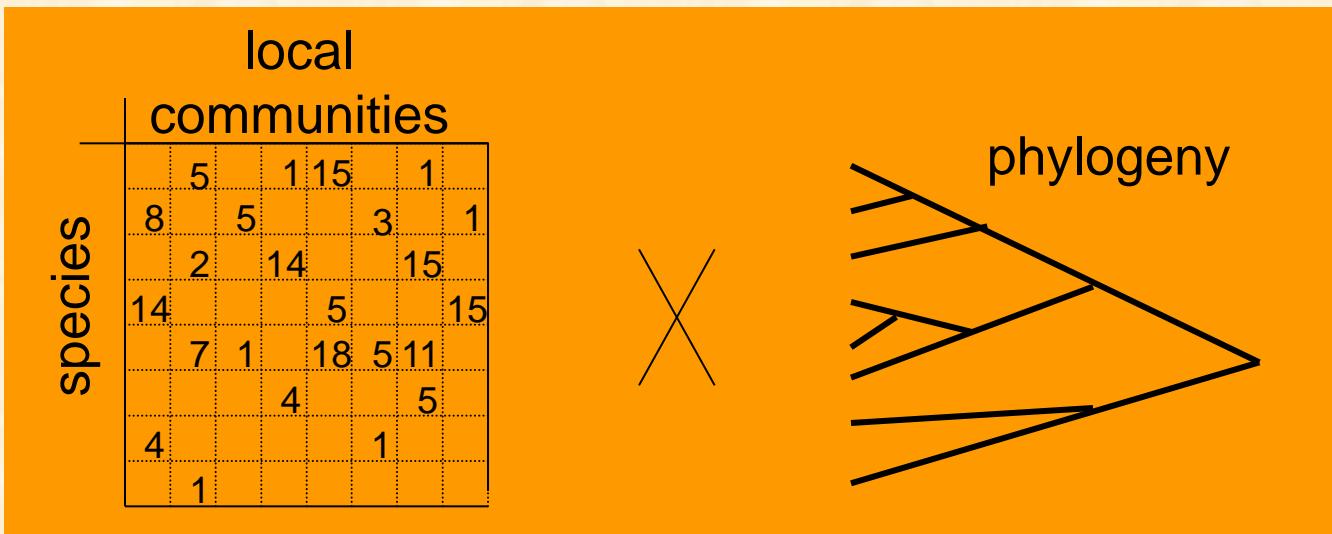
From diversity indices to community assembly processes: a test with simulated data

T. Münkemüller, F. de Bello, C. N. Meynard, D. Gravel, S. Lavergne, D. Mouillot, N. Mouquet and W. Thuiller

Ecography 35: 468–480, 2012

Testing the phylogenetic structure of local communities: which null model is adequate?

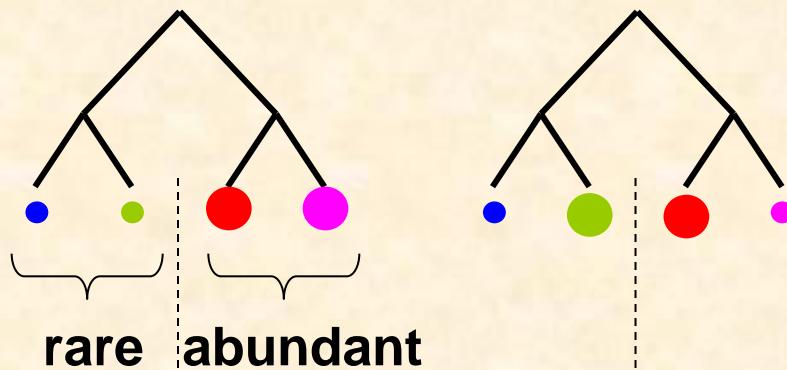
Hardy 2008 (J Ecol 96: 914–926)



⇒ evaluation of Type I error rates for various combinations of metrics and data randomization schemes (null models)

2 types of community phylogenetic structure (CPS)

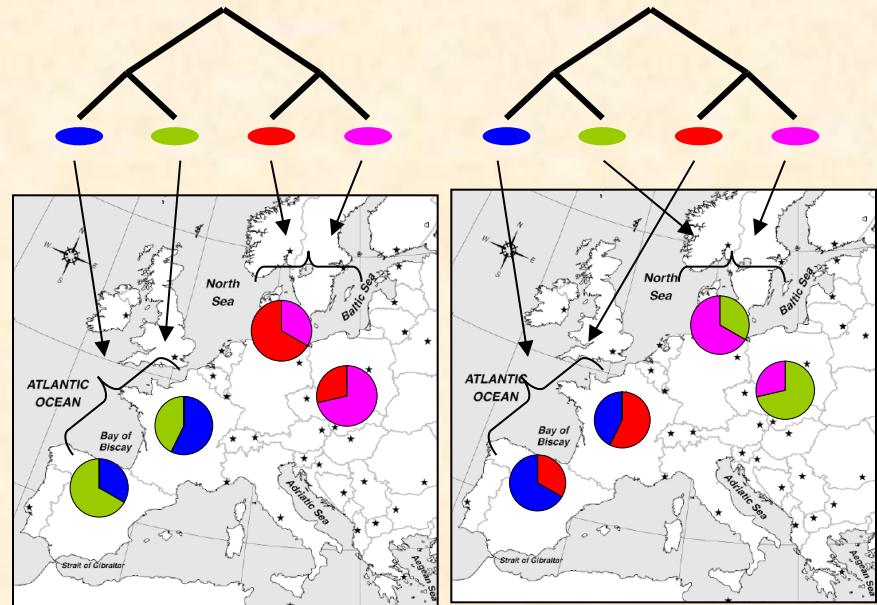
species abundance CPS



phylogenetic clustering

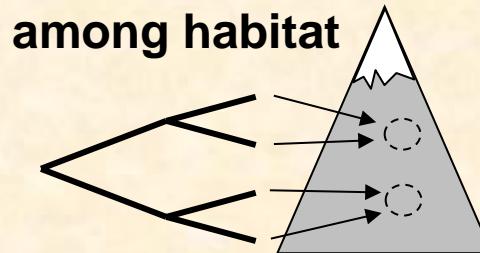
phylogenetic overdispersion

spatial CPS



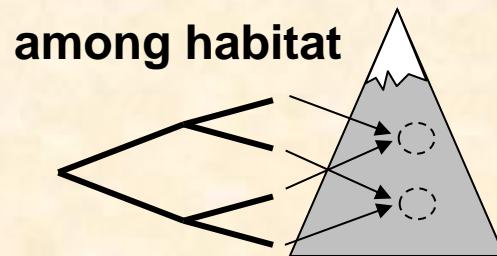
CPS caused by ecological & evolutionary processes

