

Model-based geostatistics for wildlife population monitoring : Northwestern Mediterranean fin whale population and other case studies

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joint work with

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Context and aims

- Using data from distance sampling surveys to map animal spatial distributions (line and strip transects)
- Data pooled from multiple sources. Same visual line transect protocol, Only good quality records kept
- Low densities, rare sightings
- Capture-recapture methods based on photo-identification not efficient (recapture probability too low)

Context and aims

- Sighting data are summed on small spatial cells to get count data associated with effort
- **Geostatistical methods** are applied considering : **count** data, **zero inflated** distribution, known **non-stationarity**
- To propose improved form of **Kriging** giving **maps of animal density** and associated **maps of standard error** of prediction

Spatial hierarchical non-stationary model

For all site s (a small spatial cell), Z_s is the number of sightings

$$\begin{cases} Z_s | Y_s \sim \mathcal{P}(Y_s) \\ Y_s = m_s X_s \end{cases}$$

\mathcal{P} independent (given Y) **Poisson** distributions

m_s a **deterministic drift** (habitat characteristics, historical data)

X_s a positive **stationary random field** with unit mean, variance σ_X^2 , covariance function $C_{ss'} = \text{Cov}[X_s, X_{s'}]$, and/or variogram $\gamma_{ss'}$

Drift definition and estimation

The drift m_s resumes spatial non-stationarities

- **Explained non-stationarities** (habitat characteristics)
 - environmental variables as proxies
 - knowledge on spatial potential habitat
 - cokriging of long range spatial components
- **given non-stationarities** (based on past data or a priori)
 - surveys from previous years
 - pooled data from independent sources
 - kernel smoother or filter kriging
- or else, an assumption of stationarity (constant mean)

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Variogram estimation

Let Z_α , $\alpha = 1, \dots, n$ be the n measurements of $Z(s_\alpha)$.
 An experimental variogram of the latent variable X is :

$$\gamma_X^*(h) = \frac{1}{2 N(h)} \sum_{\substack{\alpha, \beta \\ d_{\alpha\beta} \sim h}} \left(\frac{m_\alpha m_\beta}{m_\alpha + m_\beta} \left(\frac{Z_\alpha}{m_\alpha} - \frac{Z_\beta}{m_\beta} \right)^2 - 1 \right)$$

Kriging in non-stationary hierarchical context

The Kriging of Y in site s_o is given by :

$$Y_o^* = \sum_{\alpha=1}^n \lambda_{\alpha} \frac{m_o Z_{\alpha}}{m_{\alpha}} \quad \text{where } \lambda_{\alpha} \text{ are solutions of}$$

$$\begin{cases} \sum_{\beta=1}^n \lambda_{\beta} C_{\alpha\beta} + \frac{\lambda_{\alpha}}{m_{\alpha}} + \mu = C_{\alpha o} & \text{for } \alpha = 1, \dots, n \\ \sum_{\alpha=1}^n \lambda_{\alpha} = 1 \end{cases}$$

The error variance of prediction is :

$$\text{Var}(Y_o^* - Y_o) = m_o^2 \left(\sigma_X^2 - \sum_{\alpha=1}^n \lambda_{\alpha} C_{\alpha o} - \mu \right)$$