Phylogenetic clustering



environmental filtering
adaptation to altitude
+
habitat conservatism

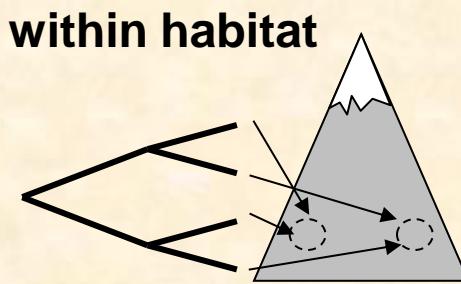
Phylogenetic overdispersion (or evenness)



environmental filtering
adaptation to altitude
+
habitat convergence

e.g. *radiation of several clades in a set of new habitats*

Phylogenetic overdispersion (or evenness)



competitive exclusion
+
niche conservatism
→ sister species cannot coexist because of niche overlap

test of spatial CPS \Leftrightarrow **test of community neutrality**
(if biogeographic effects can be neglected)

How to characterize spatial CPS?

1- Metric to quantify the signal

based on $\left\{ \begin{array}{l} \text{- species presence/absence} \\ \text{- species abundances} \end{array} \right.$

2- Adequate randomization procedure (null model)

- ↳ generates artificial datasets devoid of CPS
- ↳ distribution of metric under H_0
- ↳ test if observed value significantly different

randomization of $\left\{ \begin{array}{l} \text{- phylogenetic tree} \\ \text{- species-plot matrix} \end{array} \right.$

CPS metrics

Using species presence/absence

- ☞ **MPD**: mean phylogenetic distance between species within a site

Webb 2000, Am Nat

variants: **$mMPD_w$** mean *MPD* over a set of sites

- ☞ $\Pi_{ST} = (mMPD_a - mMPD_w) / mMPD_a$

Hardy 2007, J Ecol

relative increase of *MPD* for species sampled **among sites versus within sites**

Using species relative abundances

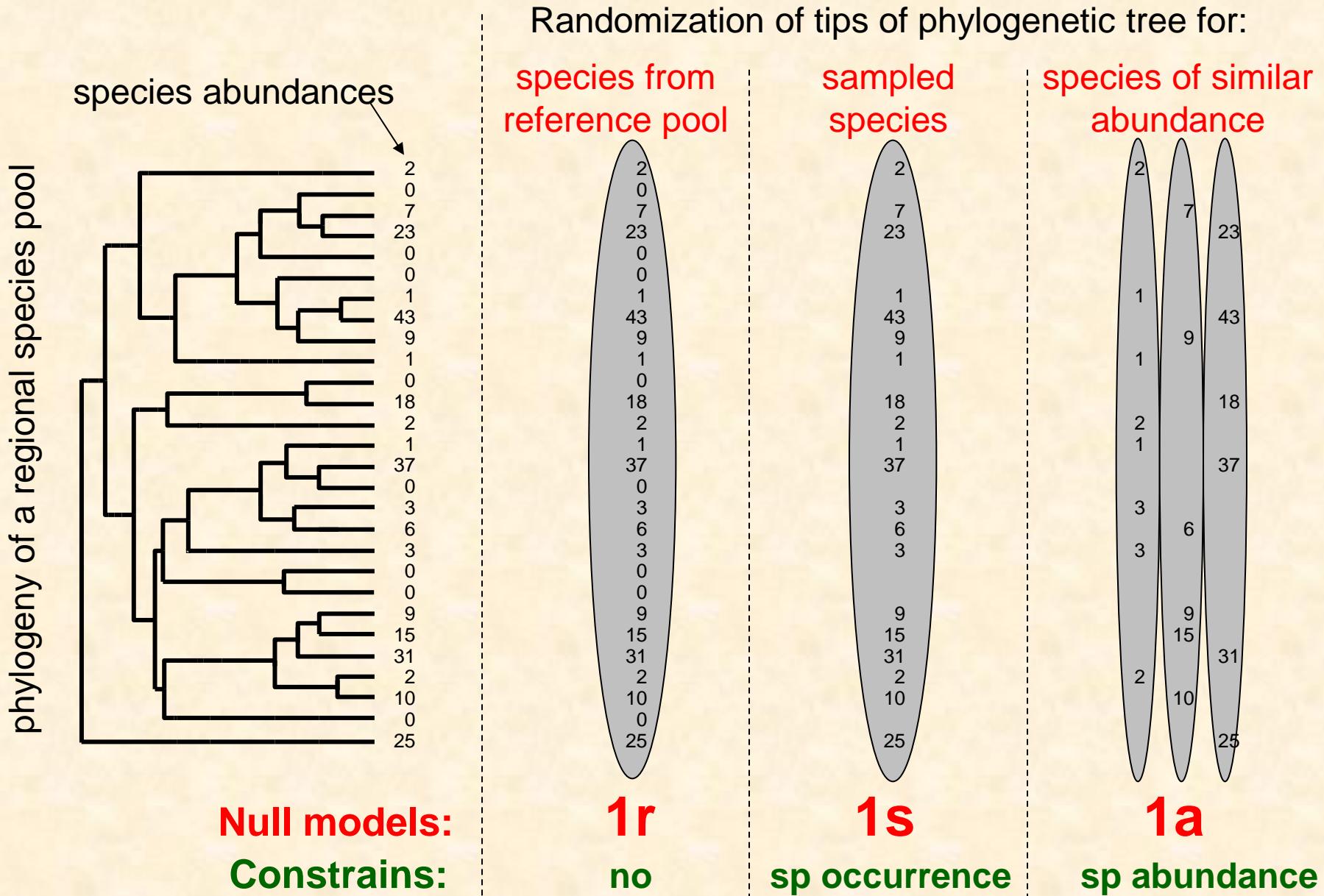
- ☞ **B_{ST}** : analogue of Π_{ST}

($\equiv P_{ST} - I_{ST}$)

Hardy 2007, J Ecol

Null models

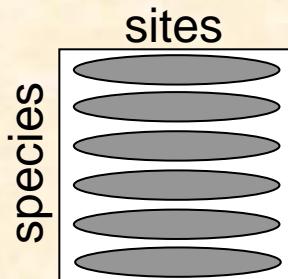
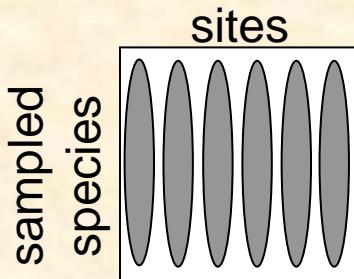
randomizing the phylogenetic tree -> 3 null models



Null models

randomizing the **species – sites matrix** -> 6 null models

independent shuffling of local abundances within species or site

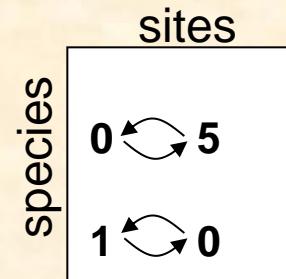


Null models: 2s

Constraints :

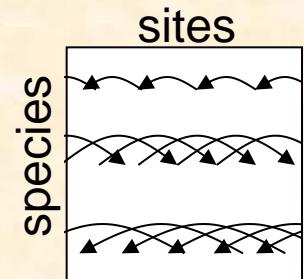
local diversity

Swap algorithm



3x

toroidal translation of local species abundances



3t

abundance distribution for each species

local sp richness

spatial autocorrelation

**All null models (NM) destroy the spatial CPS
but other structural features of the original data set can be affected**

e.g. spatial autocorrelation of species distribution

- maintained under phylogenetic tree randomization (NM 1)
- generally broken under species-site matrix randomization (NM 2 or 3)

e.g. abundance CPS

- broken under phylogenetic tree randomization (except NM 1a)
- maintained under NM 3



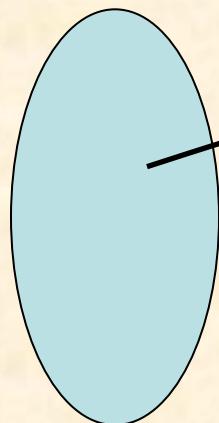
impact on the ‘validity’ of tests ?

(1 test = combination of 1 metric and 1 null model)

A 'locally neutral' subdivided community model

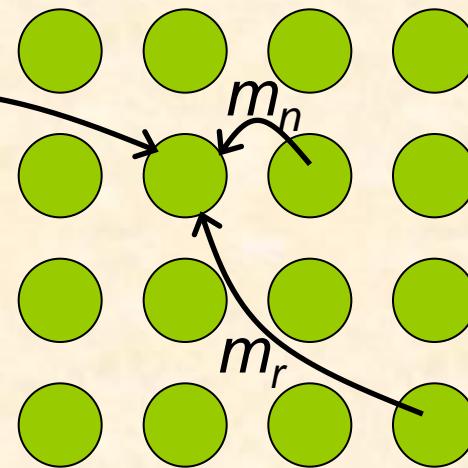
☞ **structure devoid of spatial CPS**

constant **source community**
(≈ regional pool)



$$m_s << m_n, m_r$$

focal community =
set of connected
sub-communities



S species with f_s
relative abundances
(CPS possible)

N sub-communities
containing n_i individuals

Species dispersal
immigration from:
 m_s source community
 m_n adjacent sub-comm.
 m_r random sub-comm.

☛ species assembly within focal community results from:
- species **dispersal**
- **drift** (demographic stochasticity per species)
⇒ **neutral assembly**

Simulation analysis

10 metrics X 9 null models = 90 types of test

↳ applied on 'locally neutral' community model

↳ **simulation** of 10x10 sub-communities

→ migrants from source community ($m_s=0.001$, 227 tree sp from BCI)

variable parameters:

1- sp **abundance CPS**: - **realistic** (as on BCI)

- **random** (BCI sp abund. shuffled)

- **highly clustered** (Fabaceae sp over-abundant)

2- species **dispersal**: - **random** ($m_r=0.1$ $m_n=0$)

- **limited to neighbors** ($m_r=0$ $m_n=0.1$)

- **hierarchical** (4 subsets with $m_{within}=0.1$, $m_{among}=0.01$)

3- sub-community **sizes**: - **constant** ($n_i= 100$ individuals)

- **variable** ($n_i= 40$ or 160 individuals)

↳ **18 parameter sets**

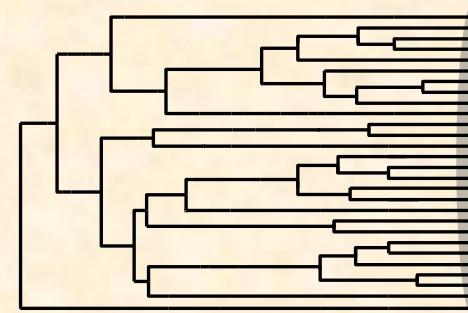
Tests on 1000
data sets →

{ 25 significant clustering
25 significant overdispersion }

expected under H_0
for $\alpha=0.05$

Null model 1s:

*Phylogenetic tree randomization
of sampled species*



data sets showing significant spatial phylogenetic clustering / overdispersion over 1000 replicates (25 expected)

| | Species abundances: | | |
|--------------|----------------------------|----------------|-----------------|
| | P Random | Realistic | P Clustered |
| <i>mMPDw</i> | 21 / 33 | 0 / 81 | 997 / 0 |
| Π_{ST} | 23 / 22 | 34 / 27 | 121 / 25 |

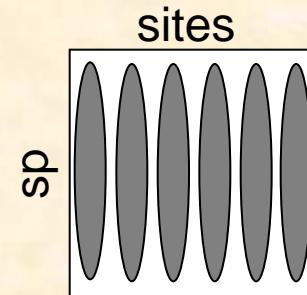
Test :
 ‘exact’
 moderately
 biased
 very
 liberal

⇒ Randomizing phylogenetic tree always OK if
species abundances are phylogenetically random

⇒ *mMPD* (or *NRI*) highly dependent on the phylogenetic structure of species abundances => problem when interpreting the test

Null model 2s:

independent shuffling of local abundances within species



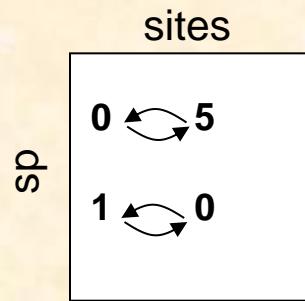
data sets showing significant spatial phylogenetic clustering/overdispersion over 1000 replicates (25 expected)

| | <i>Species abundances:</i> | | |
|-------------|----------------------------|-----------------|-----------------|
| | P Random | BCI | P Clustered |
| <i>mMPD</i> | 337 / 366 | 48 / 782 | 1000 / 0 |
| Π_{ST} | 8 / 5 | 7 / 6 | 50 / 4 |

Test :
‘exact’
moderately
biased
very
liberal

⇒ Most troublesome null model !