Mapping spatial distributions of Fin whale in Mediterranean Sea

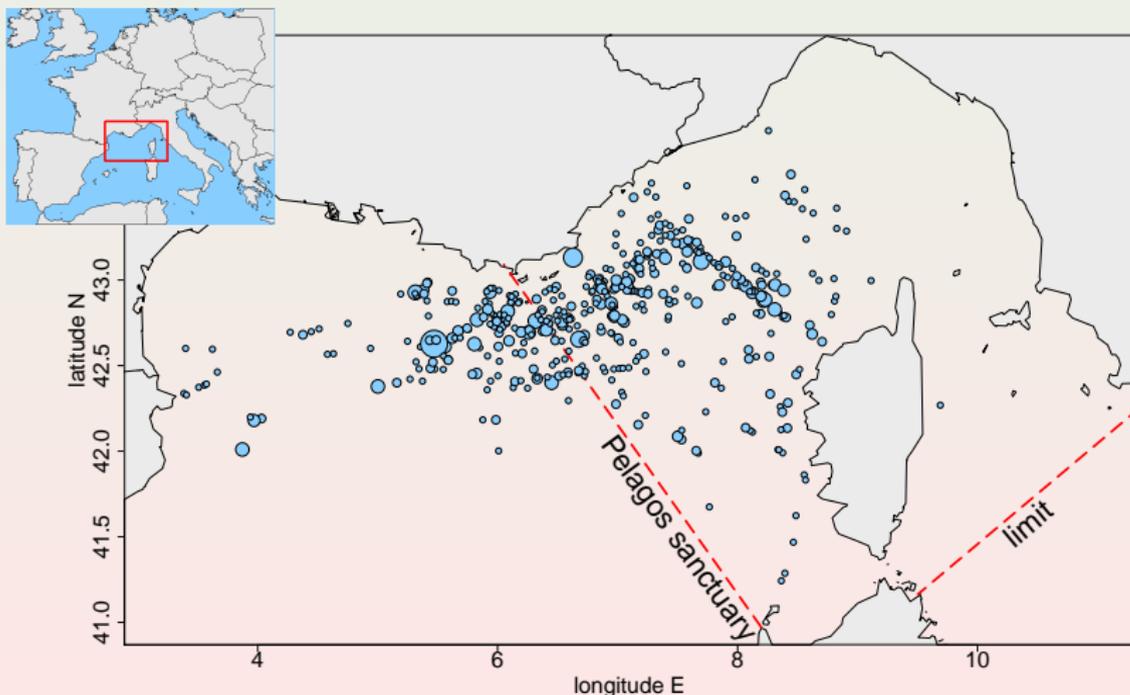
- Presence of a **resident population** in the western Mediterranean Sea (estimations range from 700 to 3500 individuals, *Forcada 1996, Gannier 2006*)
- Classified as endangered by the IUCN Red List.



- population concentrates in summer in the Ligurian basin (100.000 km²)
- An **International Sanctuary** (France, Italy, Monaco) established in 1999 and effective since 2001 (PELAGOS)

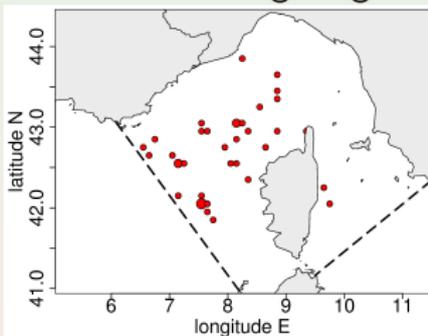
Fin whale summer spatial distribution

Fin whale sightings : raw data 1994 to 2008

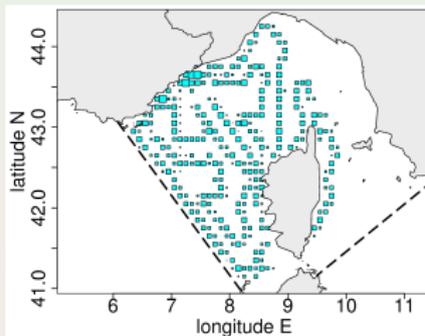


Poisson kriging and density map: Fin whales year 2001

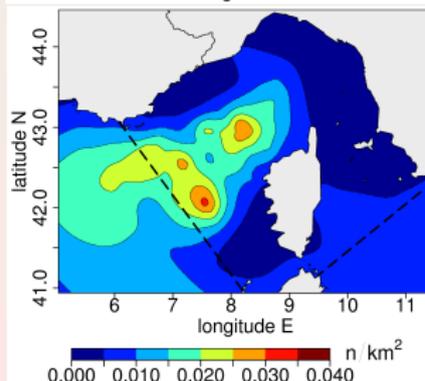
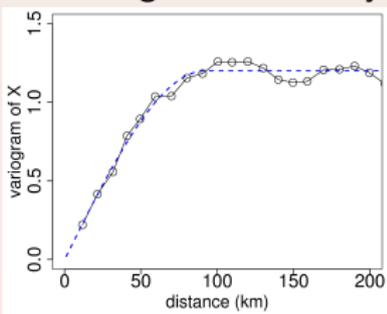
37 fin whale sightings