Null model 3x: SWAP algorithm



data sets showing significant spatial phylogenetic clustering/overdispersion over 1000 replicates (25 expected)

| | <i>Species abundances:</i> | | | |
|------------------|----------------------------|----------------|----------------|----------------|
| | P Random | P Random | BCI | P Clust |
| <i>Dispersal</i> | | | | |
| | <i>random</i> | <i>limited</i> | <i>random</i> | <i>random</i> |
| <i>mMPD</i> | 25 / 20 | 68 / 61 | 26 / 27 | 17 / 41 |
| Π_{ST} | 25 / 20 | 68 / 61 | 26 / 27 | 17 / 41 |

- ⇒ Much “better” null model
- ⇒ Still somewhat liberal under limited dispersal

Overall performances (type I error rate) over 18 simulation parameter sets

| | Null model | | | | | | | | |
|---------------|------------|------|-------|------|------|-----|------|-----|------|
| Metric | 1s | 1r | 1a | 2s | 2r | 2x | 3i | 3x | 3t |
| $mMNTD$ | red | red | pink | red | red | red | red | red | red |
| $mMPD$ | red | red | pink | red | red | red | red | red | red |
| $D_w^P = Rao$ | red | red | pink | red | red | red | red | red | red |
| Π_{ST} | red | red | green | red | red | red | red | red | red |
| B_{ST} | pink | pink | green | pink | pink | red | pink | red | pink |

Test :
 'exact'
 moderately
 biased
 very
 liberal

- ⇒ **Null model 1a** (randomization of phylogenetic tree constrained by species abundances) performs best
- ⇒ **Metrics B_{ST} more robust**

Hardy 2008 (J Ecol 96: 914–926)

Summary

Most metric X null model combinations

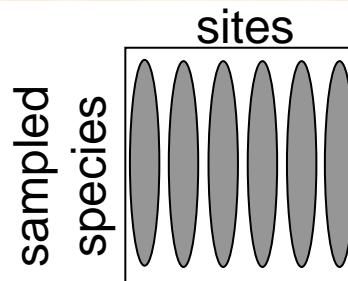
→ test with inconsistent Type I error rate (often liberal tests)

⇒ inadequate to assess if a community is locally neutral

Problems arises

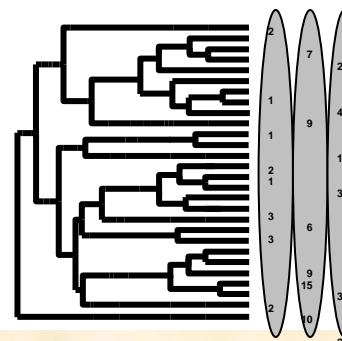
- under phylogenetically non-random sp abundances
- under spatially limited dispersal and/or variable sample sizes

Null model 2r is the least reliable



Null model 1a is the most reliable

(but rarely used in published studies)



Summary

MPD → particularly **sensitive**

↳ does not distinguish **spatial CPS** from **sp abund CPS**

Π_{ST} (and B_{ST}) → more **robust**

↳ focus on **spatial CPS**

Guidelines

Test overall species abundance CPS

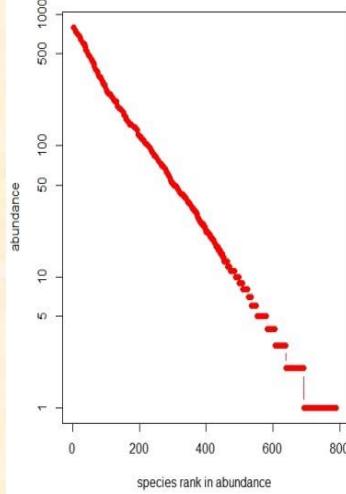
- ↳ if test significant ⇒ use null model 1a (maintains species abundance CPS)
- ↳ if test not significant ⇒ use null model 1s (more degrees of freedom)

Simulateur d'assemblage de communautés avec traits et phylogénie

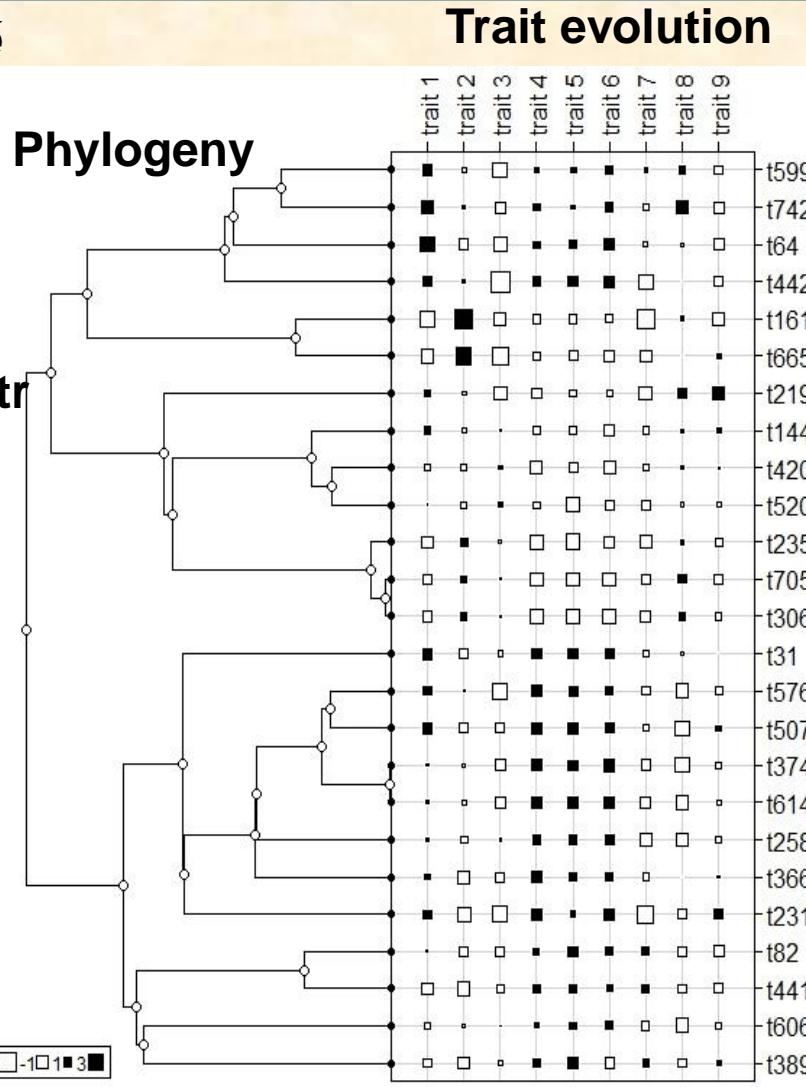
Gilles Dauby

Meta-communauté

Species abund distr (log-series)