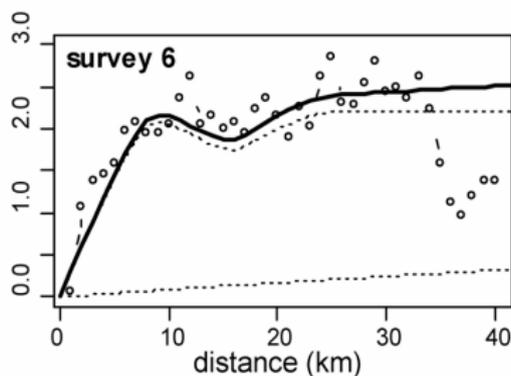
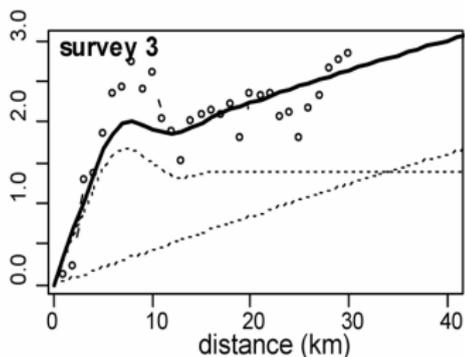
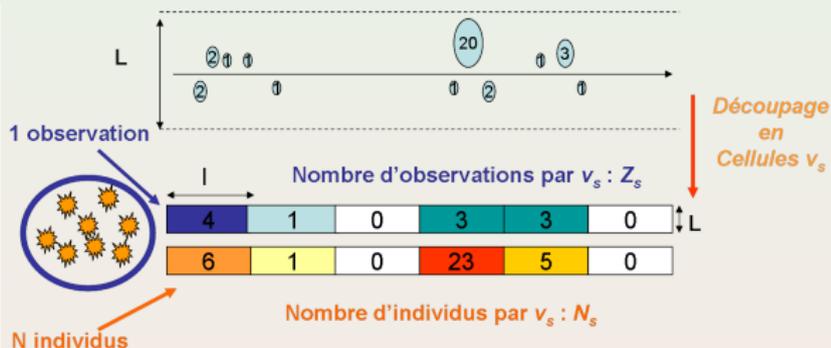
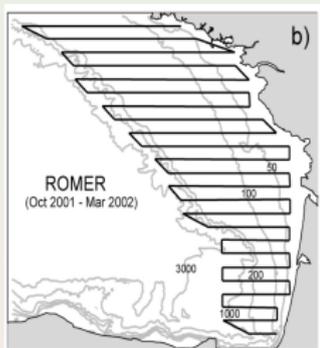
effort 220 hours or 2440 km