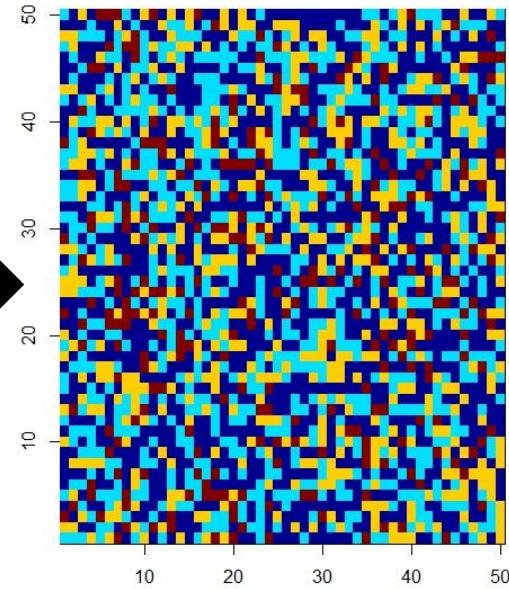
Phylogeny



Trait evolution

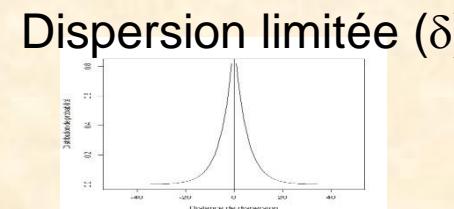
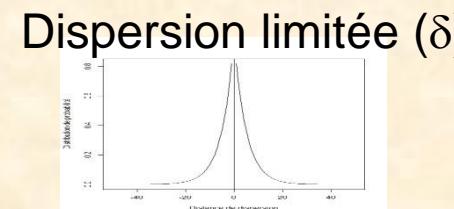
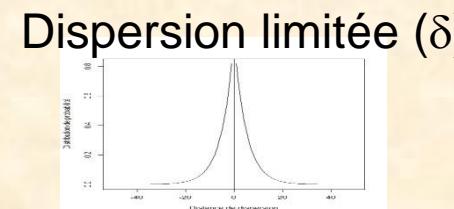
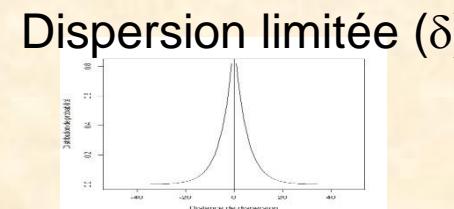
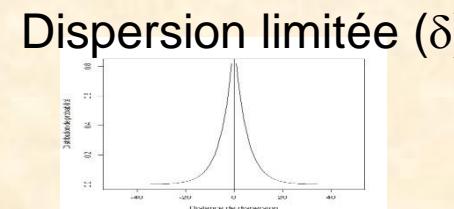
$\mu = 0.001$

Communauté locale



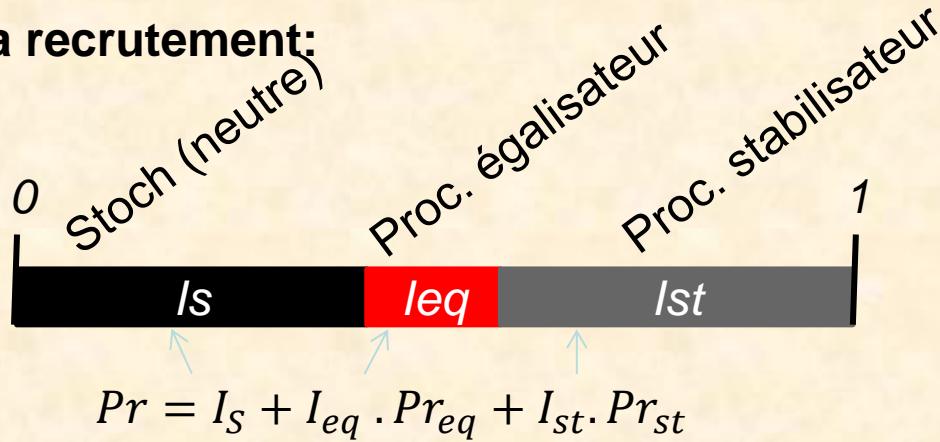
$1-\mu$

Pr, δ



Simulateur d'assemblage de communautés

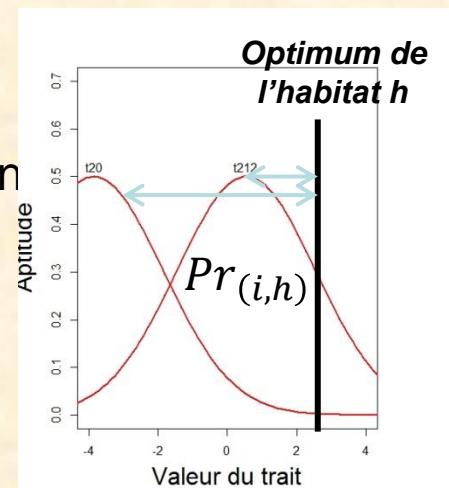
Proba recrutement:



Processus égalisateurs

(capacité compétitive selon filtre environnemental)

$$Pr_{eq(i,h)} = \sum_j f_j \min\left(1, \frac{Pr_{(i,h)}}{Pr_{(j,h)}}\right)$$

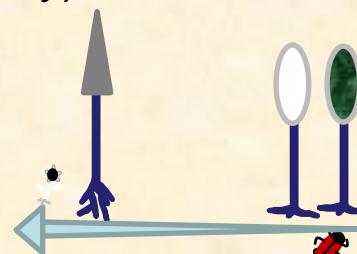


Processus stabilisateurs

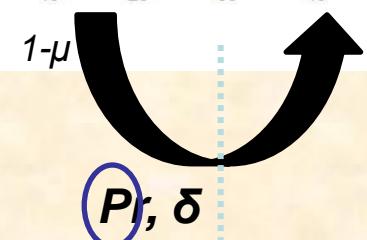
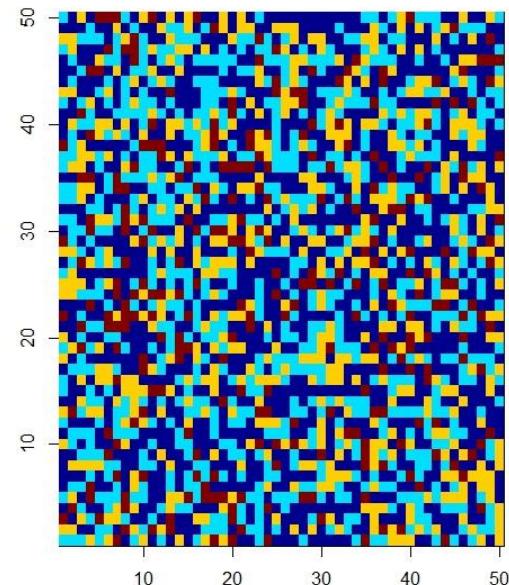
(complémentarité de niche - limiting similarity)

$$Pr_{st(i)} = \sum_j \frac{\overline{d_{i,j}}}{d_{max}}$$

Distance fonctionnelle moyenne



Communauté locale



Simulateur d'assemblage de communautés

Analyses implémentées

Structure fonctionnelle

Indices *Ust*, *Tau_{st}*

Indices intra-sites/inter-sites (diversité beta standardisé)

Hardy O.J. & Senterre B. 2007 J.Ecol.; Jost & Hardy 2008 J.Ecol.

Indices *mMFD*, *mMNTD*

Moyennes d'indices intra-sites (diversité alpha)

Vamosi S.M. et al. 2009 Mol.Ecol.; Hardy O.J. 2008 J.Ecol.

Indices *MFD*, *MNTD* , *NRI*, *NTI*

Indices intra-sites (diversité alpha)

Kembel S. 2009 Ecol.Let.; Hardy O.J. 2008 J.Ecol.

Etendus, écart-type des valeurs de traits

Indices intra-sites (diversité alpha)

Cornwell W.K. & Ackerly D. 2009 Ecol.Mono.

Indices *Ust*, *Tau_{st}*, ...

Corrélation entre indices de Beta diversité s.l. et distance spatiale (décomposée en intra et inter-habitats)

Parmentier I. & Hardy O.J. 2009 Ecography; Swenson N. et al. 2011 Proc.R.B.Soc.

Structure phylogénétique

Indices *Bst*, *Pst*

Indices *mMPD*, *mMNTD*

Indices *MPD*, *MNTD*, *NRI*, *NTI*

/

Indices *Bst*, *Pst*, *COMDIST*, ...

Simulateur d'assemblage de communautés

9 traits

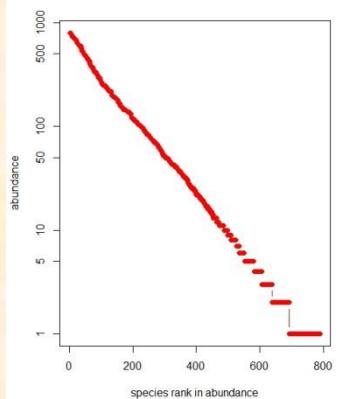
- 3 impliqués dans proc. égalisateurs
- 3 impliqués dans proc. stabilisateurs
- 3 "neutres"

Hétérogénéité environnementale

- Grain de l'habitat
- Nombre d'habitat

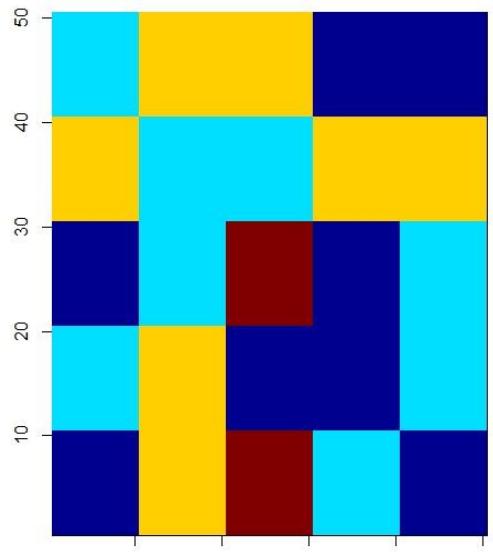
Meta-communauté

NS, Ntraits



μ

Communauté locale



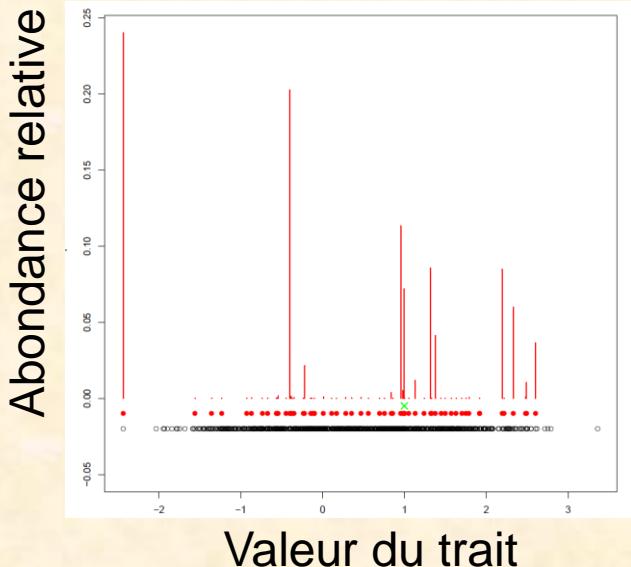
$1-\mu$

Pr, δ

Quelques résultats

1 habitat
homogène

Assemblage
→ Processus
stabilisateurs

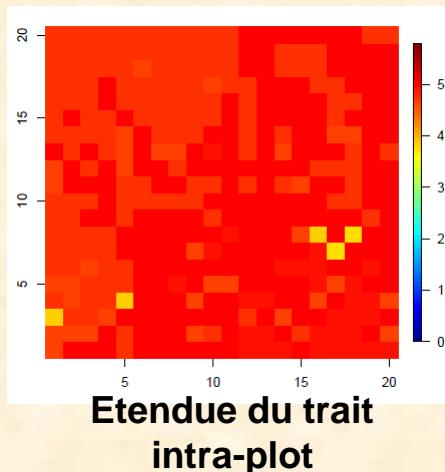
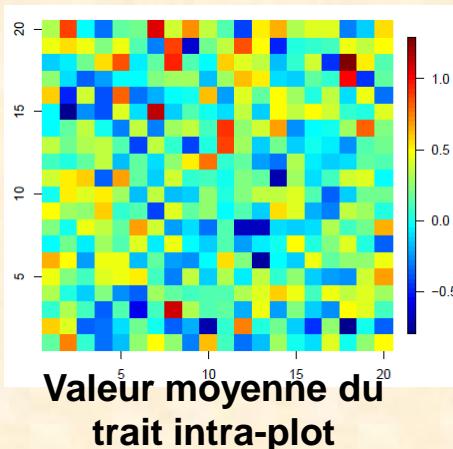


→ Dispersion des traits impliquées

Structure fonctionnelle

- Surdispersion fonctionnelle (métriques **combinant les 9 traits**) détectée pour 80 à 100% des simulations ($p\text{-value} < 0.05$)
- Parfois surdispersion détectée pour les **traits neutres** (jusqu'à 30%)