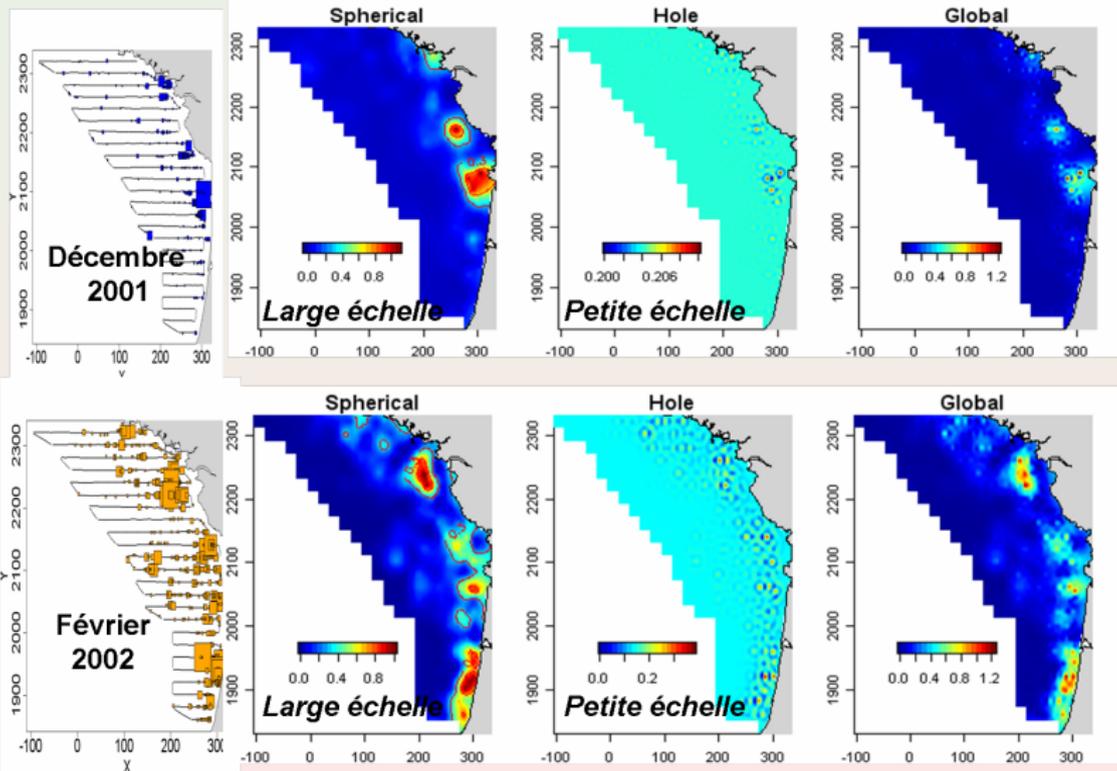
variogram of density X



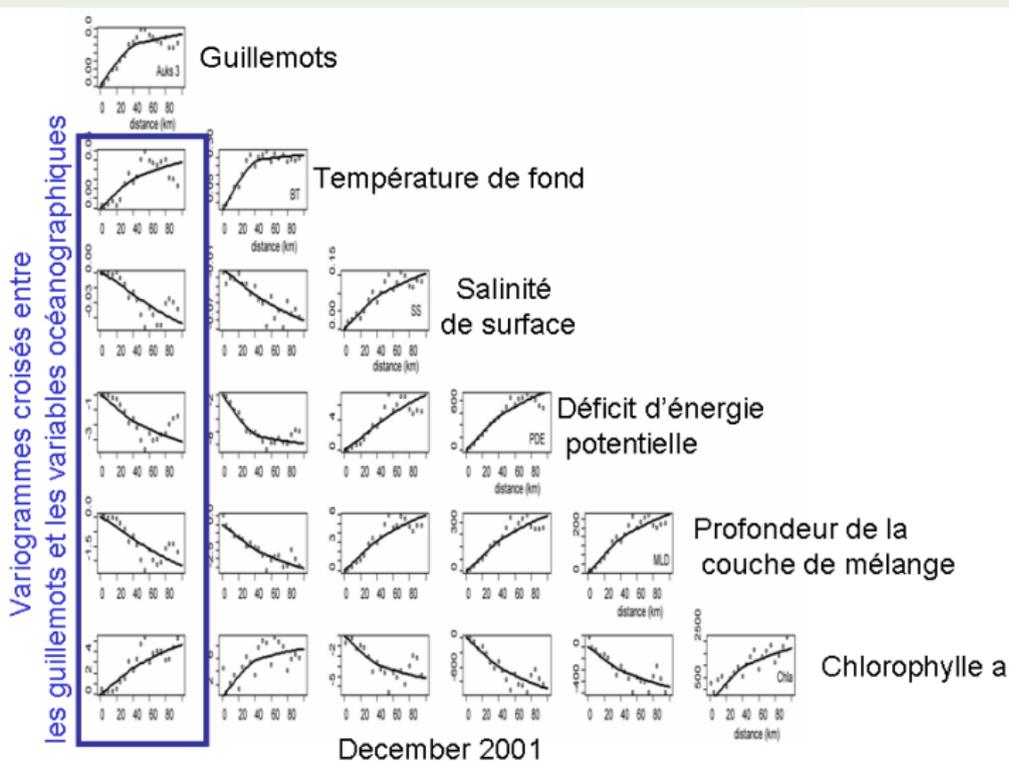
Variograms : Awks in the Bay of Biscay



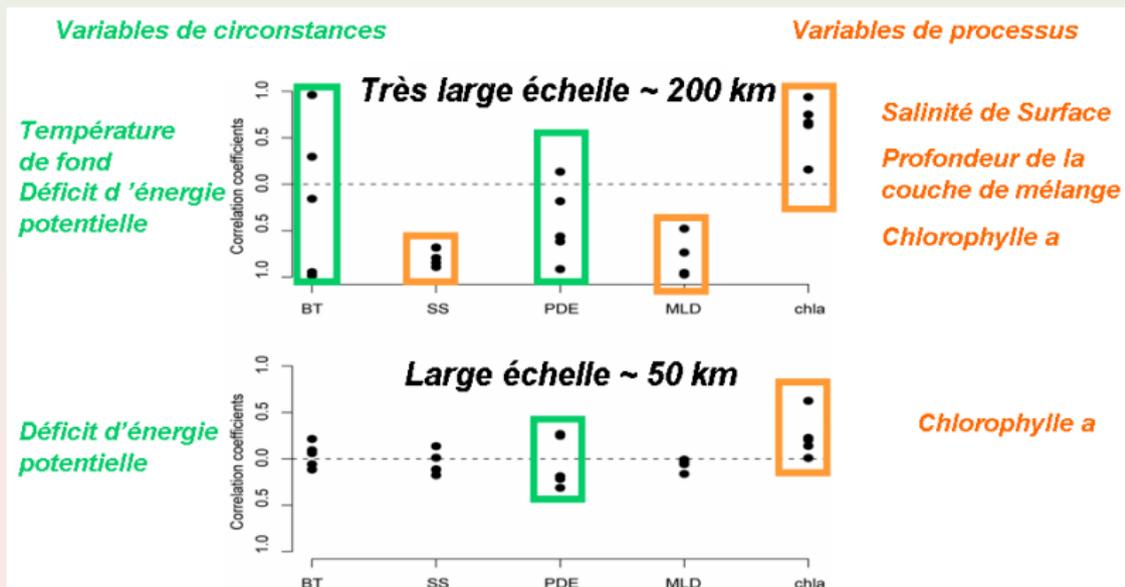