Ist



Indices Bst , Πst

→ Détection d'une surdispersion dans >70% des cas

Indices $mMPD$, $mMNTD$

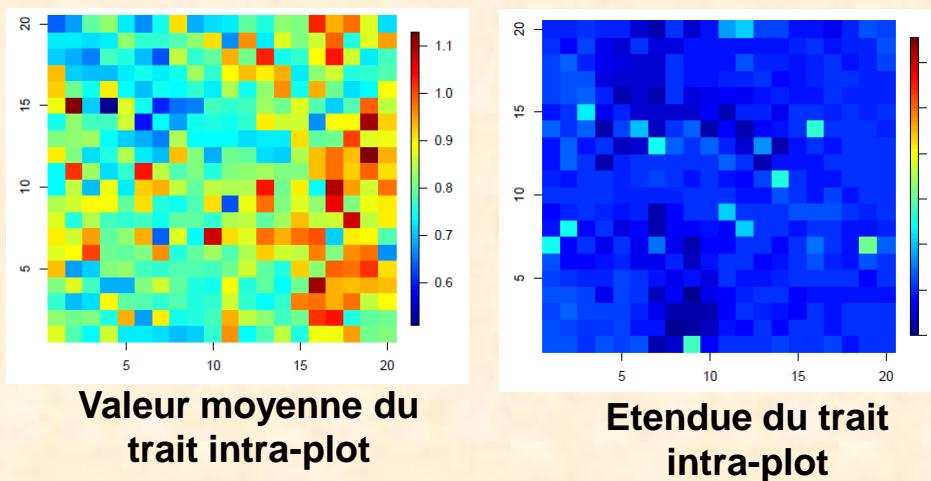
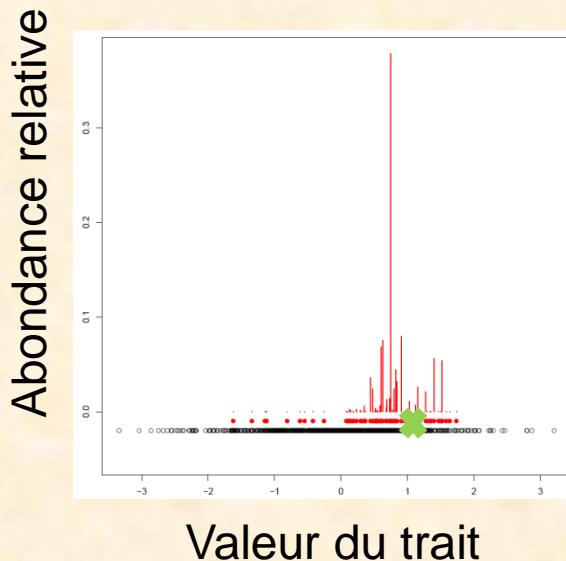
→ Pas détecté

Structure phylogénétique

Quelques résultats

1 habitat homogène

Assemblage
→ Processus
égalisateurs



Indices *mMFD*, *mMNTD*

- Détection d'une agrégation phylogénétique
- Indices biaisés par la **structure phylogénétique des abondances**

Indices *Ust*, *Tau*_{st}

- Insensible à la structure phylogénétique des abondances

→ Correction en utilisant un **modèle nul contrignant les permutations dans des intervalles d'abondances**

Structure fonctionnelle

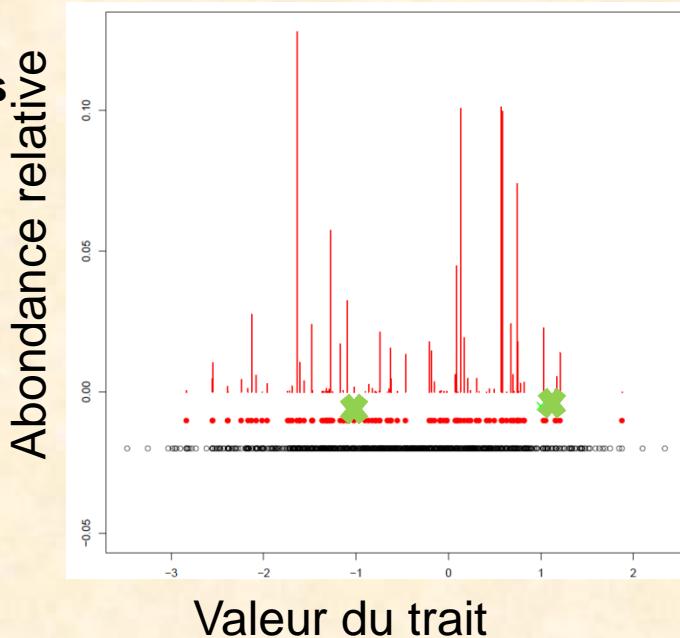
Quelques résultats

2 habitats

Grain d'habitat = 5

Assemblage

→ Processus
égalisateurs



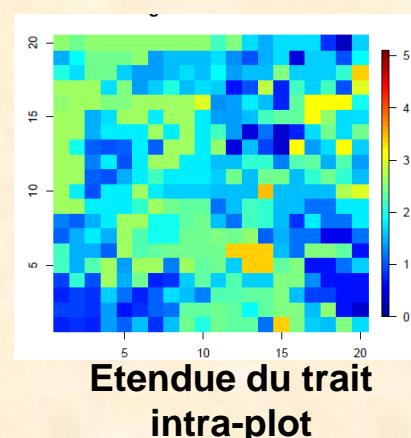
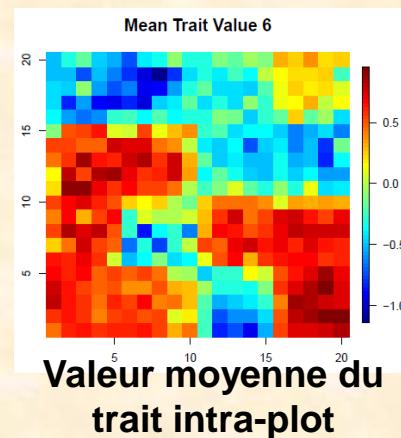
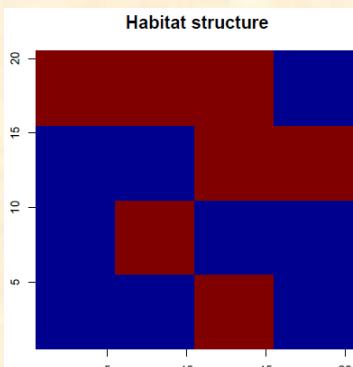
→ Effet fort du filtrage environnemental

→ Agrégation fonctionnelle (métriques **combinant les 9 traits**) détecté pour >80% des simulations ($p\text{-value} < 0.05$)

→ Parfois agrégation détectée pour les **traits neutres** (jusqu'à 30%)