Filter kriging : Awks in the Bay of Biscay



Multivariate model : Linear Model of Coregionalisation with oceanic variables



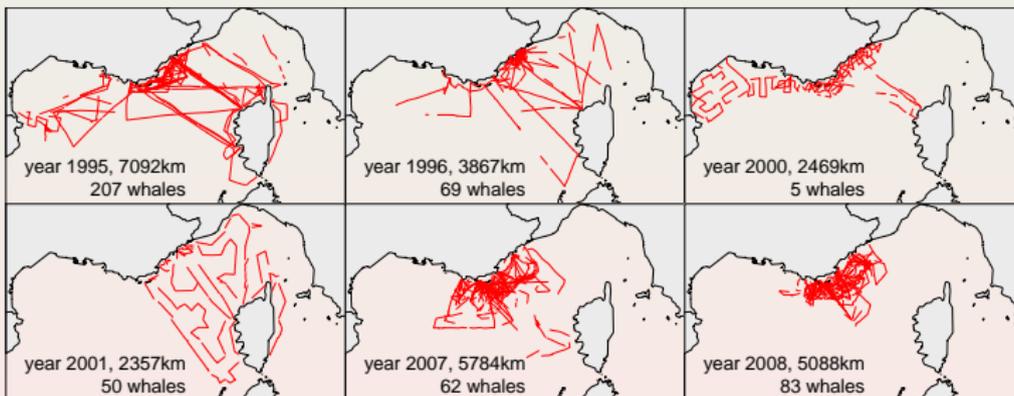
Multivariate model : time variation and/or stability



Distance Sampling : line transect approach

Data collection

- Line transect sampling with quantified efforts and standard protocol to record whale sightings and school size

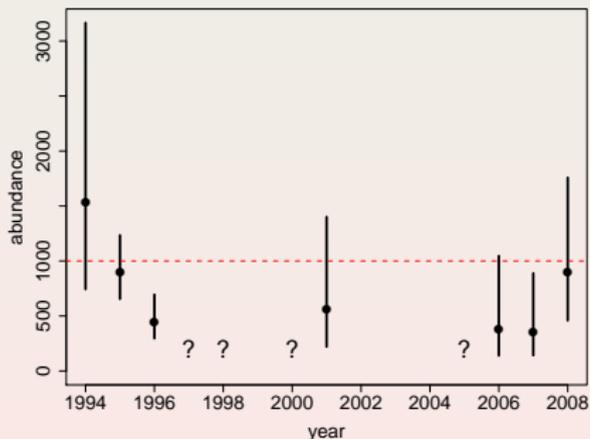


- Changes in sample designs from year to year : logistic, funding, targeted area (examples for 6 different years)

Distance Sampling : line transect approach

Data processing

- fitting the detection function \Rightarrow effective strip width
- variance and confidence intervals by block bootstrap

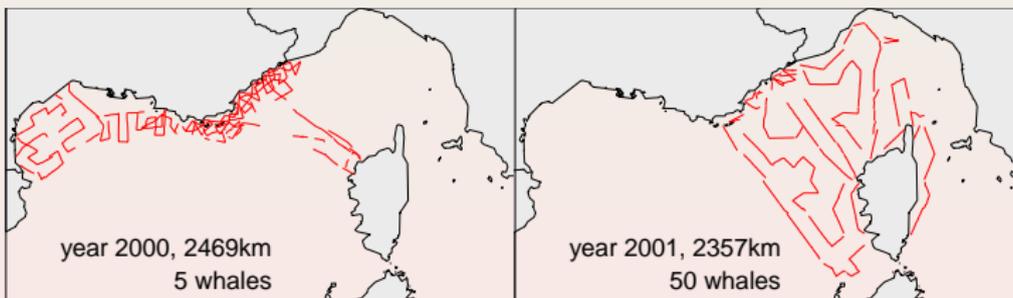


- U shape curve ? Effect of the Pelagos Sanctuary after 2001 ?

Distance Sampling : discussion

- Violation of one major assumption of Distance Sampling

"Independence between sampling scheme and whale spatial distribution"



From whale density to abundance :