Structure phylogénétique

→ Détection d'une agrégation phylogénétique dans >60% des cas



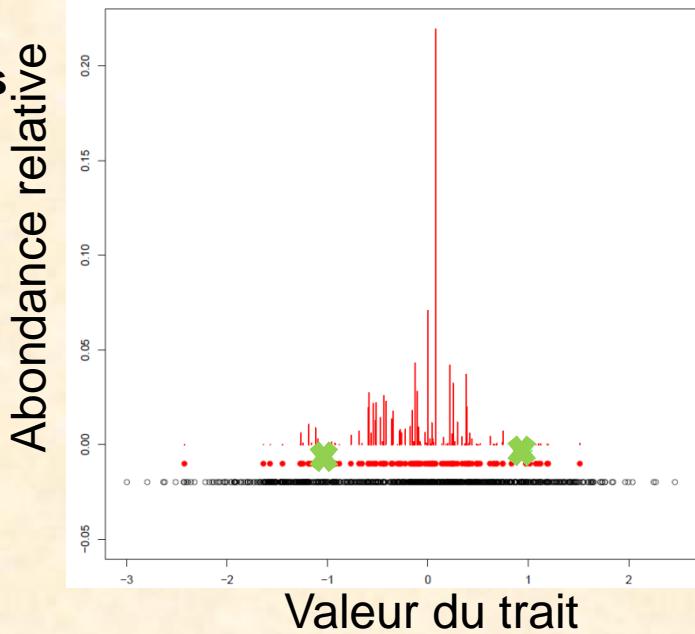
Quelques résultats

2 habitats

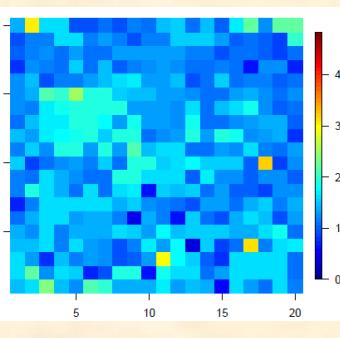
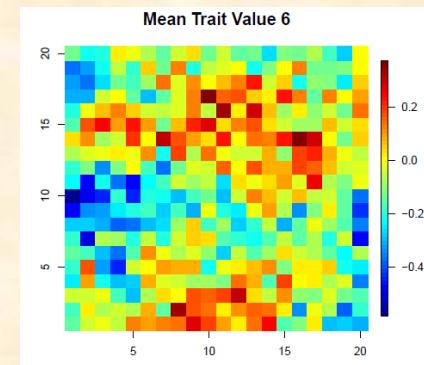
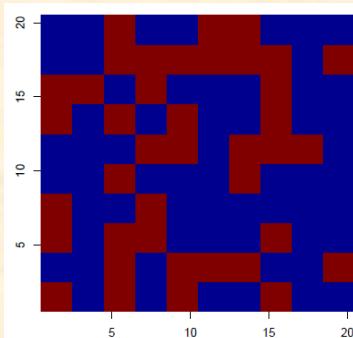
Grain d'habitat = 2

Assemblage
→ Processus
égalisateurs

0



leq



Structure fonctionnelle

Grain de l'habitat + petit
→ Espèces généralistes avantagées

→ Indices *Ust*, *Taust* détectent une agrégation pour les 3 traits impliqués dans les processus pour + de 70% des cas

Structure phylogénétique

→ Pas de structuration phylogénétique détectée

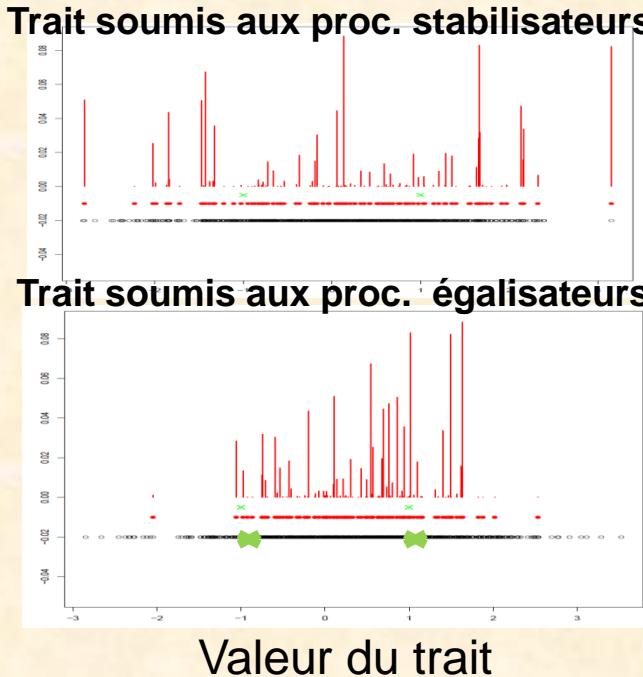
Quelques résultats

2 habitats

Grain d'habitat =5

Assemblage
→ Stoc+Equal
+Stabil

Abondance relative



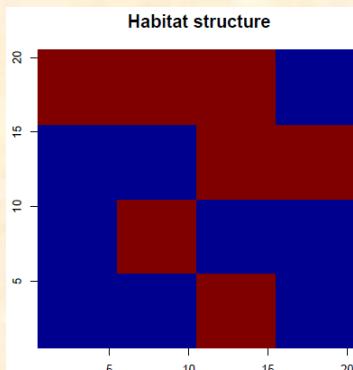
leq

0

lst

ls

↓



Structure fonctionnelle

→ Surdispersion bien détectée pour les traits stabilisateurs

→ Agrégation bien détectée pour les traits égalisateurs

Structure phylogénétique

Indices Pst , $Plst$

→ Structure (surdispersion/agrégation) détectée dans seulement 15 à 20% des simulations → Compensation entre les processus

Should assembly processes be inferred from alpha or beta (phylo/functional) diversity patterns?

Alpha (phylo)diversity: MPDw or QEw (+...)

Beta (phylo)diversity: P_{ST} or B_{ST} (+...)

Phylogenetic clustering usually interpreted as the impact of “environmental filtering” (EF)

However, EF is ubiquitous (fishes do not occur in meadows, mosquitoes do not occur in oceans,...)

-> hence, is the question “does EF occur in a particular local community” relevant?

I don’t think so, EF always occur.

The relevant question is probably: considering two (or a set of) local communities, “does differential EF among species explain different species assemblages”?

↳ this question refers to beta diversity

-> the major focus found in the current literature about phylogenetic/functional community structuring is given to “alpha” diversity where clustering/overdispersion tests rely on the problematic notion of “species pool”. My impression is that more focus should be devoted to beta diversity patterns.

Thank you for your attention