Averaged spatial mean by Block Poisson Kriging

block kriging system : $(n+1)$ equations

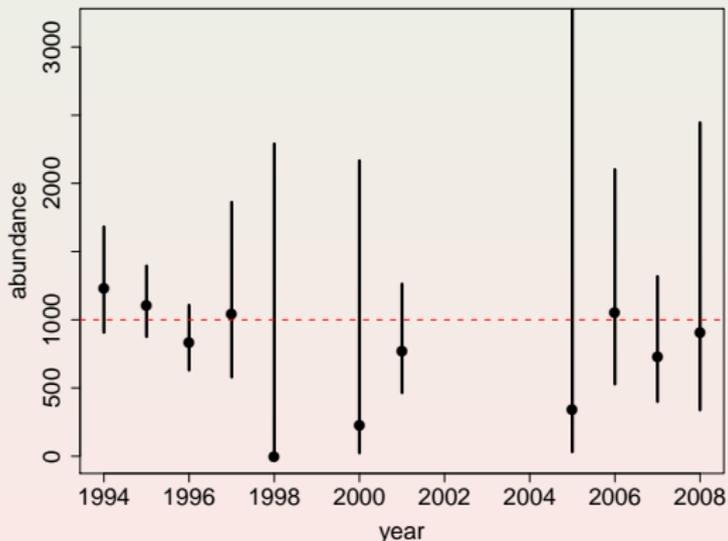
$$\begin{cases} \sum_{\beta=1}^n \lambda_{\beta} C_{\alpha\beta} + \frac{\lambda_{\alpha}}{m_{\alpha}} + \mu = \frac{1}{V} \int_V m_s C_{\alpha s} ds & \text{for } \alpha = 1, \dots, n \\ \sum_{\alpha=1}^n \lambda_{\alpha} = \frac{1}{V} \int_V m_s ds = m_V \end{cases}$$

block kriging variance

$$\begin{aligned} \text{Var}(Y_V^* - Y_V) &= \frac{1}{V^2} \iint_{V \times V} m_s m_{s'} C_{ss'} ds ds' - \sum_{\alpha=1}^n \frac{\lambda_{\alpha}}{V} \int_V m_s C_{\alpha s} ds \\ &\quad - m_V \mu \end{aligned}$$

Abundance estimations by Block Poisson Kriging

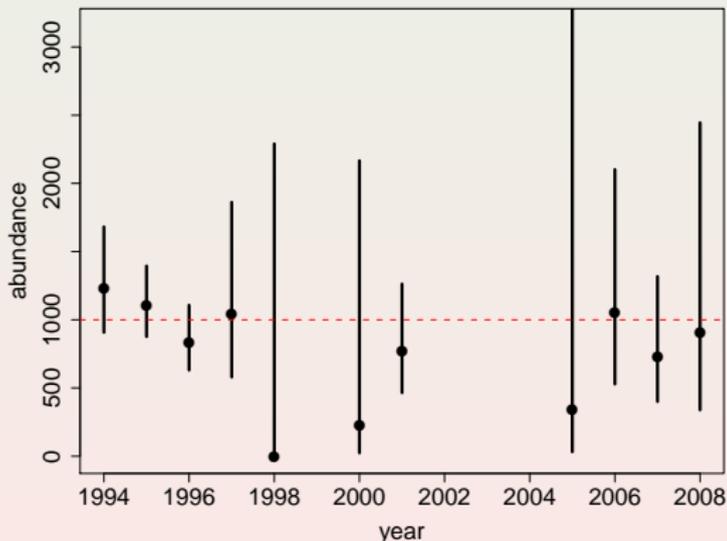
Total abundances versus years
(spatial averaged mean $Y_V^* \times \text{domain } V \text{ area}$)



No significant time trend : population size remains constant ?

Abundance estimations by Block Poisson Kriging

Total abundances versus years
(spatial averaged mean $Y_V^* \times \text{domain } V \text{ area}$)



No significant time trend : population size remains constant ?

Conclusions on methods and models

- Spatial modelling and Poisson Kriging do not totally replace Distance Sampling (some parameters of DS as the detection function, or the school sizes always needed)
- Rigorous protocols remain necessary to get total abundance
- Spatial modelling generally does not reduce confidence intervals, the main purpose is to correct bias from inhomogeneous sampling schemes.

Conclusions on monitoring

- In long term monitoring, survey designs and/or sampling schemes **never** remain constant (empirical law)
- This heterogeneity is accentuated when getting data from multiple sources
- Need to prevent bias or error underestimation due to **sample variation**, need to prevent bias due to **population spatial shifts** (changing environment)
- Then, it is crucial to **model** the animal **spatial distribution in density** (trend and stochastic part), **before concluding on long term variation** : decline, stability or recovery ?

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